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Estimating the oxidative ratio of UK peats and agricultural soils
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     Abstract
     Organic matter in the terrestrial biosphere has a fundamental role in moderating the exchange
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     of CO_2 between the atmosphere and the biosphere. One important property of organic matter
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     is its oxidative ratio (OR); that is the ratio of moles O_2 released per mole CO_2 sequestered
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     through photosynthesis i.e. the lower the OR, less O2 is released per mole of CO2 fixed. In
     global assessments of CO_2 partitioning, the failure to account for changes in OR could lead to
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     an underestimate of terrestrial carbon sequestration. It is known that OR can vary between
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     environments and management, but what other factors could be playing a role in controlling
     OR?
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             This study measured the OR of a range of peat (Histosols) and mineral soils
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(Inceptisols) under similar management from the across the United Kingdom to investigate how OR varies within and between material types. The study shows that OR values varied significantly between material types (median peat OR = 1.10, median vegetation OR = 1.03

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and median mineral soil OR = 1.14) and they also varied between study sites. Furthermore there were no significant differences in OR with peat depth.

Given the results from this study we can suggest that future sampling strategies should include sampling of the major carbon pools (i.e. vegetation, litter and soil) and that, as a first approximation, OR can be examined on the basis of these carbon pools alone. The values measured in the study give a new residence time weighted global OR estimate for the terrestrial biosphere (OR_{terra}^{global}) of 1.056 ± 0.02.

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31 Keywords

32 Terrestrial carbon cycle; elemental composition; organic matter; peatlands;

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34 1. Introduction

In 2011 anthropogenic emissions of CO₂ reached 9.5 \pm 0.5 PgC yr⁻¹ (Le Quéré et al., 2013) 35 36 and in May 2013 atmospheric concentrations of CO_2 surpassed 400 ppm for the first time in human history (Jones, 2013). By U using a carbon budget approach (e.g. Le Quéré et al., 37 2013), and assessing changes in atmospheric O2 and CO2 concentrations (Keeling et al., 38 1996), atmospheric carbon is can be partitioned between the carbon pools in the atmosphere, 39 oceans and terrestrial biosphere (residual sink). Battle et al. (2000) used changes in 40 atmospheric levels of O2 and N2 in order to calculate the sizes of annual sinks between 41 reservoirs and for the CO₂ flux to the land proposed the following equation: 42

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$$f_{land} = -\frac{OR_{ff}}{OR_{terra}} f_{fuel} + \frac{1}{(4.8 \times 0.471 \times OR_{terra})} \frac{d\binom{O_2}{N_2}}{dt}$$
 Equation

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46 where: $f_{\text{fuel}} = \text{flux of CO}_2$ due to fossil fuel combustion; $\frac{d\left(\frac{O_2}{N_2}\right)}{dt}$ is rate of change of molar ratio 47 of atmospheric O₂ and N₂; OR_{ff} is <u>_</u> the combustion stoichiometry; OR_{terra} is <u>_</u> the oxidative 48 ratio of the terrestrial biosphere.

Within equation 1 there are two terms that play an important role in understanding the 49 partitioning of atmospheric carbon, the oxidative ratio (OR) which is the molar ratio of O_2 50 and CO_2 fluxes associated with either fossil fuel combustion (OR_{ff}) or photosynthesis 51 (OR_{terra}). This ratio is an important component in equation 1 and so research has considered 52 the relative changes in O_2 and CO_2 in the atmosphere as a means of understanding the relative 53 magnitude of global sinks and sources of CO2 (Keeling and Shertz, 1992; Keeling et al., 54 1996). Until recently there has only been one estimate of OR for the terrestrial biosphere 55 (OR_{terra}) that has been used with equation 1. This and that_came from Severinghaus (1995) 56 57 who estimated the value to be 1.1. This value of 1.1 has been commonly adopted through in several global studies (e.g. IPCC, 2007). However, more recent work has shown that this 58 value may not be appropriate for equation 1. In a recent meta-analysis of global OR values, 59 Worrall et al. (2013) showed that, whilst within the range of natural occurrence, the 60 61 commonly used value of 1.1 is probably not the most accurate value. Worrall et al.- (2013) showed suggested that 1.04 ± 0.03 was a more appropriate choice and that adopting this value 62 63 meant that the sink of carbon to the land has been underestimated by up to 14%.

Direct atmospheric measurement of OR is possible through simultaneous measurements of atmospheric O_2 and CO_2 (e.g. Seibt et al., 2004), however several technical challenges have been noted with this method (see discussion in Masiello et al., 2008). An alternative and complementary approach is to directly measure the OR of biomass pools in the terrestrial biosphere. Masiello et al. (2008) detail the mathematical linkage between the oxidation state of organic carbon (C_{ox}) and OR and another fundamental properties of the carbon cycle that of the oxidation state of organic carbon (C_{ex}). It is possible to calculate C_{ox} , and therefore OR, for the carbon pools (e.g. aboveground biomass) of an ecosystem using
elemental analysis of %C, %H, %N and %O.

73 Worrall et al. (2013) based their global estimate of OR on a weighted average of OR values for different soil orders and global-scale biomes. The assumption was -but in so doing 74 assumed that the major control on differences in OR was indeed the differences between 75 carbon pools (e.g. soil vs. vegetation) and between soil orders and between vegetation 76 biomes, i.e. that the greatest control on OR variation was a difference between, for example, 77 78 Inceptisols and Mollisols, or between savannah and boreal forest. This assumption was a necessity arising from the limited amount of data available for individual environments 79 across the globe. Additionally no studies were included where both soil and vegetation were 80 81 analysed for the same site were included and given the limited number of studies targeted at OR this assumption could not be tested. 82

Therefore, t<u>T</u>his study aims to assess and understand the variation in OR between two biomes and two soil orders across one country. Thus in this study w<u>W</u>e aim to test whether the assumption that OR is controlled by differences between soil orders and biomes is true by assessing the magnitude of variation between soil orders and vegetation biomes_types_in comparison to other possible sources of variation. In this study we will consider the variation between organic matter types in comparison to the variation within a soil order and the site at which we find a soil and its associated vegetation.

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91 2. Materials and Methods

92 The approach of this study was to consider the variation in two soil types – peat soils
93 (Histosols) and minerals soils (Inceptisols) across a north-south transect through the United
94 Kingdom. For each site all the possible organic matter types were sampled with the view of

comparing the between site, between organic matter types and within site variation in thevalues of OR.

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98 2.1. Sampling sites and methodology

InDuring the summer of 2011 peat cores were taken from 8-eight sites across a elimatic 99 gradient through the UK (Figure 1; Table 1). Within each peat site the actual location of 100 sampling was chosen as being the least disturbed location available i.e. little or no visible 101 102 land management in recent years. At each sampling location two peat cores of up to 1_m 103 depth were taken using a 70 mm diameter gouge auger. Each core was subdivided into 50 104 mm sections in the field and placed into sealed plastic sample bags. - In addition to collecting two peat cores at each site, Rrepresentative samples of dominant vegetation types and surface 105 litter were also collected at each site. The exact vegetation composition varied amongst sites 106 107 (Table 1) but typically at each site the following were sampled: mosses (e.g. *Sphagnum* spp.); 108 sedges (e.g. Eriophorum spp.) and shrubs (e.g. Calluna vulgaris). The vegetation data were 109 split into seven functional groups - shrubs, grasses, sedges, Sphagnum mosses, non-Sphagnum mosses, cropland vegetation, and litter. 110

111 To act as a comparative sample fE or each peat site, two nearby locations not on peat 112 soils_were also selected for sampling – both-locations were on mineral soils, but one was chosen under arable and the other under pasture land use. For these comparator sites, soils, 113 litter, and vegetation were sampled. In mineral soils no profile samples were taken because 114 115 of the paucity of organic carbon at depth in most mineral soils but soil samples were taken from the upper 100 mm-using a trowel, whilst litter and vegetation were sampled in the same 116 way as for the peatland sites. For the purposes of analysis and reporting, Tthe vegetation data 117 were split into seven functional groups shrubs, grasses, sedges, Sphagnum mosses, non-118 Sphagnum mosses, cropland vegetation, and litter. 119

120	All samples were dried at 105°C for 48 hours prior to further analysis. Peat Bbulk
121	density was then calculated on a dry weight basis using the volume of the core section and
122	mass of dry soil solids. The mineral soils were pre-treated using a 2% HF acid solution based
123	on the methods of Mathers et al. (2002) and Skjemstad et al. (1994). Approximately 5g of
124	mineral soils were treated with five 50mL aliquots of 2% HF acid and shaken. Supernatants
125	were centrifuged and decanted between treatments. Soils were then rinsed with deionised
126	water at least 3 times and then dried at 75°C. HF-treated soils, peat, litter and vegetation
127	samples were all ground using a Spex 6770 Cyromill.

For comparative purposes, three standard, naturally-occurring organic materials were considered: lignin, humic acid, and cellulose. The lignin and humic acid were from supplied by Aldrich and the cellulose was taken from ash-free paper. The standards were analysed for their elemental composition (C, H, N and O) and their energy content (gross heat value, ΔH_e).

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133 2.1.1. CHNO Analysis

All samples were analysed for CHNO elemental content. Samples were analysed for their 134 135 carbon, hydrogen and nitrogen (CHN) and oxygen (O) concentration on a Costech ECS 4010 Elemental combustion system with pneumatic autosampler. Oxygen concentrations were 136 analysed on a separate set up to the CHN set up. It was set up for CHN analysis where 137 Reactor 1 consisted of chromium (III) oxide/Silvered cobaltous cobaltic oxide catalysts @ 138 950°C and Reactor 2 consisted of reduced high purity copper wires @ 650°C. Helium was 139 used as the carrier gas at a flow rate of 95 ml min⁴. This was filtered for hydrocarbons 140 upstream of the instrument. A packed 3m GC column was used for separation of the gases. 141 A thermal conductivity detector (TCD) was used to calculate the signal of each sample. For 142 oxygen (O) concentration, the Costech ECS was also used but was set up for O analysis. 143 consisted of a nickelised carbon/ silica chips/nickel wool pyrolysis tube @ 1060°C 144

whilst Reactor 2 was left empty. Helium was used as the carrier gas at a flow rate of 130 ml
min⁻¹-but no oxygen was used. A 2m packed oxygen GC column was used for separation of
the gases. Chloropentane vapour was added to the carrier gas to enhance decomposition of
the oxygen compounds and to reduce possible memory effects from previous samples
(Kirsten, 1977).

150Computer software used was EAS Clarity (DataApex Ltd, Prague, Czech Republic). For-151both CHN and O setups a calibration curve of $r^2 > 0.999$ were created using acetanilide as the152standard. Samples of acetanilide were included within each run as unknown samples to act as153internal quality control checks. Each sample (peat, soil or vegetation) was analysed in154triplicate... i.e. three times on the CHN setup and a further three times on O set up, and a mean155calculated for C, H, N and O.

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157 2.1.2. Calorimetry

158 Energy content, as gross heat value (ΔH_c), was measured for all peat, vegetation and litter samples. Masiello et al. (2008) has shown that it is possible to derived Cox values (and 159 160 therefore OR values) from calorimetry data. Analysis was performed on a 6200 Isoperibol Calorimeter (0.1% Precision Classification, Parr Instrument Company, Illinois, USA) with 161 1108(P) Oxygen Bomb. Calibration was performed as a rolling average of 10 measurements 162 using benzoic acid standards. Samples were placed in crucibles and compressed to stabilise 163 the peat surface and weighed following compression, with a weight of approximately 0.8 g 164 used. Where sample amount was deficient, a benzoic acid spike was used. Following 165 analysis, fuse corrections were performed by measuring the length of fuse wire remaining, 166 measured in calories and converting to MJ/Kg. The difference was taken away from the 167 energy content recorded during analysis. Limited organic matter contentsample meant that 168 169 gross heat values could not be calculated for mineral soils.

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171 2.1.3. C_{ox} and oxidative ratio (OR) calculation

172 A value of OR can be calculated from the carbon oxidation state (C_{ox}) which in turn can be 173 calculated from elemental compositions of organic matter as follows (Masiello et al., 2008): 174

$$C_{OX} = \frac{2[O] - [H] + 3[N]}{[C]}$$
 Equation 2

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Where: [X] = molar concentration of C, H, N or O, and assuming the majority of organic
nitrogen exists as amine groups in amino acids. Furthermore, sulphur is not included in this
equation as it is assumed to form < 0.25% of biomass (Charlson et al., 2000).

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As C_{ox} and OR are related through the balancing of organic matter synthesis, the OR value is calculated as the ratio of O_2 and CO_2 coefficients (for further details see Masiello et al., 2008). (for further details see Masiello et al. (2008)). Simplified, it is then calculated as:

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$$OR = 1 - \frac{c_{ox}}{4} + \frac{3[N]}{4[C]}$$
 Equation 3

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Equation 3 assumes that there is no contribution to the C_{ox} from S or P, and it has been shown that the error in the OR of making such an assumption would be only ± 0.002 (Hockaday et al., 2009). This equation also assumes that the nitrogen source in carbon fixation is N₂. There are two further possible nitrogen (N) conversions (Masiello et al., 2008):



195For the purposes of this paper, Equation 3 was used as N_2 is the dominant form in the196peatland ecosystem. The agricultural soils will likely received N in other forms in addition to197 N_2 , but no data were available for these sites and other studies have shown minimal changes198in OR when using alternative assumptions for the reason that [N]/[C] is always likely to be <</td>1990.1-(Gallagher et al., in review).

As a quality control check, OR values were only calculated for those samples that had measured data for C, H, N and O; if one of these data were missing (e.g. lost sample), no OR value was calculated.

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204 2.2. Statistical analysis

205 2.3. Analysis of Variance (ANOVA)

206 The design of the study allows for a several statistical comparisons to be made using an 207 ANOVA approach. Firstly, one-way ANOVA was used to test whether there were 208 significant differences within the organic matter types being considered. This set of ANOVA 209 could be sub-divided into several separate ANOVA: the difference in OR for mineral soils 210 under arable and under pasture; the difference in OR between vegetation functional groups 211 where the factor levels were: shrubs, grasses, sedges, Sphagnum mosses, non-Sphagnum mosses, cropland vegetation, and litter. The second set of ANOVA that could be performed 212 was for the peat soils only. A two-way ANOVA was used to determine the statistical 213 214 significance of the factors – site and peat depth. Finally, the variation in organic matter types between sites was examined. All types of the organic matter types were considered and all 215 216 sites but for comparison with the mineral soil samples, only the surface samples of peat soils from each site were included (depth < 20 cm). Further, for reasons of cross-classification 217

between the factor levels the depth in the peat profile, the vegetation functional group, andthe land use of the mineral soil were not considered as a separate factor.

220 Response variables used were energy content, Cox, and OR. If necessary tThe 221 response variables were log-transformed to ensure data normality prior to ANOVA. tested for normality prior to ANOVA using the Anderson Darling test; if the response variable failed 222 the test it was log-transformed and re-tested. Further transformation to ensure normality of 223 the response variable did not prove necessary. The Post hoc testing of the results was 224 225 performed using the Tukey test at 95% level was used to determine significant differences between levels of any factor. The magnitude of the effects of each significant factor and 226 interaction were calculated using the generalized ω^2 (Olejnik and Algina, 2003). Response 227 228 variables used were: C/N ratio, H/C ratio, O/C ratio, energy content, Cox, and OR.

A power analysis was used to assess the minimum effect size that could be detected within this latter comparison of organic matter type and site. The study was fully factorial with respect to each of 2 factors, 3 centre points were assumed; the standard deviation was estimated as the square root of the mean square difference; and the required experimental power was set at 80%.

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235 2.4. OR and ΔH_c

236 Masiello et al. (2008) use<u>d</u> ΔH_c from calorimetry to calculate C_{ox} values for a range of 237 standard materials. If then there is a relationship between C_{ox} and ΔH_c it might be reasonable 238 to expect there to be a relationship between ΔH_c and OR values from this study. If so it may 239 be able to shed some light on underlying mechanisms leading to OR variation—and also 240 provide a simpler method of calculating OR for many materials. Therefore, ΔH_c values were 241 plotted against OR values for the peat soils and vegetation along-together with the standard **Comment [F1]:** 80% is like 95% in stats test – I SHALL TRY AND FIND A REFERENCE OTHERWISE I WILL USE A FOOTNOTE

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materials – cellulose, lignin and humic acid. Mineral soils were excluded from this analysis as no ΔH_c values could be calculated.

- 244
- 245 2.5. Global OR values

The data from this survey can be used to update the estimation of global OR made by Worrall et al. (2013). Worrall et al. (2013) have proposed a weighted average <u>approach:based upon</u> the residence time of carbon in vegetation and in soil as this gives a greater importance to the faster turnover of carbon in the vegetation pool. Therefore:

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$$OR_{terra}^{global} = f_{soil}^{terra} OR_{soil}^{global} + f_{veg}^{terra} OR_{veg}^{global}$$
Equation 64

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253 Where: OR_{terra}^{global} = the oxidative ratio of the global terrestrial biosphere; OR_{soil}^{global} = the 254 oxidative ratio of the global soils; OR_{veg}^{global} = the oxidative ratio of the global vegetation; 255 f_{soil}^{terra} = the proportion of the terrestrial biosphere C annual flux that is due to soils; and 256 f_{veg}^{terra} = the proportion of the terrestrial biosphere C annual flux that is due to vegetation.

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The annual flux from the soils or vegetation was based upon the size of reservoir divided by 258 the average residence time. The comparative sizes of the soil and vegetation reservoirs was 259 were estimated from Eswaran et al. (1993) and Olson et al. (2001), where the proportion of 260 261 carbon in the vegetation reservoir was 0.28 and in the soil reservoir as-0.72. The average 262 carbon residence time for soils was taken as between 20 and 40 years based upon a study by Jenkinson and Rayner (1977). Mills et al. -(2014)(2014)- examined radiocarbon results for 263 133 UK₇ soils and found that the carbon turnover was best modelled as two pools – a fast 264 265 pool with 20 year residence time and a slow, 1000 year turnover pool. -The average carbon 266 residence time for vegetation was taken as between 2 and 5 years (e.g. Gaudinski et al.,

2000). We recognize that the OR of soil fluxes and soil pools may not be identical (just as the 267 268 carbon isotopic values of the bulk soil carbon pool rarely match the carbon isotopic values of 269 the soil CO_2 flux). However, we must start with the assumption of equivalence between soil 270 OR flux values and OR pool values because no data yet exist comparing soil pool and flux <u>OR values</u>. Given the above approach the values of $f_{soil}^{terra} = 0.27$ and $f_{veg}^{terra} = 0.73$. The 271 value of OR_{veg}^{global} is the weighted average of the expected values of the 16 global biomes 272 where the weighting is the area of each biome (Loveland and Belward, 1997).__and_OR_soil 273 is the weighted average of the expected values of each of the soil orders of USDA soil 274 taxonomy where the weighting is the organic carbon content of each order (Eswaran et al., 275 276 1993). Since in both cases the data for any one biome or soil order is are scarce, then the expected value is taken as the median of each biome or soil order. 277

Worrall et al. (2013) applied Equation <u>6-4</u> based upon the classifying organic material samples into one of either the 12 USDA soil orders (although Gelisols and Histosols were combined into on<u>ce because of lack of data</u>) or 16 global biomes (Loveland and Belward, 1997). The soil samples collected as part of this study were classified as either Inceptisols or Histosols. The vegetation samples can be classified <u>asin one of the following</u>: Grassland, Cropland, Shrubland or Permanent wetland.

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285 **3. Results**

Overall OR could be calculated for 251 peat samples, 49 vegetation samples and 14 mineral soil samples. Table 2 shows the data for the individual elemental concentrations for each of the material types whilst Tables 3 and 4 show the parameters for vegetation and peat soils respectively.__and tThe composition of the standard materials , naturally occurring organic types are listed in Table 5. None of the datasets needed to be transformed prior to ANOVA. Of the three material types, vegetation samples had the lowest OR values followed by peat soils; mineral soils showed the highest OR values (Table 2). The values for mineral soils are within the range of previously reported values (Hockaday et al., 2009) though the peat and vegetation means were lower than 1.1 though-but still within the range of results reported by Worrall et al. (2013).

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297 *3.1. Variation within organic matter type*

298 Mineral soils

Within the mineral soil dataset it was possible to determine whether there was a significant difference between land uses, i.e. between soils under arable and soils under pasture. Results from the one-way ANOVA shows that there were no significant differences in elemental ratio data (C/N, O/C, and H/C), C_{ox} or OR between mineral soils (n = 14) under different land uses.

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305 *Vegetation types*

306 Within the vegetation data, there were significant (p < 0.05)-differences in all elemental 307 ratios, energy content (p < 0.001), C_{ox} (p < 0.018) and OR (p < 0.001) values between 308 vegetation functional groups (Table 6) from the one-way ANOVA. The post hoc testing showed a great deal of variation in where the significant differences lay. For example, for 309 C/N ratio the only difference lay between non-Sphagnum mosses and arable crops and 310 grasses (Table 6) with grasses and crops having the lower C/N ratio (Table 3). However, for 311 312 other elemental ratios and energy content there were a number of significant differences between functional groups (Table 6). Tthe highest OR values were found in shrubs (e.g. 313 Calluna vulgaris, Erica tetralix) whilst the lowest values were found in mosses, both 314 Sphagnum and non-Sphagnum (Table 3). The reverse of this is-was true for Cox values where 315

the lowest values were found on shrubs and the highest on *Sphagnum*. However, the variation within these functional groups meant that only certain combinations of groups were statistically different from one another (Table 6).

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318

320 Peat Soils

In the ANOVA model, site and peat depth were included but due to a collinearity betweensite and depth, the interaction term could not be plotted.

323 For all the measured parameters, the site factor was found to be significant (Table $\frac{76}{2}$). Post hoc testing of site factor for each parameter showed considerable variation between sites 324 325 (Table 8). For example, with C/N ratios, Auchencorth and Forsinard showed the lowest values whilst the highest values were found on Westhay Moor (Table 4). The extremes for 326 O/C ratios were found between Whixhall and Dartmoor, whilst for H/C ratios the largest 327 328 ratios were found on Forsinard and the lowest on Thorne (Table 4).showed T the lowest 329 energy contents were found at Auchencorth with the highest at Dartmoor (Table 4). There were significant differences between peatland sites in terms of Cox and OR values explaining 330 331 21% and 39% of the variation in the data respectively (Table $\frac{76}{2}$). *Post hoc* testing (Table 8) 332 showed that The lowest OR values gave some similar patterns to the elemental data 333 withwere found at Thorne resulting in the lowest OR values whilst and Forsinard and Bodmin having had the highest OR values. 334

Depth was a significant factor for <u>all parametersbulk density and energy content</u> <u>although not significant for with the exception of C_{ox} and OR (Table 76). In terms of down</u> <u>core profiles generally t There were increases in C/N ratio and energy content down the core</u> <u>and decreases in O/C and H/C ratios with depth (Figure 2). These profiles were as would be</u> <u>expected as the peat becomes more carbon rich with depth. whilst t</u>The observed data for bulk density <u>was-were</u> more complex-<u>and</u>; across all cores the bulk density generally-increased in
the upper 30 centimetres before decreasing with depth (Figure 2).

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343 *3.2. Organic matter type vs. site*

It was possible to analyse the differences between surficial peat (0 - 20 cm), vegetation and 344 345 surface mineral soils across all the sites considered in study for the element ratios and OR. Bulk density and energy content were not considered because neither could be measured for 346 347 all sample types. - but because the analysis covered all the organic matter types being considered then it was not possible to include the bulk density and the energy content. The 348 349 analysis was performed with the caveats that the above analysis found the following 350 significant differences: there were significant differences in the OR between vegetation 351 functional group; there were no significant differences between OR for different land uses on 352 the mineral soils; and there were no significant differences with depth for the OR of peat soils. The power analysis shows that this design was capable of detecting a difference of 0.02 353 354 in the OR at the 80% probability.

355 When comparing all organic matter types and all sites for elemental ratios and OR 356 there were significant differences between the site and organic matter type factors (Table 97). 357 For the OR values the most important of the factors was the difference between organic matter types (explaining 70% of the original variance) and post hoc testing showed that there 358 were significant differences between all the organic matter types considered. The highest OR 359 values given by the mineral soils (1.15 ± 0.01) followed by peat soils (1.079 ± 0.006) with 360 the lowest OR represented by vegetation (1.037 ± 0.007) where the values are given as the 361 362 estimated marginal means (averages accounting for all other factors and covariates) and the uncertainty in each is given as the standard error. The variation between organic matter types 363 was greater than the variation between sites with the site factor explaining only 21% of the 364

original variance. The *post hoc* testing between the individual sites shows that only one site,
<u>Thorne</u>, was significantly different from all the others <u>— Thorne was significantly lower than</u>
all other sites (Figure 3; Table 10). When the data for Thorne were removed then
Auchencorth was found to have OR values significantly higher than all other sites: there were
no other significant differences between any other sites considered in this study.

370 When the elemental ratios were considered then the first observation was that there were no significant differences between the sampled sites for the O/C ratio (Table 9); this 371 372 study could find no evidence that O/C varied across the UK. For both the C/N and H/C ratios, and as for the OR values, the organic matter type factor was more important than the site 373 374 factor (Table 9). Post hoc testing of the elemental ratios shows two distinct patterns. For the 375 C/N ratio the post hoc testing shows that the significant difference between sites was been Westhay (C/N = 43 ± 3) and the samples from both Dartmoor (C/N = 26 ± 3) and Bodmin 376 377 24 ± 4). For the H/C ratio the significant difference between sites lay between Thorne Auchencorth, Bodmin, Dartmoor, Forsinard and Whixall with Thorne having 378 379 significantly lower values. The pattern of the differences in the H/C ratios is distinctly closer 380 to that observed for the OR values than the pattern observed for C/N ratios and thus implying 381 that the difference at the Thorne site was due to the H/C ratios and not due to differences in O 382 or N.

There was no significant interaction between the site and organic matter type factors for any of the parameters considered in this study. The lack of significant interaction between the two factors means that the difference between organic matter types does not vary with site suggesting that Inceptisol is different from a Histosol regardless of the position within the UK and as such there is a fixed relationship between the organic matter types. Regressing the mean vegetation OR and mean bulk peat OR for each of the eight sites shows no significant relationship between them.

391 *3.3. Variation in Organic Matter Composition*

392 A comparison of OR and ΔH_c with respect toin peat soil samples and the vegetation samples show several possible patterns (Figure 4a and b). Given the result of Masiello et al. (2008) it 393 394 would be expected that OR would increase with ΔH_c and this is true for such a relationship is discernible in these plots as the line between the organic material standards (humic acid, 395 cellulose and lignin). With respect to the sampled peat sites some of the soils sampled from 396 397 Thorne, Westhay and the Peak District plot below this line with lower ΔH_c values than would be expected for their OR values (Figure 4a). Conversely, all the samples from all the other 398 399 sites (Auchencorth, Bodmin, Dartmoor, Forsinard and Whixall) plot above this line with OR greater than the equivalent mix of the organic matter standards. Furthermore, it could be 400 401 proposed that majority of samples form a 3 end- member triangle the end-members of which are marked by humic acid, lignin and a high OR end-member represented by peat soil 402 403 samples from Forsinard (Figure 4a). The latter end-member with high OR represents organic 404 matter that is as reduced as the lignin standard but does not have the calorific value. Some 405 soil samples plotted at even lower values of ΔH_c than those from Forsinard and these come from Auchencorth. One possible explanation of this is that the peat soil at Auchencorth has 406 received inputs of mineral matter, had visible evidence of trace mineral matter in the field. 407 Inputs of fine silt or clay into the peat would not alter the measurement of OR as it is based 408 409 on a ratio of elements in the organic matter but it would in effect-dilute out the energy value of any sample.- It it-may also explain large values of bulk density observed for this site 410 which contribute to inflexion in the bulk density profile (Figure 2). 411

With respect to vegetation type mMost vegetation types were better constrained
within the line defined by the organic matter standards (Figure 4b) and the peat samples from
Forsinard than was observed for the peat soils (Figure 4a). The end-member represented by

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the peat soils of Forsinard plots close to the samples of *Erica tetralix*. Some samples of both
grass and *Sphagnum* plot at lower values of OR than would be expected from a combination
of the organic matter standards and this type of plant material which is more oxidised than
any combination of the standards but at similar calorific value.

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420 **4. Discussion**

This study has been able to comment onevaluated the oxidation status of peat (Median OR = 1.10) and agricultural soils (Median OR = 1.14) across a latitudinal transect across the United Kingdom. Significant differences were found in the oxidative ratio of major terrestrial carbon pools (soil and vegetation). This is perhaps not unsurprising given the different processes operating in each carbon pool but this study has been able to quantify the difference with respect to OR.

427 Within-group variation of OR generated a number of interesting results. Firstly, soils under agricultural management did not vary significantly in their OR value. One might expect 428 429 that different management practices, and different vegetation types, would influence soil 430 processes that which would, in turn, affect OR. Even from this study is possible to see the large contrast between cellulose and lignin, and so woody vegetation that is woodier would 431 be expected to have a higher OR than non-woody types. Furthermore, N is a component of 432 433 the oxidation state of an ecosystem (Equation 2) and so diversity of N inputs and N sources 434 (e.g. organic wastes vs. inorganic fertiliser) might be expected to shift the OR of an 435 environment. However, this result is for the ecosystem and not the components of that ecosystem (e.g. soil or vegetation) and the lack of observed difference between grasslands 436 and croplands in this study may reflect totality and not just processes that affect vegetation 437 438 alone.

439	Secondly, OR did not vary significantly with depth and there was not not consistent
440	OR depth profile variation between across-all the sites. Peat depth was an important factor in
441	the elemental ratio datasets and many of these trends can be used to identify structural
442	changes in the peat and to infer changes in the peat decomposition process. The C/N ratio
443	can be used to infer decomposition rates, specifically the loss of mass, whilst the H/C and
444	O/C ratio are commonly used to infer humification rates . Decreases in H/C ratios and O/C
445	ratios are specifically linked to dehydrogenation and decarboxylation respectively. The lack
446	of a-significant OR trends is perhaps one of the more unusual observations from this study.
447	Given the classical explanation of peat formation it would be expected that C_{ox} would decline
448	with depth as the peat profile becomes more anaerobic, and therefore an increase in OR
449	would be expected. The classical explanation of peat soils is that they rapidly become
450	anaerobic-In peat soils anaerobic conditions result from due to excess organic matter and
451	slow ingress of air due to persistent waterlogged conditions, leading to successive use and
452	exhaustion of redox couples (Reddy and D'Angelo, 1994). However, this succession is noted
453	mainly for inorganic <u>chemical</u> species (e.g. Fe(III)) in soil water and for peat soil water
454	concentrations are often low compared to mineral soils giving them very little buffering
455	capacity with respect to redox conditions meaning that species in solutions are rapidly
456	transformed, e.g. Fe(III) to Fe(II). However, the concentration of inorganic redox species that
457	can be reduced is very lowamount of reduction occurring is minimal -in comparison to the
458	amount of organic matter that could be oxidised and so therefore dramatic changes in
459	inorganic soil solution chemistry are not reflected in the substrate.
460	Thirdly, when examining the variation in organic matter composition a number of
461	samples plotted outside of the 3 end member space delineated by humic acid, lignin and the

- 462 high OR end member (Figure 4a). Of particular note is that many of the samples of peat soil
- 463 from the site at Auchencorth plot at values of ΔH_e outside this triangle and this is probably

due to inputs of silt. Furthermore, some of the vegetation samples also plotted at lower than
expected OR values. Whatever component lends low OR values to these grass and *Sphagnum* samples may be part of the explanation of low OR peat samples from the sites at
Thorne, Westhay and the Peak District. Equally, the composition of the proposed third endmember is not known.

By considering natural organic carbon reservoirs in a structured survey, this study has examined the variation in OR and it can assess whether the global assessment of OR proposed by Worrall et al. (2013) was appropriate. Firstly, the study has shown that it is-was possible to distinguish between the naturally-occurring organic matter types (e.g. soil and vegetation) when looking at OR; that is there is-are a-significant differences between organic carbon pools that are easy to sample and model over large scales.

Secondly, Worrall et al. (2013) could only use the data that was were available in the 475 476 published literature so had little control on the soils orders available and their geographic distribution. This study has been able to take one soil order (Histosols) and compare the OR 477 478 across a range of locations under similar management in order to assess the relative source of variation in OR. However, the study can show<u>It has shown</u> that there is more variation 479 480 between carbon pools than between sites but that the variation between carbon pools is independent of the changes between sites, i.e., tThis is good statistical justification to support 481 the approach used in Worrall et al. (2013) that as a first approximation, and at large scales, 482 483 OR is better understood by considering soil and vegetation separately. Thirdly, the study has shown a significant difference between the two soil orders considered by this study (Histosols 484 485 and Inceptisols) which is an underlying assumption of Worrall et al. (2013) and of the calculation of OR_{soil}^{global} . But as discussed above there was no significant difference but two of 486 487 the biomes directly analysed, i.e. grassland and croplands.

Fourthly, we can start to address the question of what kind of sample is truly 488 489 representative of the OR of an environment. Results from the study showed that if OR does 490 not vary with peat depth, then it may be appropriate just to sample the surface peat rather than coring. If peat can be simplified to just surface samples, can the OR of the environment be 491 simplified to just one sample? This study has looked at the vegetation associated with the 492 soils and has shown that vegetation and peat samples were not significantly similar to each 493 other to warrant only one sample being taken. Therefore future sampling should concentrate 494 on sampling both the vegetation and soils carbon pools. 495

Finally, this study is a continuation of earlier work (Worrall et al., 2013) that 496 calculated a global OR value from a database of OR values. and Tthis study is able to add to 497 that database and is able to update the global OR figure accordingly. The peat soils from this 498 499 study are classified as Histosols_in the USDA soil taxonomy. and in Worrall et al. (2013) list 500 Histosols as having an OR of 1.03 (range 0.92 - 1.11) based on 8 studies, 23 locations and 501 345 samples - updating these values with data from this study gives an OR for Histosols of 502 1.08 with an interquartile range of 1.05 - 1.11, based now on 9 studies, 31 locations and 596 503 samples. The mineral soils for this study could be classed as Inceptisols. <u>-iIn</u> the review of 504 Worrall et al. (2013) the Inceptisols had an OR of 1.07 (range 1.03 to 1.10) based on 2 soils 505 from 1 study this can now be updated to be from 2 studies and 7 sites with 18 samples to give a median value of 1.15 (IQR 1.10 to 1.15). Coupling this information with new estimates for 506 507 Alfisols (Worrall et al., in review) that updated the OR of 1.10 (range = 1.07 to 1.12) based upon 1 study and 4 soil samples to that of 1.12 (range = 1.08 to 1.19) based upon 2 studies 508 and 2 soils but 84 samples. This would give a new value of $OR_{soil}^{global} = 1.084 \pm 0.002$, where 509 the uncertainty is interquartile range. Similarly, the values for vegetation types can be 510 511 updated, for grassland the new estimate would be 1.02 with an interquartile range of 1.00 -512 1.05; croplands would now have a median value of 1.00 (0.99 - 1.05); shrublands with a median value of 1.10 (1.08 to 1.12); and permanent wetlands with a median of 1.02 (0.99 – 1.04). This gives an $OR_{veg}^{global} = 1.049 \pm 0.2$ (error as IQR). A new residence time weighted global OR estimate for the terrestrial biosphere (OR_{terra}^{global}) of 1.056 (IQR = 1.054 to 1.058) can be calculated; the previous value of OR_{terra}^{global} was 1.04 ± 0.03.

This study has been able to validate the sampling approach of Worrall et al. (2013) <u>but</u> i.e. sampling the major carbon pools, but this work is able to now suggest that this can be improved further. <u>Bb</u>y increasing the level of sophistication and characterising individual components of the carbon pools e.g. <u>lignin vs. cellulose components</u>, active vs. passive soil organic matter, or dissolved organic carbon (DOC) fractions, it may be possible to elucidate the underlying controls on OR in the terrestrial environment. <u>Future research should</u> therefore explore these factors.

524

525 5. Conclusion

This study has shown that there are significant differences in oxidative ratio (OR) between mineral soils, peats and vegetation. Whilst there were significant differences in OR between different sites and material types, there was no significant interaction between the factors. Furthermore there was no significant difference in OR with peat depth. This suggests, on a large scale at least, that it the most important factor in OR variation is material type rather than site location and the OR of an environment can be determined by simple sampling of the major terrestrial carbon pools.

533

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612	Figure 1.	Location	of study sites	within	Great Britain.
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- 613
- Figure 2. Interval plot of elemental ratios, energy content (MJ/kg), bulk density (g/cm³) and
- 615 Oxidative Ratio with peat depth. 95% confidence interval of the mean
- 616
- Figure 3. Interval plot of OR for surficial peat samples for each site. 95% confidence interval
- 618 of the mean
- 619
- $\label{eq:G20} Figure \ 4. \ Plot \ of \ OR \ vs. \ \Delta Hc. \ a) \ Peat \ soil \ samples \ highlighted; \ b) \ Vegetation \ samples$
- 621 highlighted.
- 622
- 623

624	Table 1. Location information for each of the sites
625	
626	Table 2. Median values (inter-quartile range in parentheses) for each measured or derived
627	variable for the three material types.
628	
629	Table 3. Median values (inter-quartile range in parentheses) for each measured or derived
630	variable for functional plant groups
631	
632	Table 4. Median values (inter-quartile range in parentheses) for each measured or derived
633	variable for peat soils by site.
634	
635	Table 5. Median values (inter-quartile range in parentheses) for each measured or derived
636	variable for the three standards
637	
638	Table 6. One-way ANOVA within vegetation. p = probability of factor being zero, post hoc
639	testing where ≠ denotes a significant difference between levels.
640	
641	Table 6. ANOVA for peat samples energy content, C_{ox} and OR values, and bulk density. df =
642	degrees of freedom, p = probability of factor being zero, ω^2 = generalized proportion of
643	variance explained
644	Table 7 ANOVA for peat samples elemental ratios and OR values energy content and bulk
645	density $df = degrees of freedom, n = probability of factor being zero, \omega^2 = generalized$
646	proportion of variance explained
647	proportion of variance explained
047	

Table 8. <i>Post hoc</i> testing of site factor from Table 7 (ANOVA) where \neq denotes a			
significant difference between levels			
Table 97. ANOVA for the comparison of surficial peat and mineral soils with vegetation. df			
= degrees of freedom, p = probability of factor being zero, ω^2 = generalized proportion of			
variance explained			
Table 10. Post hoc testing of site factor from Table 9 (ANOVA) where \neq denotes a			
significant difference between levels.	Fi	eld Code Chan	ged
	Table 8. Post hoc testing of site factor from Table 7 (ANOVA) where ≠ denotes a significant difference between levels Table 97. ANOVA for the comparison of surficial peat and mineral soils with vegetation. df = degrees of freedom, p = probability of factor being zero, ω² = generalized proportion of variance explained Table 10. Post hoc testing of site factor from Table 9 (ANOVA) where ≠ denotes a significant difference between levels.	Table 8. Post hoc testing of site factor from Table 7 (ANOVA) where ≠ denotes a significant difference between levels Table 97. ANOVA for the comparison of surficial peat and mineral soils with vegetation. df = degrees of freedom, p = probability of factor being zero, ω² = generalized proportion of variance explained Table 10. Post hoc testing of site factor from Table 9 (ANOVA) where ≠ denotes a significant difference between levels.	Table 8. Post hoe testing of site factor from Table 7 (ANOVA) where ≠ denotes a significant difference between levels Table 97. ANOVA for the comparison of surficial peat and mineral soils with vegetation. df = degrees of freedom, p = probability of factor being zero, ω² = generalized proportion of variance explained Table 10. Post hoc testing of site factor from Table 9 (ANOVA) where ≠ denotes a significant difference between levels.