

1 **Competitive displacement and agonistic character displacement, or the ghost of**
2 **interference competition**

3

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24 **ABSTRACT**

25 Interference competition can drive species apart in habitat use through competitive displacement
26 in ecological time and agonistic character displacement (ACD) over evolutionary time. As
27 predicted by ACD theory, sympatric species of rubyspot damselflies (*Hetaerina* spp.) that
28 respond more aggressively to each other in staged encounters differ more in microhabitat use.
29 However, the same pattern could arise from competitive displacement, if dominant species
30 actively exclude subordinate species from preferred microhabitats. The degree to which habitat
31 partitioning is caused by competitive displacement can be assessed with removal experiments.
32 We carried out removal experiments with three species pairs of rubyspot damselflies. With
33 competitive displacement, removing dominant species should allow subordinate species to shift
34 into the dominant species' microhabitat. Instead, we found that species-specific microhabitat use
35 persisted after the experimental removals. Thus, the previously documented association between
36 heterospecific aggression and microhabitat partitioning in this genus is most likely a product of
37 divergence in habitat preferences caused by interference competition in the evolutionary past.

38

39 INTRODUCTION

40 Niche partitioning can arise from competition between species in both ecological and
41 evolutionary time (Price and Kirkpatrick 2009; Pfennig and Pfennig 2012). When the presence of
42 one species narrows the realized niche of another in ecological time, this is referred to as
43 competitive displacement (Debach 1966; Reitz and Trumble 2002). Competitive displacement is
44 reversible in that the affected species' realized niche would expand if the competitor was
45 removed. In contrast, when competition between species causes niche divergence over
46 evolutionary time, this is referred to as character displacement, and removal of the competitor
47 might have no immediate effect (Connell 1980; Martin and Ghalambor 2023).

48 Competitive displacement can be a source of selection driving character displacement in
49 habitat preferences, but habitat preferences can also diverge between species in response to
50 selection caused by interspecific competition without competitive displacement. All that requires
51 is heritable variation in habitat preferences and for individuals with habitat preferences closer to
52 the other species' mean habitat preference to have lower fitness as a consequence of interspecific
53 competition. Thus, competitive displacement and character displacement can occur together or
54 separately.

55 Two distinct types of character displacement arising from interspecific competition can
56 drive niche divergence. Ecological character displacement (ECD) is caused by exploitative
57 competition (i.e., resource depletion) and usually leads to reduced resource overlap between
58 species (Schluter 2000; Pfennig and Pfennig 2012). Agonistic character displacement (ACD) is
59 caused by interference competition (e.g., territorial aggression) and involves shifts in traits that
60 affect the rate, intensity or outcome of interspecific encounters (Grether et al. 2009). Both types
61 of character displacement can cause species to diverge in habitat preferences, but the

62 mechanisms of selection differ. With ECD, habitat preferences diverge because of resource
63 depletion where the species overlap (Schluter 2000). With ACD, habitat preferences diverge
64 because interspecific encounters are costly (Grether et al. 2013). ECD is widely regarded as a
65 major driver of niche differentiation between species that compete exploitatively for resources
66 (Schluter 2000; Dayan and Simberloff 2005; Pfennig and Pfennig 2012; but see Stuart and Losos
67 2013). ACD is best known as a mechanism of divergence between species in competitor
68 recognition (Grether et al. 2009, 2017; Seddon and Tobias 2010; Pasch et al. 2017; Moran and
69 Fuller 2018a, 2018b), but it could also be an important cause of habitat divergence, particularly
70 in territorial species (Grether et al. 2013; McEachin et al. 2021).

71 Although interference competition has often been inferred to be the cause of species
72 replacements along elevational and habitat gradients (Heller 1971; Cody and Walter 1976;
73 Schoener 1983; Robinson and Terborgh 1995; Jankowski et al. 2010, 2012; Freeman et al. 2019;
74 Patterson and Drury 2023), the types of manipulative field experiments required to distinguish
75 competitive displacement from ACD are rarely done. Most such experiments that have been
76 published implicate competitive displacement (e.g., (Reed 1982; Garcia 1983; Ebersole 1985;
77 Robertson and Gaines 1986; Ziv et al. 1993; Robertson 1996; Martin and Martin 2001; Stewart
78 et al. 2002; Harmon et al. 2007; Pasch et al. 2013; Edgehouse et al. 2014; Eurich et al. 2018;
79 Martin and Ghalambor 2023), but this could reflect a longstanding bias against publishing
80 negative results. Offering character displacement as an explanation for negative results of
81 competition experiments is what Connell (1980) derisively referred to as invoking the “Ghost of
82 Competition Past”. As Connell (1980) argued, other types of evidence are required to build a
83 convincing case for character displacement. Here we present the results of field experiments to

84 test for competitive displacement in a system for which the requisite evidence already exists (i.e.,
85 negative results would implicate ACD).

86 Interspecific territoriality is a form of interference competition in which individuals of
87 different species compete over space. Quite common in vertebrates (e.g., Orians and Willson
88 1964; Myrberg and Thresher 1974; Wolff et al. 1983; Shimoyama 1999; Maruyama et al. 2010;
89 Suwanvecho and Brockelman 2012; Benson and Patterson 2013; Harris and Siefferman 2014;
90 Cowen et al. 2020; Drury et al. 2020), interspecific territoriality also occurs in certain
91 invertebrate taxa, including the insect order Odonata (damselflies and dragonflies). Some
92 instances of interspecific territoriality in Odonata might be maladaptive byproducts of
93 intraspecific territoriality (Singer 1989; Schultz and Switzer 2001; Tynkkynen et al. 2004), but in
94 rubyspot damselflies (*Hetaerina* spp.) it appears to be an adaptive response to reproductive
95 interference (Drury et al. 2015, 2019; Grether et al. 2020). Male rubyspot damselflies compete
96 for territories in areas with flowing water and submerged vegetation where females oviposit
97 (Johnson 1963; Córdoba-Aguilar et al. 2009; Guillermo-Ferreira and Del-Claro 2011). Territorial
98 fights are energetically demanding and losers are evicted (Córdoba-Aguilar and Cordero-Rivera
99 2005). Males recognize females using visual cues and the females of most sympatric species of
100 *Hetaerina* look very similar (Drury et al. 2015, 2019; Grether et al. 2015). Consequently, males
101 try to mate with heterospecific females, resulting in local mate competition between species
102 (Drury et al. 2015; Grether et al. 2020). Territory holders respond more aggressively to male
103 intruders of species with which they compete more intensely for females (Drury et al. 2015;
104 Grether et al. 2020). Nevertheless, territorial fights are costly and species that overlap more in
105 microhabitat use fight more frequently (McEachin et al. 2021). Thus, this is a system in which
106 selection would be expected to cause the habitat preferences of sympatric species to diverge.

107 Sympatry-allopatry comparison is the canonical method for detecting character
108 displacement in morphology and coloration (Brown Jr. and Wilson 1956), but this is not a
109 suitable method for detecting character displacement in habitat preferences due to an inherent
110 bias toward finding greater habitat differences between species in allopatry than sympatry
111 (Figure 1). This can be understood by considering that, unlike phenotypic traits that are measured
112 on the organism itself, habitat preferences are usually inferred from habitat use, which is
113 constrained by habitat availability. Species with partially overlapping geographic ranges tend to
114 overlap in habitats that are suitable for both species and to occupy different habitats where they
115 do not overlap (Martin and Ghalambor 2023). All habitats at sympatric sites are available to both
116 species, while the habitats at allopatric sites are only available to one species. Consequently, the
117 null expectation is for sympatric species to be more similar in habitat use than allopatric species
118 (Figure 1). Selection could potentially counteract this bias sufficiently to generate a geographic
119 pattern of greater habitat differences in sympatry than allopatry, and cases in which that pattern
120 has been found may be fine examples of character displacement. But when habitat differences
121 between allopatric sites exceed the range of habitats available in sympatry, as must often be the
122 case, there is little scope for selection to generate the predicted pattern of greater species
123 differences in sympatry.

124 Confusingly, some researchers use the term ‘character displacement’ to refer to the
125 pattern of greater species differences in sympatry than allopatry, but most researchers use it to
126 refer to a category of evolutionary processes, and it is well known that those processes can
127 produce other geographic patterns (Schluter 2000; Pfennig and Pfennig 2012; Grether 2018). To
128 test for ACD in the habitat preferences of rubyspot damselflies, McEachin et al. (2021) tested for
129 a predicted pattern of variation among sympatric populations of 10 species. Specifically, they

130 collected data on species differences in territory microhabitat and the responses of territory
131 holders to heterospecific intruders in 25 pairs of sympatric populations. As would be expected if
132 habitat preferences diverged through ACD, species that responded more aggressively to each
133 other differed more in territory microhabitat (McEachin et al. 2021). However, the same pattern
134 could be caused by competitive displacement, if some species are dominant and exclude others
135 from preferred microhabitats (McEachin et al. 2021).

136 To test for competitive displacement, we carried out removal experiments with three of
137 the sympatric species studied by McEachin et al. (2021). If species differences in microhabitat
138 are due to competitive displacement, removal of dominant species should allow subordinate
139 species to settle where they were previously excluded and shift in microhabitat use toward
140 dominant species (Pasch et al. 2013; Eurich et al. 2018). We did not have a priori knowledge of
141 which species of *Hetaerina* are dominant, and in any case if males of only one species were
142 removed, territory holders of the unremoved species would continue defending their territories
143 against newcomers, thereby biasing the outcome in favor of newcomers settling in the
144 microhabitat of the removed species (i.e., in favor of the competitive displacement hypothesis).
145 Instead, we removed all territory holders of both species from selected sections of river and
146 monitored resettlement of these initially competitor-free zones. New males were free to settle in
147 removal sections without interference from established territory holders of either species. To
148 examine how this temporary respite from interference competition affected microhabitat use, we
149 compared the species differences in microhabitat prior to removal to that during the early
150 resettlement period. This experimental design is similar to classical succession experiments in
151 which multiple species of sessile organisms were removed (or clean substrate was introduced) to

152 gain insight into the mechanisms that structure ecological communities, including interference
153 competition (e.g., Sousa 1979).

154

155 **MATERIAL AND METHODS**

156 **Removal experiments**

157 We successfully carried out removal experiments at three sites, each with a different focal
158 species pair, in Costa Rica in 2016 and 2017 (Table 1). (At a fourth site, the experiment was
159 aborted because of heavy rain and flooding.) River transects 200-300 m in length, spanning the
160 river and both banks, were marked at 1-m intervals with numbered flags. We marked all of the
161 males in the transects individually using paint pens (Anderson et al. 2011) and released them
162 where they were captured. Starting > 1 week prior to the removals, two to four observers walked
163 along the transects between 0900 and 1700 h recording the locations of marked males to the
164 nearest 0.1 m using hand-held computers. Males were classified as territory holders if they
165 perched repeatedly in the same location (± 2 m) close to the surface of the water (< 1 m) on at
166 least two consecutive days (Anderson and Grether 2010). Damselfly territories do not have
167 discrete boundaries, but prior to the experimental removals, the area within which individual
168 territory holders could fly and perch without being attacked by neighboring territory holders
169 ranged in size from ca. 4 to 9 m². In narrow sections of river (< 3 m) a male's territory could
170 span both banks, but in wider sections a transverse line from one bank to the other could cross
171 the territories of multiple males. The transects were divided into sections and removals were
172 done in every other section (4-6 removal sections per site; Table 1). All *Hetaerina* spp. territory
173 holders in removal sections were captured with aerial nets and released on the same river > 100
174 m above or below the transect, which was sufficient to keep them from returning during the

175 experiment. We erred on the side of removing males whose territorial status was unclear, but
176 females and males without territories were not removed. We continued marking new males and
177 recording the locations of marked males for 1 week post removal. The decision to monitor post-
178 removal settlement for 1 week was based on our knowledge of the study system and the goal of
179 the experiment, which was to document territory settlement in the absence of interference
180 competition. Stopping much sooner would not have allowed sufficient time to document territory
181 settlement; it takes at least 2 days to determine whether a male is a territory holder, and we
182 expected the removal sections to be settled gradually. Continuing to monitor settlement for
183 longer than necessary would have defeated the purpose of the experiment because as occupancy
184 increased further, later settlers would have fought with early settlers.

185

186 **Microhabitat measurements**

187 We measured the same four microhabitat variables as McEachin et al. (2021): stream width,
188 current speed, canopy cover, and perch height. Stream width (to 0.1 m) and current speed were
189 recorded at 2-m intervals and overstory canopy cover was recorded at 5-m intervals along the
190 transects. Current speed was assessed visually on a 5-point ordinal scale, with 0 for “still”, 1 for
191 “slow” (no waves), 2 for “wavy” (standing waves), 3 for “rippling” (interfering waves), and 4 for
192 “turbulent” water (breaking waves), near both banks and in the middle of the river. A concave
193 spherical densiometer (Forestry Suppliers, Inc.) was used to measure percent canopy cover in the
194 middle and near both banks where the river was ≥ 10 m wide, or near both banks where the river
195 was 3-10 m wide, or in the middle where the river was < 3 m wide.

196 Horizontal coordinates along the transect were used to obtain mean values of stream
197 width, current speed, and canopy cover for each territory holder by interpolation.

198

199 **Data analysis**

200 To examine whether territories in the removal sections differed in microhabitat from those
201 outside the removal sections, we constructed multivariate linear models with the microhabitat
202 variables as the dependent variables and location (inside versus outside) as a categorical
203 predictor variable.

204 Euclidean distances in microhabitat space (i.e., the four-dimensional space defined by the
205 microhabitat variables) were used to quantify differences between species before and after the
206 experimental removals. We used the same metric to test the prediction that one species at each
207 site shifted toward the other species' microhabitat during the resettlement period. In principle,
208 this approach might detect species differences and shifts in microhabitat that would not be
209 detected by analyzing the microhabitat variables separately. To weight the four microhabitat
210 variables equally, they were converted to z-scores by subtracting the site mean and dividing by
211 the site standard deviation. Euclidean distances between territory holders in microhabitat space
212 (henceforth, "distances") were calculated using the z-scores.

213 Each territory holder contributed two data points to the analysis of species differences in
214 microhabitat: a mean distance to conspecific territory holders and a mean distance to
215 heterospecific territory holders. Non-parametric Wilcoxon paired tests were used to compare the
216 conspecific and heterospecific distance distributions because the data were not normally
217 distributed. For testing the competitive displacement prediction that one species at each site
218 shifted toward the other species' microhabitat, each new territory holder in the removal sections
219 contributed one data point: a mean distance to heterospecific territory holders that were present
220 before the removals. Non-parametric Wilcoxon two-sample tests were used to compare those

221 mean distances to the distribution of mean distances between territory holders of the two species
222 before the removals. To account for testing the prediction twice at each site (i.e., once for each
223 species), the p -values were adjusted using Holm's method, which is also known as the sequential
224 Bonferroni method (Holm 1979; Hochberg and Benjamin 1990). In practice, however, the results
225 were unaffected by this adjustment at $\alpha = 0.05$.

226 For an overall test of the effect of the experimental removals on the species difference in
227 microhabitat, data from all sites were combined and a linear mixed-effects model was fit with the
228 'lmer' function in R package 'lme4' (version 1.1-34). To meet model assumptions, log Euclidian
229 distance was used as the dependent variable. The model's fixed effect terms were species
230 comparison (i.e., conspecific vs. heterospecific), period (i.e., pre- vs. post-removal) and their
231 interaction, and the random effect terms were site and individual. Thus, the model took into
232 account that each territory holder contributed two values (mean distances to conspecifics and
233 heterospecifics) and also the nesting of individuals within sites.

234 The data analysis was carried out in R version 4.2.2 (The R Foundation for Statistical
235 Computing).

236

237 **RESULTS**

238 Territories in the removal sections did not differ in microhabitat from those outside the removal
239 sections before or after the experimental removals, at any of the three sites (multivariate linear
240 models, before removal: site GO01 $t = 0.85$, $p = 0.4$; site MV05 $t = 0.91$, $p = 0.4$; site RT02
241 before removal, $t = 1.02$, $p = 0.3$; after removal: site GO01 $t = 0.12$, $p = 0.9$; site MV05 $t = 0.18$,
242 $p = 0.9$; site RT02 $t = -1.01$, $p = 0.3$).

243 Males began settling in the removal sections almost immediately, and the number of new
244 territory holders continued increasing for the duration of the experiment (Figure S1). The number
245 of new territory holders varied, but in none of the removal sections did the number of new males
246 of either species reach pre-removal levels by the end of the monitored resettlement period (Table
247 1). The species differed significantly in territory microhabitat before the experimental removals
248 (Table 2, Figure 2) with one species perching higher and in areas with greater canopy cover than
249 the other at all three sites (Figure 3). The species largely overlapped on the other two
250 microhabitat axes (current speed and stream width; Figure S2). Significant species differences in
251 territory microhabitat were found in the removal sections during the early resettlement period at
252 two sites (Table 2, Figure 2), and the species differences in perch height and canopy cover were
253 in the same directions as prior to the removals at all three sites (Figure 3). Contradicting the
254 competitive displacement hypothesis, none of the species shifted significantly toward the other
255 species' microhabitat during the early resettlement period (Figure 4). With data from the three
256 sites combined, the species difference in microhabitat was highly significant (linear mixed
257 model; $t = 28.95, p < 0.0001$) but was not affected by the experimental removals (species
258 comparison \times period interaction, $t = -1.44, p = 0.15$). The mean distance in microhabitat space
259 between conspecifics decreased between the pre-removal and post-removal periods ($t = -2.87, p =$
260 $= 0.02$) but there was no change in the mean distance between heterospecifics ($t = -1.12, p =$
261 0.68). Thus, the only overall effect of the experimental removals was a reduction in the mean
262 microhabitat distance between conspecifics.

263

264

265

266 **DISCUSSION**

267 We carried out removal experiments at three sites to determine whether microhabitat partitioning
268 in rubyspot damselflies is caused by ongoing competitive interactions (i.e., competitive
269 displacement). While the experiments were small in spatial scale and short in duration by
270 vertebrate standards, they were designed for insects with territory sizes in the range of 4-9 square
271 meters and adult life spans of weeks. The species differences in microhabitat were unaffected by
272 the experimental removals at two sites (Figures 2 & 3; Table 2). At a third site, the species did
273 not differ in microhabitat during the resettlement period, but the initial species difference was
274 also smaller than at the other two sites (Figures 2 & 3; Table 2). In no case did one species shift
275 significantly toward the other species' microhabitat (Figure 4). Males that established new
276 territories in the removal sections often perched where conspecific territory holders had perched
277 previously, even though other perches were available (S.M. pers. obs.). However, the mean
278 microhabitat distance between conspecifics decreased somewhat after the experimental removals
279 (Table 2), perhaps because in the absence of competition, males were free to settle according to
280 their species-typical habitat preferences.

281 Evidently, the differences between these species in microhabitat are the product of
282 differing habitat preferences, not ongoing competitive displacement. This result alone does not
283 implicate agonistic character displacement (ACD) because the species could differ in habitat
284 preferences for a multitude of reasons unrelated to interference competition. However, a previous
285 study of 25 pairs of sympatric populations of rubyspot damselflies showed that microhabitat
286 partitioning increases with the level of heterospecific aggression (McEachin et al. 2021).
287 Heterospecific aggression refers not to the actual rate of interspecific fighting but instead to how
288 aggressively territory holders respond to heterospecific intruders relative to conspecific intruders.

289 Species differences in microhabitat reduce the frequency of interspecific encounters that result in
290 interspecific fights (McEachin et al. 2021), and thus the positive correlation between
291 microhabitat partitioning and heterospecific aggression is a predictable consequence of selection
292 against interspecific fighting (i.e., ACD). Note that if the direction of causality were reversed,
293 and heterospecific aggression evolved in response to pre-existing species differences in
294 microhabitat, the correlation between microhabitat partitioning and heterospecific aggression
295 would be negative (i.e., species that differ more in microhabitat would be less aggressive to each
296 other). The purpose of carrying out removal experiments on a subset of the same species pairs
297 was to test an alternative hypothesis, namely that the positive correlation arose from dominant
298 species excluding subordinate species from preferred microhabitats. While ACD and competitive
299 displacement are not mutually exclusive processes, and both could have contributed to the
300 species differences in microhabitat, we found no evidence for competitive displacement.

301 Species sorting (i.e., differential extinction) is another process that could potentially
302 produce a positive correlation between habitat partitioning and heterospecific aggression if
303 interspecific fighting reduced the probability of species co-occurring (Pigot and Tobias 2013;
304 Rybinski et al. 2016; McEachin et al. 2021). The likelihood that species sorting could have
305 generated the geographic pattern documented by McEachin et al. (2021) cannot yet be fully
306 assessed due to a paucity of research on the effects of species interactions at the adult stage on
307 coexistence in Odonata (Svensson et al. 2018; Grether et al. 2023). Nevertheless, interspecific
308 fighting over mating territories seems unlikely to be a strong mechanism of competitive
309 exclusion in this system because there is no parental care, and males do not provide females with
310 resources. In some other groups of territorial damselflies, females have to copulate with the
311 resident male to use the oviposition sites in his territory, but that is not the case in *Hetaerina*

312 (Grether 1996; Anderson and Grether 2011; Córdoba-Aguilar and González-Tokman 2014).
313 Thus, while interspecific fighting reduces male fitness (Drury and Grether 2014), and species
314 that differ more in microhabitat fight with each other at lower rates (McEachin et al. 2021), it
315 seems very unlikely that the correlation between species differences in microhabitat and
316 heterospecific aggression is the product of differential extinction. By comparison, the hypothesis
317 that the geographic pattern was caused by divergent selection on microhabitat preferences (i.e.,
318 ACD) is quite plausible.

319 The microhabitat differences that have been documented in rubyspot damselflies can be
320 understood to reduce interspecific interference by reducing spatial proximity, either horizontally
321 between different sections or sides of a river (i.e., canopy cover, current speed and stream width)
322 or vertically (i.e., perch height) (Anderson and Grether 2011; McEachin et al. 2021). The species
323 pairs on which we carried out removal experiments tend to perch at different heights and in areas
324 with different amounts of canopy cover (Figure 3). These relatively small differences in habitat
325 preferences could potentially be modified rapidly by selection if the competitive environment
326 changed. For example, if a newly arriving species' canopy cover preference was intermediate
327 relative to the species already present, selection against interspecific fighting might merely
328 reinforce (i.e., narrow) the other species' canopy cover preferences; conversely the
329 disappearance of one species might allow the remaining species' habitat preferences to broaden
330 within their existing environmental tolerances. More generally, however, shifts in habitat use
331 could change the abiotic environment experienced by the animals in ways that cause other types
332 of traits, such as thermal thresholds and morphological traits, to diverge between species (Wcislo
333 1989; Stuart et al. 2014). In addition to obscuring the original mechanism of divergence,

334 adaptation to the abiotic environment could make habitat preferences harder for selection to
335 modify if the competitive environment changes in the future.

336 It is worth emphasizing again that competitive displacement and character displacement
337 in habitat preferences are not mutually exclusive processes. The European flycatchers *Ficedula*
338 *hypoleuca* and *F. albicollis* offer a possible example of competitive displacement leading to
339 evolutionary divergence in habitat preferences. In allopatry both species prefer to breed, reach
340 higher population densities, and achieve higher reproductive success in deciduous forest than
341 coniferous forest (Qvarnström et al. 2009; Veen et al. 2010). In sympatry males of the two
342 species compete for breeding territories and *F. albicollis* usually prevails (Veen et al. 2010).
343 Since arriving on two Baltic islands < 160 years ago, *F. albicollis* males have been outcompeting
344 *F. hypoleuca* males for territories in deciduous forest and forcing them to establish territories in
345 lower quality mixed forest and coniferous forest (Qvarnström et al. 2009; Vallin et al. 2012a;
346 Rybinski et al. 2016). The same type of habitat partitioning occurs in an older (post-Pleistocene)
347 contact zone in Central Europe, but aviary experiments showed that *H. hypoleuca* in the old
348 contact zone actually prefer coniferous vegetation (Adamík and Bureš 2007). Competitive
349 displacement presumably occurred when the species first came into contact in Central Europe,
350 just as it is occurring now on the Baltic islands. Thus, the shift in the habitat preference of *H.*
351 *hypoleuca* seems likely to be an evolved response to competitive displacement in the past (i.e.,
352 ACD) but other possible mechanisms have not been ruled out (Qvarnström et al. 2009; Vallin et
353 al. 2012b, 2012a).

354

355 **Concluding thoughts**

356 Interspecific aggression is common in many animal groups and could be a major cause of
357 divergence in habitat preferences (Ebersole 1985; Linnell and Strand 2000; Grether et al. 2009,
358 2017; Rybinski et al. 2016; Jones et al. 2020), but as yet there are few compelling examples. We
359 offer our evidence that interference competition has contributed to species differences in habitat
360 preferences in rubyspot damselflies as one of the clearest examples to date. We also hope this
361 paper serves as a reminder that ignoring the ghost of competition past does not make it
362 disappear. Competition experiments with negative results should only be taken as evidence that
363 competition is not occurring in the present.

364

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374

375 **Data and Code Accessibility:** The data and code are archived in a public repository
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377

378 **Author Contribution Statement:** SM and GFG designed the study. GFG obtained funding and
379 permits and organized the fieldwork. SM carried out the experiments, the initial data analysis,
380 and wrote the first draft of the manuscript under GFG's guidance. GFG reanalyzed the data and
381 rewrote the manuscript prior to submission. JPD helped with data collection and with the final
382 revision of the manuscript.

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582

584 **Table 1** Study sites, species pairs, time periods, and number of territory holders outside and
 585 inside removal sections.
 586

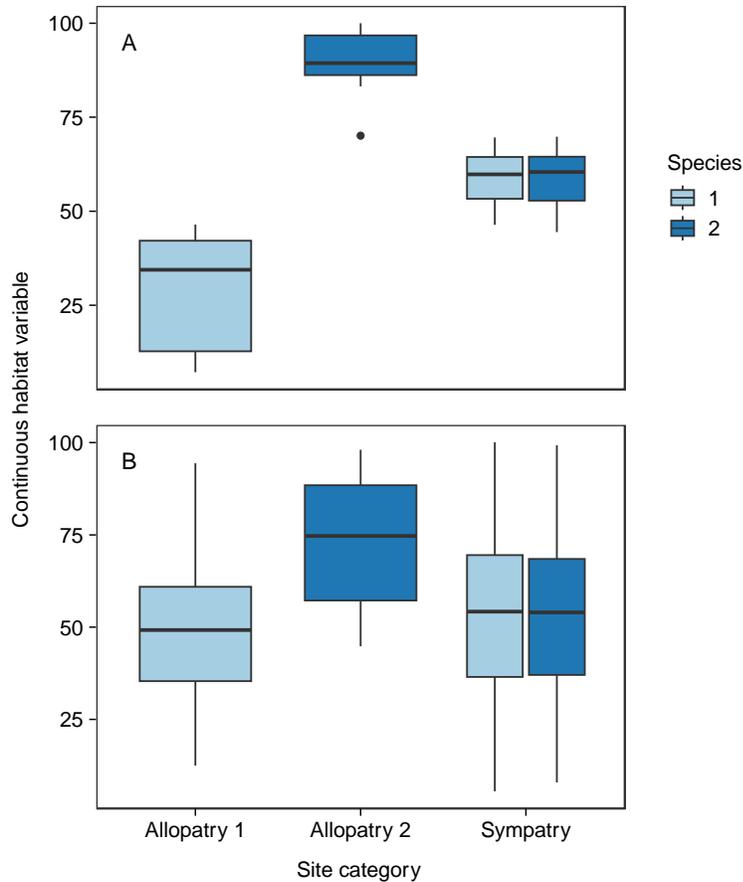
Site	Latitude Longitude	Species	Month Year	No. of territory holders			
				Outside before removal	Outside after removal	Inside before removal	Inside after removal
GO01	8.643	<i>H. occisa</i>	April	12	13	18	12
(4)	-83.1953	<i>H. fuscoguttata</i>	2017	8	11	22	8
MV05	10.278	<i>H. cruentata</i>	April	20	28	20	13
(6)	-84.8189	<i>H. capitalis</i>	2016	20	25	21	5
RT02	10.949	<i>H. occisa</i>	May	33	43	45	18
(5)	-85.5116	<i>H. capitalis</i>	2016	24	32	65	20

587 The number of removal sections is shown in parentheses. “Outside” refers to parts of the study
 588 transects that were designated as removal sections. As explained in greater detail in the Methods,
 589 all *Hetaerina* spp. territory holders were removed from the removal sections and released outside
 590 the study transect. Two other *Hetaerina* species (*H. caja* and *H. titia*) were present at low
 591 densities at site GO01; no other *Hetaerina* species were present at the other sites.
 592

593 **Table 2.** Mean Euclidean distances in microhabitat space between conspecific and heterospecific
 594 territory holders pre- and post-removal.

Site	Period	Conspecific		Heterospecific		<i>n</i>	<i>p</i>
		Mean	SE	Mean	SE		
GO01	Pre	2.44	0.05	2.81	0.06	60	<0.0001
	Post	2.11	0.08	3.04	0.08	20	<0.0001
MV05	Pre	2.42	0.05	2.84	0.07	81	<0.0001
	Post	2.26	0.08	2.74	0.14	18	0.0056
RT02	Pre	2.51	0.05	2.61	0.06	167	<0.0001
	Post	2.37	0.10	2.36	0.10	38	0.94

595 The tabled means are the means of the values plotted in Figure 2, each of
 596 which is the mean distance between one territory holder and the other
 597 conspecific or heterospecific territory holders present in the same period. Only
 598 territories in removal sections were included in post-removal comparisons. The
 599 *p*-values are from paired Wilcoxon paired tests comparing conspecific and
 600 heterospecific distances; *n* is the number of territory holders.
 601



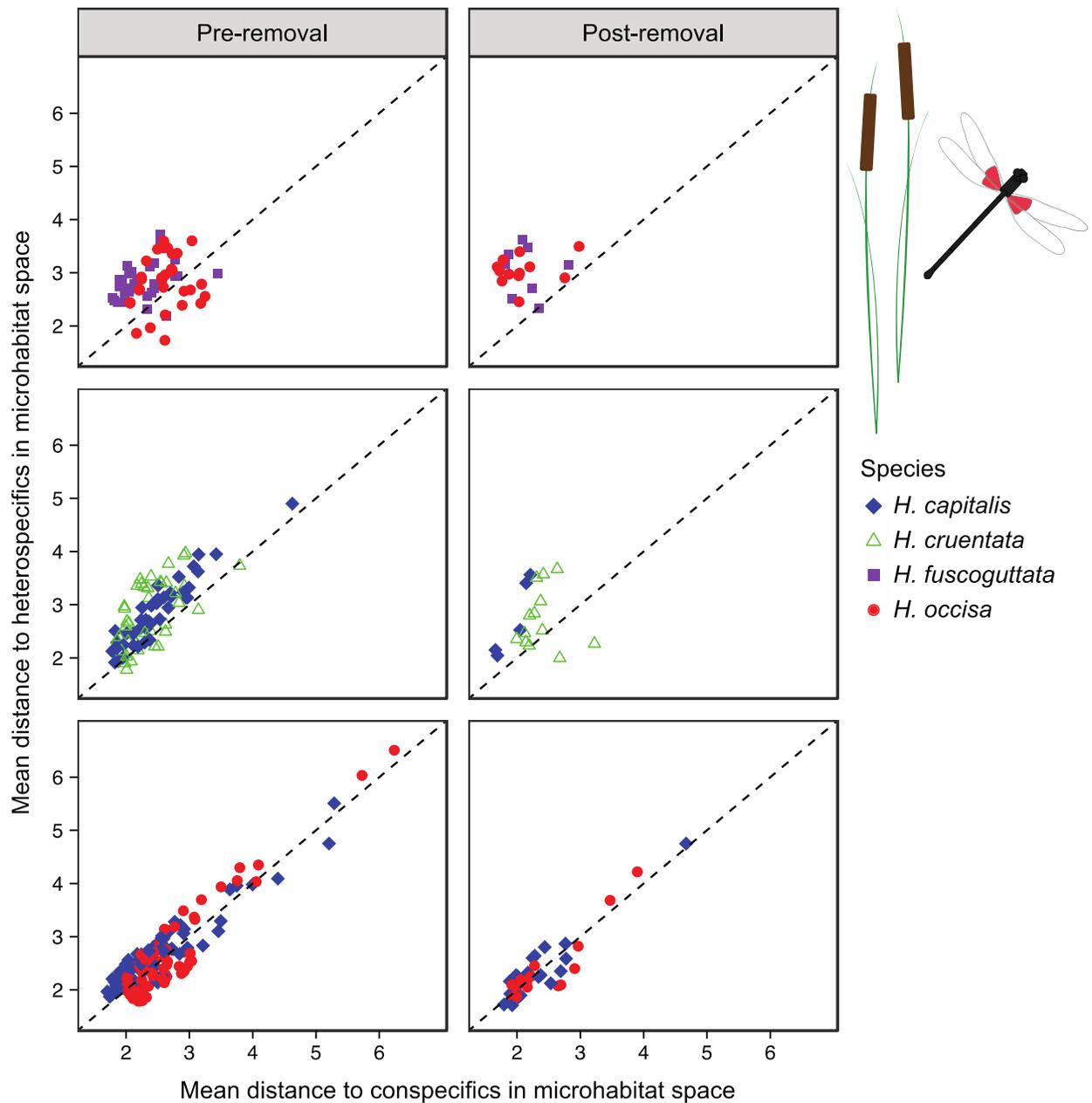
602

603 **Figure 1.** Monte Carlo simulations illustrating that sympatry-allopatriy comparisons of habitat
 604 use are biased toward finding greater differences in allopatriy than sympatry. Two extreme
 605 situations were modeled. (A) Linear habitat gradient with areas of allopatriy around a contact
 606 zone in intermediate habitat. In this simulation, sites were assigned to allopatriy-sympatry
 607 categories so as to mimic the common situation in which species ranges overlap in transitional
 608 habitat. (B) Irregular habitat gradient or mosaic with no correlation between species and habitat.
 609 In this simulation, sites were assigned to allopatriy-sympatry categories at random to mimic a
 610 situation in which species ranges are independent of habitat. In both simulations, 200 individuals
 611 settled at random into different cells (representing territories) at each of 30 sites. The sites varied
 612 in the mean of a continuous habitat variable (site means were randomly drawn from a uniform
 613 distribution with values ranging from 5 to 100); within sites, habitat heterogeneity was normally

614 distributed around the site mean with a standard deviation of 10 (to represent within-site
615 variation in habitat among territories). Prior to settlement, sites were sorted from low to high
616 mean values of the habitat variable for simulation A and left unsorted for simulation B. After
617 settlement, the first 10 sites had only species 1 (Allopatry 1), the next 10 sites had both species in
618 equal proportions (Sympatry), and the last ten sites only had species 2 (Allopatry 2). The same
619 simulated dataset was used for both simulations; only the order of the sites, and hence the
620 sympatry-allopatry categories assigned to the sites, differed. Boxplots depict the median
621 (horizontal line within the box), interquartile range (box), and lower and upper adjacent values
622 (whiskers). Simulations were run in R version 4.2.2 (The R Foundation for Statistical
623 Computing). The simulation code is available for download (see Data and Code Accessibility).

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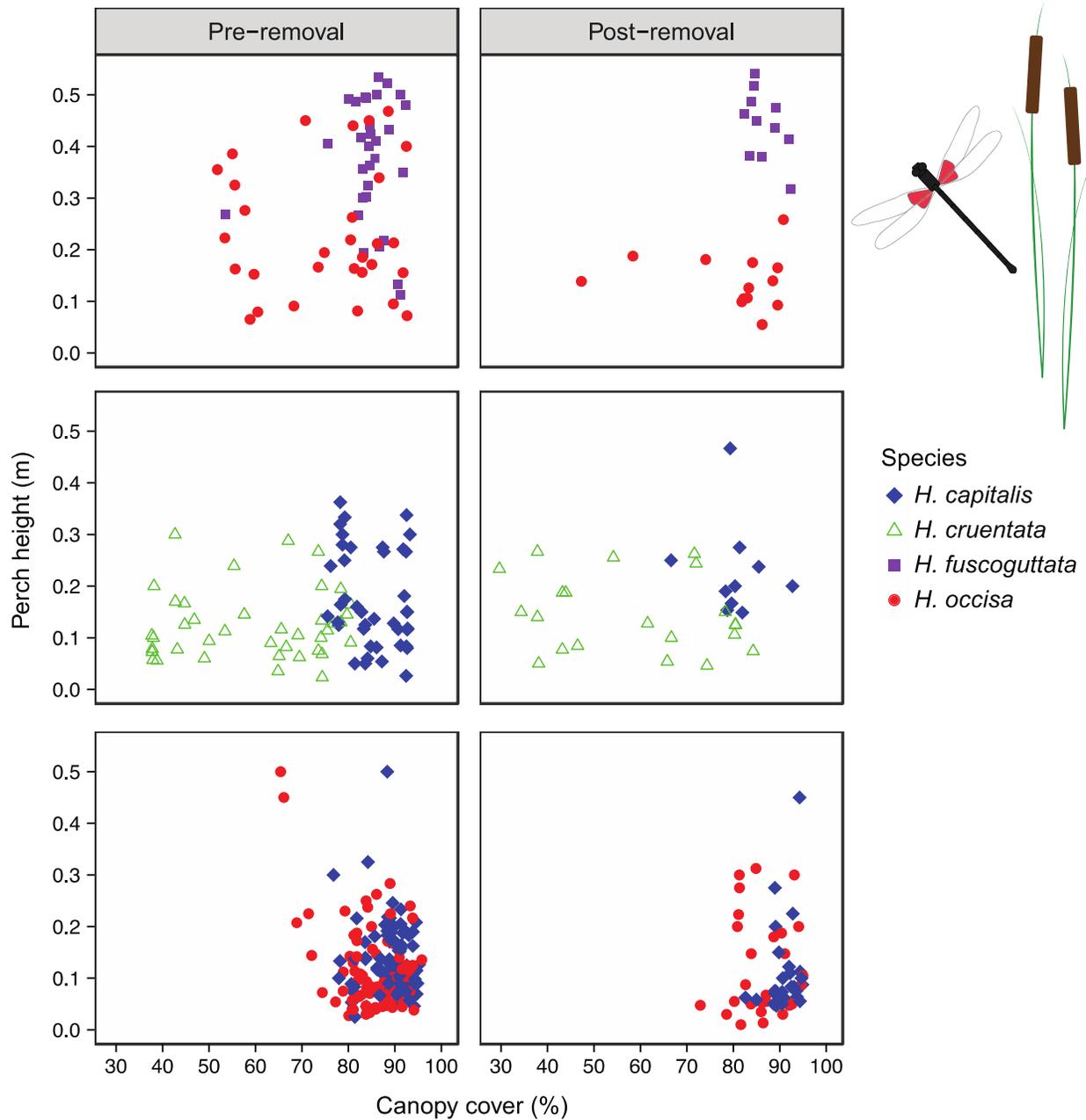
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626

627 **Figure 2.** Distances between conspecific and heterospecific territory holders in microhabitat
 628 space at each of three sites, before and after experimental removals. Each row of panels
 629 represents a different site. Each point represents one territory holder's mean distance to
 630 conspecific and heterospecific territory holders. Dashed lines indicate equal distances to
 631 conspecific and heterospecific territory holders. Points above the dashed lines are cases in which

632 the male's mean distance to heterospecific territory holders was greater than his mean distance to
633 conspecific territory holders. Thus, a preponderance of points above the dashed line indicates
634 that the species differ in microhabitat (for statistical tests, see Table 2). Only territories in
635 removal sections are included in the post-removal panels. For information about the study sites,
636 see Table 1. Sites are in the same vertical order in all figures and tables.
637



638

639 **Figure 3.** Scatterplots showing variation in territory microhabitat of two species at each of three

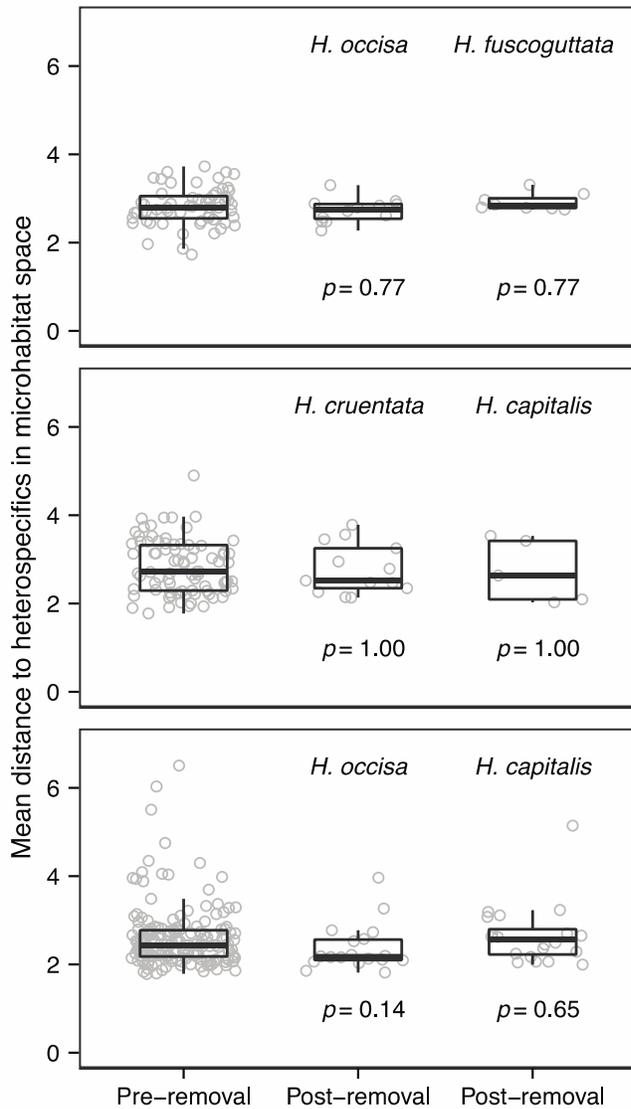
640 sites, before and after experimental removals. Each row of panels represents a different site.

641 Each point represents one territory holder's mean perch height and the mean canopy cover on his

642 territory. Only territories in removal sections are included in the post-removal panels. For

643 scatterplots of the other two microhabitat variables, see Figure S2. For information about the
644 study sites, see Table 1. Sites are in the same vertical order in all figures and tables.

645



646

647 **Figure 4.** Boxplots illustrating that none of the species shifted toward the other species'

648 microhabitat after the experimental removals. Each panel represents a different site. Each point

649 (gray circle) represents one territory holder's mean distance in microhabitat space to territory

650 holders of the other species that were present before the experimental removals. The pre-removal

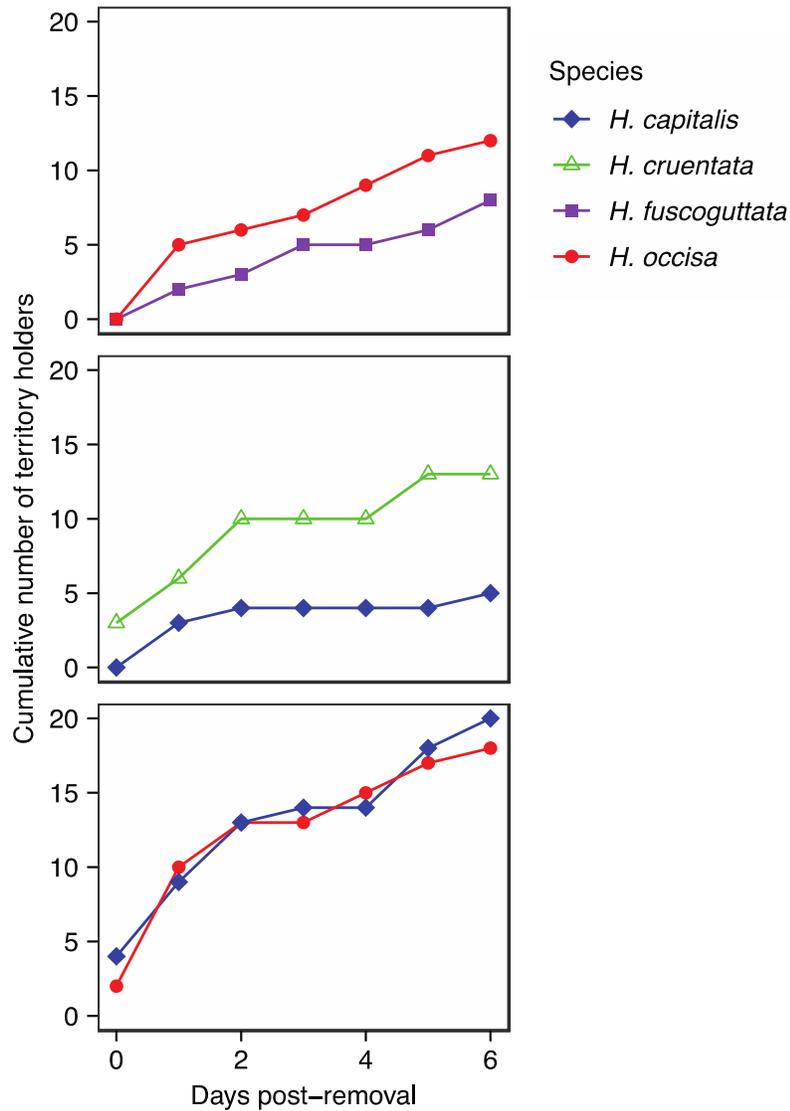
651 groups include all territory holders that were present before the removals. The post-removal

652 groups only include males that established territories in removal sections after the removals.

653 Wilcoxon tests were used to compare post-removal groups to pre-removal groups. The *p*-values

654 were corrected for multiple tests (within sites) using Holm's method. Boxplots depict the median
655 (horizontal line within the box), interquartile range (box), and lower and upper adjacent values
656 (whiskers). Points were jittered horizontally to reduce overlap. Sites are in the same vertical
657 order in all figures and tables.

658



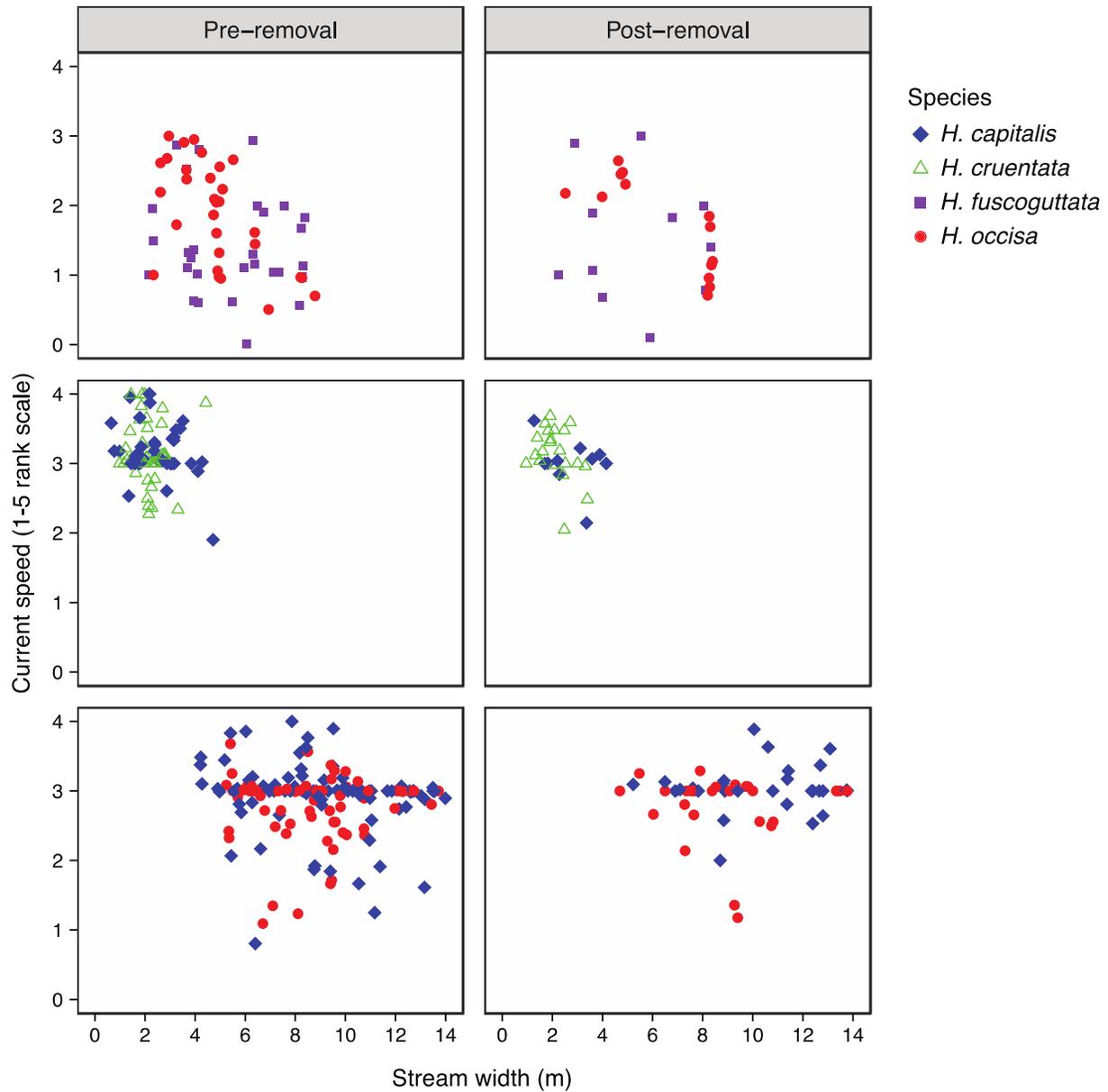
660

661 **Figure S1.** Post-removal territory settlement. Each panel represents a different site. Each point
 662 represents the cumulative number of new territory holders of each species on each day following
 663 experimental removals at three sites. Only territories in removal sections are included. For
 664 information about the study sites, see Table 1. Sites are in the same vertical order in all figures
 665 and tables.

666

667

668



669

670 **Figure S2.** Variation in territory microhabitat of two species at each of three sites, before and
 671 after experimental removals. Each row of panels represents a different site. Each point represents
 672 the mean current speed and stream width on one male's territory. Only territories in removal
 673 sections are included in the post removal panels. For information about the study sites, see Table
 674 1. Sites are in the same vertical order in all figures and tables.

675

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