

# 1 **A method of estimating in-stream residence time of water in rivers**

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9

## 10 **Abstract**

11 This study develops a method for estimating the average in-stream residence time of  
12 water in a river channel and across large catchments, i.e. the time between water  
13 entering a river and reaching a downstream monitoring point. The methodology uses  
14 river flow gauging data to integrate Manning's equation along a length of channel for  
15 different percentile flows. The method was developed and tested for the River Tees  
16 in northern England and then applied across the United Kingdom (UK).

17 i) The study developed methods to predict channel width and main channel  
18 length from catchment area.

19 ii) For an 818 km<sup>2</sup> catchment with a channel length of 79 km, the in-stream  
20 residence time at the 50% exceedence flow was 13.8 hours.

21 iii) The method was applied to nine UK river basins and the results showed that  
22 in-stream residence time was related to the average slope of a basin and its  
23 average annual rainfall.

24 iv) For the UK as a whole, the discharge-weighted in-stream residence time was  
25 26.7 hours for the median flow. At median flow, 50% of the discharge-

26 weighted in-stream residence time was due to only 6 out of the 323  
27 catchments considered.

28 v) Since only a few large rivers dominate the in-stream residence time, these  
29 rivers will dominate key biogeochemical processes controlling export at the  
30 national scale.

31 vi) The implications of the results for biogeochemistry, especially the turnover of  
32 carbon in rivers, are discussed.

33

34 **Keywords:** transit time; reaction kinetics; DOC; BOD

35

## 36 **1. Introduction**

37 The time water spends travelling through a catchment is an important control of  
38 biogeochemical cycling and contaminant persistence. Water spends most time  
39 moving through subsurface storage before it enters the river channel (McGuire and  
40 McDonnell, 2006). Nevertheless, for a number of reasons it is important to  
41 understand how long water spends in a river channel, this can be called the in-  
42 stream residence time. This is not the same as the residence time or age of the  
43 water in the catchment since that encompasses the entire time between water  
44 entering the catchment as precipitation and leaving at the river mouth (McGuire and  
45 McDonnell, 2006; Heidbüchel et al., 2012). Here we are only concerned with the time  
46 between water entering the river channel and it passing a point of interest. In-stream  
47 residence time will be important if, for example, we wish to predict: how much of a  
48 pollutant will be lost in-stream; the in-stream turnover of a nutrient (eg. Honti et al.,  
49 2010); the emissions of greenhouse gases from riverwater to the atmosphere (eg.  
50 Battin et al., 2009); or, the in-stream algal abundance (Talling and Rzoska, 1967). It

51 is often possible to know the kinetics of in-stream processes (eg. Köhler et al., 2002)  
52 but knowing the rate of a process is only part of the solution as we need to know the  
53 amount of time over which the process will work, thus the in-stream residence time is  
54 critical. For example, soil and groundwaters are often highly concentrated in  
55 dissolved CO<sub>2</sub> with respect to the atmosphere (Worrall and Lancaster, 2005): when  
56 soil water containing excess dissolved CO<sub>2</sub> enters a river it will begin to degas CO<sub>2</sub>  
57 to the atmosphere (Billett and Moore, 2008). At the same time organic matter in the  
58 river water will be mineralised to produce dissolved CO<sub>2</sub> (Wickland et al., 2007).  
59 Rates of CO<sub>2</sub> degassing are known (Liss and Slater, 1974) and rates of DOC  
60 turnover in-stream are known (eg. del Giorgio and Pace, 2008), but it is only  
61 possible to estimate the amount of CO<sub>2</sub> entering the atmosphere if the in-stream  
62 residence time over which rates of processes are to be integrated is also known.

63 In-stream residence time ( $t_r$ ) can be defined as:

64

$$65 \quad t_r = \int_{x_e}^{x_m} \frac{x}{v} dx \quad (i)$$

66

67 where:  $v$  = the mean cross-sectional velocity at point  $x$ ;  $x$  = the downstream distance  
68 along the river channel;  $x_m$  = the downstream monitoring point; and  $x_e$  = the point  
69 along the river length where the water enters the river. For example,  $x_m$  could be the  
70 river mouth and  $x_e$  would be the point at which, on average, water enters the river.  
71 The distance  $x_m - x_e$  represents the length of the river travelled by water and  
72 henceforward we refer to this as the expected length of the river. Equation (i)  
73 therefore shows that, if we are able to estimate the change in mean river velocity  
74 along a river length, we can also estimate the in-stream residence time.

75 Mean cross-sectional velocity is commonly estimated as part of the  
76 consideration of hydraulic geometry. Leopold and Maddock (1953) proposed a series  
77 of power law equations that relate channel depth and mean velocity to stream  
78 discharge. This approach has the advantage that continuity constrains the constant  
79 and exponent terms. The power law approach has been popular and several studies  
80 have published the empirical fit of these equations for many rivers worldwide (e.g.  
81 Griffiths, 2003) and related the form of these equations to flow resistance (e.g.  
82 Ferguson, 2007). In some early studies, discharge was related to depth and to a  
83 residence time (Leopold et al., 1964). However, these equations do not tend to  
84 consider independent variables other than discharge, if this the focus were changed  
85 to consider in-stream residence time, then this would view downstream river length  
86 as the key independent variable (Equation (i)).

87 There have been a number of approaches to estimate the distribution of in-  
88 stream residence times using transient storage models (Bencala and Walters, 1983),  
89 but these approaches have a number of limitations. Firstly, they tend to rely on tracer  
90 studies and these have their own limitations - for example, irreversible adsorption of  
91 rhodamine dye (Lin et al., 2003). Secondly, the studies are based on solute transit  
92 times, i.e. they consider distribution of travel times from one point to another and, as  
93 observed by Gomez et al. (2012), these distances are typically short (of the order of  
94 1000m) rather <10 to >100 km which maybe the scale of interest for large-scale  
95 biogeochemical processes. Thirdly, not only have studies not considered scales of  
96 interest, they have not used these results to scale up to larger catchment areas or  
97 indeed to a wider range of flows. Wondzell (2011) has shown that transit storage  
98 becomes negligible when considering catchments greater than approximately 1 km<sup>2</sup>

99 and so either if they were or could be applied at larger catchments that would not be  
100 of much benefit.

101 Alternatively, some studies have considered transit times for water in whole  
102 catchments. Boning (1974) developed an empirical model of water transit times  
103 based on measured solute transit times from dye tracer tests. Soballe and Kimmel  
104 (1987) estimated annual average transit time ( $t_w$ ) for a series of east-coast US rivers  
105 based on the following empirical formula from Leopold et al. (1964):

106

$$107 \quad t_w = 0.08A^{0.6}Q_{ave}^{-0.1} \quad (ii)$$

108

109 where:  $A$  = catchment area ( $\text{km}^2$ ); and  $Q_{ave}$  = arithmetic mean annual discharge  
110 ( $\text{m}^3/\text{s}$ ).

111 A similar approach to calculate a transit time for flood peaks was proposed by  
112 Pilgrim (1987) and used by Robinson and Sivapalan (1997) and Sivapalan et al.  
113 (2002) where the mean channel response time ( $t_n$  - hours) is:

114

$$115 \quad t_n = \tau A^\omega \quad (iii)$$

116

117 where:  $A$  = catchment area ( $\text{km}^2$ ); and  $\tau, \omega$  = constants which for the case of  
118 Sivapalan et al. (2002) were 0.28 and 0.5 respectively.

119 Van Nieuwenhuysse (2005) proposed a method to calculate the transit time of surface  
120 water from its source as the water enters the river channel. Van Nieuwenhuysse  
121 (2005) showed there was a significant relationship with transit time based on dye  
122 tracer studies or average velocity at gauged sites based on discharge characteristics  
123 and catchment area. However, this empirical approach to the calculation of transit

124 time has some limitations. Firstly, the method had to consider average conditions  
125 where “average” was defined as arithmetic mean rather than the expected value of  
126 the true distribution of the river discharge. Thus, an estimate of average transit time  
127 could not be used to consider actual (expected) in-stream residence time or its  
128 distribution as is also the case for the methods illustrated in Equations (ii) and (iii)  
129 above. Understanding the distribution of transit times is important because it is often  
130 the extreme values that represent the greatest risk. At low values of transit time there  
131 is a risk of causing excess pollution: a risk of exceedence causing excess release of,  
132 for example, greenhouse gases; or conversely, underestimating pollutant retention  
133 as short-term storage is ignored (Drummond et al., 2012). Second, Van  
134 Nieuwenhuyse (2005) admits that the proposed approach estimated transit time and  
135 not in-stream residence time. While transit time is useful for predicting the flushing  
136 time of a pollutant along a given reach, it is not the in-stream age of the water  
137 passing any point, as transit time can only consider one point to one point, whereas  
138 water enters the river along a continuum at an infinite number of locations stretching  
139 back along the length of river to the channel. Indeed, Equation (i) could be used to  
140 estimate a transit time if  $x_e$  is a fixed point rather than the length of the river  
141 experienced by the water flowing past the point of interest. What is needed is a  
142 means of predicting the point at which the “average” water enters the river. The point  
143 at which the “average” water can be taken to enter the river could be understood in  
144 terms of the expected value of the downstream discharge profile of the river, i.e. it is  
145 the discharge weighted “average” river length. By using a discharge weighted  
146 approach, the “average” length is assessed on the basis of river length experienced  
147 by the volume of water passing down the channel.

148 Therefore, there is gap between the application of the transient storage  
149 models (eg. Gooseff et al., 2005) and the empirical models used to predict in-stream  
150 residence time (eg. Van Nieuwenhuyse (2005). The purpose of this study was to  
151 develop a method for estimating in-stream residence time of water in river channels  
152 where the method should work across a range of flows and across the full length of  
153 the river but rely on readily available information. The method developed needs to  
154 be applicable in different catchments and here it is applied across the United  
155 Kingdom (which includes the countries of England, Scotland, Wales and Northern  
156 Ireland – UK).

157

## 158 **2. Approach & Methodology**

159 The approach of this study is (i) to develop a method for calculating in-stream  
160 residence time; (ii) apply this method to a UK river where there is sufficient high-  
161 frequency flow data to test the method; and (iii) apply the method to other UK rivers.

162

### 163 **2.1. In-stream residence time**

164 The in-stream residence time can be defined as in Equation (i). The mean velocity of  
165 a river at any point can be estimated from the Manning equation (Manning, 1891):

166

$$167 \quad v = \left(\frac{1}{n}\right) \left(\frac{a_{cross}}{p}\right)^{\frac{2}{3}} s^{\frac{1}{2}} \quad (iv)$$

168

169 where:  $a_{cross}$  = cross-sectional area of the river at point x;  $p$  = the wetted perimeter;  $s$   
170 = the water surface slope; and  $n$  = the Manning coefficient. If Equation (iv) is

171 expressed in terms of  $x$ , i.e. the down-channel distance along the river, then  
172 Equation (i) can be used to estimate velocity as a function of down-channel distance.  
173 This assumes that the river is not impacted in any substantial way by impoundment.

174 It is common for the longitudinal slope profile of a river to be expressed as an  
175 exponential function of river length (Putzinger 1919):

176

$$177 S_x = S_0 e^{-\phi x} \quad (v)$$

178

179 where  $S_x$  = the bed slope at point  $x$ ;  $S_0$  = the bed slope at source;  $\phi$  = a constant. At  
180 the scale of the entire river length and at steady state, then it can be assumed that  
181 bed slope is a good approximation of the water surface slope in Equation (iv)  
182 (Wilson, 1994). Equation (v) can be readily calibrated for any catchment; here this  
183 was done by reference to altitudes of gauging stations on studied rivers.

184 If it is assumed that the river has a rectangular cross-sectional area then:

185

$$186 \frac{a_{cross}}{p} = \frac{dw}{(2d+w)} \quad (vi)$$

187

188 where  $d$  = river channel depth and  $w$  = river channel width. For a rectangular cross-  
189 section, the width of the river does not vary with discharge and so it is only  
190 necessary to find an expression for river width change with river length. The  
191 assumption of a rectangular section is the simplest possible formulation but could be  
192 readily replaced if more complex formulations of the river cross-section were  
193 required. A possible alternative formulation for equation (vi) is to consider a v-  
194 shaped, or triangular cross-section: :

195

196  $\frac{a_{cross}}{p} = \frac{dw}{\sqrt{w^2 + 4d^2}}$  (vii)

197

198 Other formulations of the channel-section, eg. trapezoidal, would mean that  
 199 additional paramters would be required to calculate cross-sectional area, eg. the  
 200 angle of the river bank. Since the angle of channel banks could not readily be known  
 201 for any individual catchment, this cannot be a general approach.

202 The further advantage of using the formulation in equation (vi) is that river  
 203 width does not vary with river depth. To calibrate equation (vi) with respect to width,  
 204 we used data collected by Dangerfield (1997) to create an empirical equation for  
 205 river width variation with catchment area. Dangerfield (1997) lists the bankfull width  
 206 of 124 UK rivers and these data were augmented with data from the River Tees  
 207 (Figure 1) to give the following equation (Figure 2):

208

209  $w = 0.061C + 9.0 \quad r^2 = 0.73, n= 129$  (viii)

210

211 where C = catchment area (km<sup>2</sup>); and w<sub>0</sub> = river channel width at source (m).

212 River channel depth, the other component of equation (vi), will vary with flow  
 213 and we propose the following form of equation:

214

215  ${}^f d_x = {}^f d_m - \beta e^{\left(\frac{x}{\gamma}\right)^\delta}$  (ix)

216

217 where:  ${}^f d_x$  = depth at exceedence flow f (eg. 10% exceedence) at river length x (m);

218  ${}^f d_m$  = depth of the river at the monitoring point m for exceedence flow f; and  $\beta, \gamma, \delta =$

219 constants where  $\beta$  approximates to  ${}^f d_m - {}^f d_0$  . Equation (ix) can be calibrated

220 against of observations of river depths at a given point for a given exceedence flow;  
221 furthermore, a Weibull function has a physical interpretation where a simple power  
222 law approach does not. For example, a Weibull function can represent a range of  
223 shapes of response, including sigmoidal, and the paramters in the equation can have  
224 physical meaning and be read directly from observations, eg. the minimum and  
225 maxium values observed are explicitly included in the equation.

226 One problem remains: relative to the monitoring point (at distance  $x_m$ ) at what  
227 point, on average, does the water enter the river system? In other words what is the  
228 average length travelled, what is the value of  $x_e$ ? We propose that average length  
229 travelled is the expected value of the function of discharge with river length: this is a  
230 discharge weighted length of the river. The form of the equation was taken as a  
231 Weibull function:

$$233 \quad {}^f Q_x = {}^f Q_m - \varepsilon e^{\left(\frac{x}{\theta}\right)^\mu} \quad (x)$$

234

235 Therefore the expected value is:

236

$$237 \quad l_e = {}^f Q_m \log_e(2)^{\frac{1}{\mu}} \quad (xi)$$

238

239 where:  ${}^f Q_x$  = discharge at river length  $x$  at exceedence discharge  $f$ ;  ${}^f Q_m$  = discharge  
240 of the river at the monitoring point  $m$  for the exceedence discharge  $f$ ; and  $\varepsilon, \theta, \mu =$   
241 constants. Again, equation (x) could be calibrated against records from river gauging  
242 stations.

243

## 244 **2.2. Testing**

245 The above approach was calibrated for the River Tees given data readily available  
246 for gauging stations in the UK as reported within the National River Flow Archive  
247 ([www.nrfa.ac.uk](http://www.nrfa.ac.uk)) and the Flood Studies Report (NERC, 1975 - Table 1). The data  
248 required were: mainstream river length to the gauge; altitude of the gauging station;  
249 flow duration curve (values for  $Q_{10}$ ,  $Q_{50}$ ,  $Q_{95}$  and  $Q_{bf}$  are routinely reported for river  
250 flow gauging stations in the UK); and the bankfull width and depth.

251 It is not possible to validate the above approach directly because there is no  
252 direct method of measuring in-stream residence time. However, it is possible to  
253 estimate the travel time of a storm hydrograph peak between two gauging stations if  
254 flow records of sufficient detail are available for stations at sufficient distance apart.  
255 Of course, the peak travel time is not the same as the in-stream residence time and  
256 so this cannot be strictly considered a validation, but it can at least be used to test  
257 whether the proposed method produces results of the correct order of magnitude. On  
258 the River Tees 15-minute flow records are available from 1982 for 3 gauging  
259 stations. Using the 2 stations that were furthest apart on the River (Broken Scar and  
260 Middleton-in-Teesdale – Figure 1, Table 1), the 15-minute flow record was examined  
261 for almost 5 years (1982-87) and each peak in flow at the upstream site was  
262 examined to see at what time it occurred at the lower stream site. The time of travel  
263 for each peak between the upper and lower gauging site was calculated and  
264 compared to the percentile flow at the upper and lower sites. This time of travel was  
265 then compared to the calculated in-stream residence time.

266

## 267 **2.3. Application to UK rivers**

268 The UK's National River Flow Archive (NRFA) was examined and all rivers where  
269 there were 5 or more gauging stations along the main stream length were  
270 considered; for each of these gauging stations the same data as for the River Tees  
271 were collected. For those rivers where it was possible to apply the above method,  
272 other catchment characteristics were recorded, including: catchment area to the  
273 lowest gauging station; maximum altitude within the catchment; and average annual  
274 rainfall (1961-1991) – these are all catchment characteristics reported as standard  
275 within the National River Flow Archive. The main stream river length to each gauging  
276 station from the start of the river was available from the Flood Studies Report  
277 (NERC, 1975); using its definition of a river start and by combining these data, the  
278 average slope of the river was calculated. The in-stream residence time ( $t_r$ ) was  
279 estimated at each of the flow exceedences ( $Q_{10}$ ,  $Q_{50}$ ,  $Q_{95}$ , and  $Q_{bf}$ ) for each of the  
280 selected rivers and compared to the selected catchment characteristics to develop a  
281 linear model of in-stream residence time that may be applied more broadly,  
282 particularly to rivers where the necessary catchment characteristics were available  
283 but where there were insufficient gauging stations for a separate calculation of the in-  
284 stream residence time. If an understanding of what controls in-stream residence time  
285 can be achieved, then it can be applied across regions. The catchments identified  
286 were amongst the largest in the UK and henceforward will be referred to as basins.

287 Linear equations developed for predicting in-stream residence time were  
288 applied across the UK. Since the aim of this study was to assess how long it takes  
289 water to travel through the river channel network across large catchments, it was the  
290 gauging stations furthest downstream that were examined. There are 323  
291 “downstream” gauging stations across the UK. Results from individual catchments  
292 were both discharge- and area-weighted in order to give an average value of in-

293 stream residence time for the UK. It should be noted that no river flow data were  
294 available for Northern Ireland and so strictly all data were for Great Britain and not  
295 the UK.

296

### 297 **3. Results**

#### 298 **3.1. Calibration for the River Tees**

299 The method was applied to the River Tees (Table 1). Equation (v) was fitted to the  
300 available slope data (Figure 3):

301

$$302 \quad S_x = 0.033e^{-0.022x} \quad r^2 = 0.93, n = 6. \quad (\text{xii})$$

303

304 Dangerfield (1997) did not include data from the River Tees and so data from the 5  
305 gauging stations on the Tees were used to augment Dangerfield's dataset. The  
306 smallest catchment area included by Dangerfield (1997) was 13 km<sup>2</sup>; this could only  
307 be marginally improved with data from the Tees to 11.4 km<sup>2</sup> (Table 1). Equation (vii)  
308 shows a significant linear relationship between catchment area and river width for  
309 catchments to 11.4 km<sup>2</sup> (5 km river length) but this equation suggest that rivers  
310 would be over 9 m wide at source. In order to correct for this overestimation in small  
311 catchments, it was assumed that Equation (v) applied for catchments larger than  
312 11.4 km<sup>2</sup> but for smaller catchments a second function (Equation xiii) was assumed  
313 to give a more suitable value of width at river sources:

314

$$315 \quad w = 0.68C + w_0 \quad (\text{xiii})$$

316

317 Equation (vii) can be calibrated against measurements for the Tees gauging stations  
318 (Figure 4):

319

$$320 \quad {}^{bf}d_x = 2.43 - 2.33e^{\left(\frac{x}{16.6}\right)^{1.47}} \quad \text{rmse} = 0.02 \quad (\text{xiv})$$

321

322 where rmse is the root mean square error. For the range of flows, Equation (ix) can  
323 be fitted against the available flow duration curves for the gauging sites along the  
324 Tees, for example for the 50% exceedence flows:

325

$$326 \quad Q_{50} = Q_{50m} - 8.1e^{\left(\frac{x}{40.1}\right)^{4.8}} \quad \text{rmse} = 0.11 \quad (\text{xv})$$

327

328 The good fit of the calibrated equations (equations xiv and xv) helps justify using the  
329 Weibull function. Given the fit of Equation (x) to the range of flows, the expected  
330 length and the depth correction are given in Table 2. As the expected length is a  
331 discharge-weighted length, it is not surprising that it will vary with the flow, in this  
332 case as measured by the % exceedence flow. The surprising result here is that the  
333 expected length of the river is relatively insensitive to changes in flow with only a  
334 decline in the expected length as bankfull discharge is approached, i.e. the average  
335 point at which water enters the river relative to the monitoring point moves closer to  
336 the source at maximum flows. For the River Tees the in-stream residence time  
337 varied from 46 hours for the 95% exceedence flows to 4 hours at bankfull. For each  
338 exceedence flow, Equation (ii) can be solved, in this case by numerical integration,  
339 to get the longitudinal velocity profile of the River Tees to the monitoring point at  
340 Broken Scar (Figure 5). It is notable that there is a maximum in the velocity for this  
341 river which is more pronounced with decreasing percentile exceedence flow.

342 For the period from the start of 15-minute flow records (February 1982) until  
343 December 1987, there were 531 events for which a transit time could be estimated.  
344 These 531 events covered percentile exceedence flows from 0 to 100% based upon  
345 all daily flows measured from 1961 to 2011. The measured peak transit times show a  
346 limiting curve from a peak transit time of 16.5 hours at 97.1% exceedence flow to a  
347 peak transit time of between 2.75 and 5.75 hours at 0.2% exceedence flow (Figure  
348 6). The calculated in-stream residence times for the same distance varied from 4  
349 hours at 1% exceedence flow to 36 hours at 95% exceedence flow. The estimated  
350 in-stream residence times match well to the measured peak transit times for flows  
351 greater than, approximately, the 50% exceedence flow but there is divergence  
352 between the measured transit times and the estimated in-stream residence time with  
353 in-stream residence time estimates curving upwards while transit time varies  
354 approximately linearly with flow. As noted previously, this comparison is not a true  
355 validation of the method as transit time represents the kinematic wave travel time  
356 while the in-stream residence time is the solute or particle travel time. Firstly, the  
357 data clearly show very short transit times occurred for flows that would have been  
358 different by orders of magnitude; this can easily be explained if the geometry of the  
359 catchment is considered. The assessment of transit time assumes that the flood  
360 wave enters from the river reach of interest through the upstream site but, depending  
361 upon the nature of the storm causing the increase in flow, this assumption may not  
362 be valid. The River Tees is predominantly a west-to-east flowing river and so any  
363 rainstorm which has a resolved component east to west will mean that a proportion  
364 of rain will enter the system below the upstream monitoring point causing a short  
365 circuit in the river reach between monitoring points, and would thus invalidate the  
366 assumption of the transit time calculation. Secondly, as noted by Van Nieuwenhuyse

367 (2005), a transit time is not an in-stream residence time. Transit time is a peak to  
368 peak comparison whereas in-stream residence time is the amount of time the  
369 average water spends in the river. If the method of Soballe and Kimmel (1987)  
370 (Equation (ii)) is applied to the Tees, a transit time of 3.5 hours would be predicted  
371 while observations from this study would suggest values between 4.25 and 9.25  
372 hours. Equally, Equation (iii) would suggest a value of 8 hours but it is not known for  
373 what percentile flow this is a prediction for. Although this was not a strict validation,  
374 the comparisons do provide some evidence that the method is capable of producing  
375 sensible results.

376

### 377 **3.2. Application to the UK**

378 There are 9 rivers in Great Britain where the main stream has 5 or more gauging  
379 stations upon it and, fortuitously, they cover much of the UK from north to south and  
380 thus span the range of land uses, hydroclimatic conditions and geomorphological  
381 settings found in the UK (Table 3 - Figure 7). The 9 selected catchments include the  
382 5 longest rivers in the UK and 8 of the 11 longest rivers with only the Tees being  
383 outside the top 20. The chosen catchments cover 43,000 km<sup>2</sup> out of a total UK area  
384 of 244000 km<sup>2</sup>. The catchments cover altitude ranges up to 1303 m above sea level  
385 while the extreme altitude range in the UK is 1343 m above sea level. The 9  
386 catchments include sub-catchments that are in top 25 wettest gauged catchments in  
387 the UK and the 25 driest gauged catchments in the UK out of 1453 gauged  
388 catchments. The method was applied to each of these basins and the results show a  
389 broad variation in estimated residence times (Table 3). The longest in-stream  
390 residence times was calculated for the largest basin considered (River Thames)

391 which is also the largest catchment in the UK with a predicted in-stream residence  
392 time of 151 hours (6.3 days) at median flow.

393 Using the readily-available catchment characteristics it was possible to  
394 produce significant relationships predicting in-stream residence times at different  
395 exceedence flows:

396

$$397 \ln(t_{r95}) = 6.8 - 1.5 \ln(\text{slope}) \quad r^2 = 90\%, n=9 \quad (\text{xv})$$

398 (0.3) (0.19)

$$399 \ln(t_{r50}) = 16.4 - 0.86 \ln(\text{slope}) - 1.7 \ln(\text{rain}) \quad r^2 = 96.1\%, n=9 \quad (\text{xvi})$$

400 (4.4) (0.22) (0.68)

$$401 \ln(t_{r10}) = 24.5 - 3.2 \ln(\text{slope}) \quad r^2 = 78\%, n=9 \quad (\text{xvii})$$

402 (4.3) (0.6)

$$403 \ln(t_{r5}) = 3.3 - 0.94 \ln(\text{slope}) \quad r^2 = 65\%, n=9 \quad (\text{xviii})$$

404 (0.4) (0.26)

405

406 where: slope = the average slope of the catchment to the downstream gauging  
407 station (m/km); rain = the annual average rainfall 1961 – 1990 (mm). Only those  
408 variables found to be significant at least at the 95% probability of being greater than  
409 zero were included and the numbers in the brackets are the standard errors in the  
410 regression coefficients and y-intercept. Equations (xv – xviii) all show a significant  
411 effect due to slope, in-stream residence time decreasing with increasing slope. It is  
412 possible to recalculate Equation (xvi) so as to include slope only and therefore  
413 Equations (xv – xviii) can all plotted together (Figure 8). It is not clear to the authors  
414 why a rainfall term should be significant only for the 50% exceedence flows but it  
415 may be that rainfall is collinear with slope at the national scale.

416 Equations (xv - xviii) were applied to 323 rivers across the UK to sites on  
 417 those rivers that represent the most downstream gauging station in their respective  
 418 catchments. The catchments cover an area of 149,000 km<sup>2</sup> out of possible 244,000  
 419 km<sup>2</sup> (65% of total area); catchment areas range from 1 to 9,948 km<sup>2</sup> with a geometric  
 420 mean of 147 km<sup>2</sup>. The unsampled catchments are most likely to be small and close  
 421 to the coast and, for most of the gauging stations being considered, the most  
 422 downstream gauging station is not precisely at the tidal limit. For 222 catchments no  
 423 mean stream length was reported; for the 111 catchments where a mainstream  
 424 length was known, the best fit equation with catchment area was found to be:

425

426 
$$l = \frac{\alpha C_{1/2max} C}{(\alpha C + C_{1/2max})} \quad r^2=0.90, n=111 \text{ (xix)}$$

427

428 where:  $\alpha$  = a constant (km/km<sup>2</sup>); and  $C_{1/2}$  = the area constant (km<sup>2</sup> – the catchment  
 429 area at which half the maximum rate of length increase is achieved) (km<sup>2</sup>). When  
 430 expressed in this manner, the constant represents the initial rate of change of river  
 431 length with catchment and for the best-fit equation  $\alpha = 0.142$  km/km<sup>2</sup>. The best-fit  
 432 value of  $C_{1/2}$  for the UK was 226 km<sup>2</sup>.

433 Equations (xv - xviii) were applied to all 323 catchments and their calculated  
 434 in-stream residence time was calculated at the 50% exceedence flow. The  
 435 discharge-weighted average in-stream residence time for the UK at 50%  
 436 exceedence flow was 26.7 hours (Table 4). The cumulative distribution of the flow  
 437 weighted in-stream residence time at 50% exceedence flow shows that 50% of  
 438 discharge-weighted average in-stream residence time for the country was accounted  
 439 for by only 6 out of the 323 catchments considered (Thames, Ely Ouse, Severn,  
 440 Trent, Tweed, Wye – Figure 7 and 9a). The distribution of in-stream residence time

441 at 50% exceedence flow shows that the UK almost divides exactly east-west with all  
442 the long-residence time rivers in the east (Figure 9b); this distribution represents the  
443 topography of the UK with eastward-flowing rivers being longer and coming from  
444 lower altitudes regions compared to shorter, steeper west-flowing rivers. It should be  
445 noted that none of these rivers are in Scotland where high slopes and high rainfall  
446 may give rise to high discharges but also short in-stream residence times. The  
447 Thames accounts for 14% of the discharge weighted in-stream residence time for  
448 the entire country at median flows. The longest in-stream residence time calculated  
449 was for the River Glen which is a 37 km stream but has a mean slope of only 0.34  
450 m/km; however, when discharge weighted, the in-stream residence time of the River  
451 Glen represents only 0.7% of the national in-stream residence time. At 10%  
452 exceedence flow the in-stream residence time decreases to 2 hours; and is 67 hours  
453 at 95% exceedence flow. At the lowest flows 32% of the discharge weighted in-  
454 stream residence time is contributed by only two rivers (Thames and Ely Ouse –  
455 Figure 7 and 9).

456 When area-weighted, the UK in-stream residence time at 50% exceedence  
457 flow is 56 hours (Table 4) with 50% of the area-weighted in-stream residence time  
458 accounted for by only 5 catchments. At 10% exceedence flow the area-weighted in-  
459 stream residence time is 2.5 hours with only 12 catchments accounting for 50% of  
460 the area-weighted in-stream residence time of the entire country. At 95%  
461 exceedence flow the area-weighted in-stream residence time is 156 hours with 50%  
462 of this value contributed by only 3 rivers

463

#### 464 **4. Discussion**

465 The method presented in this study includes changing flows across large-scale  
466 (10,000 km<sup>2</sup>) basins but does so using information often readily-available in  
467 developed countries, i.e. multiple rated sections along the course of the river, and  
468 thus the approach can be considered as a clear advance on the empirical methods  
469 as represented by Equation (ii). The question is: how good is the approach relative to  
470 the more physically-based approaches used in transient storage models? Firstly, this  
471 approach does work across large catchments and basins even scalable to the size of  
472 the UK which has not been done for transient storage approaches. Secondly, the  
473 approach did not require tracer studies but could use river flow and topographic data.  
474 The expected effect of transient storage within a stream would be to increase the in-  
475 stream residence time with the increased time being spent in dead-zones, pools and  
476 the hyporheic zone. The importance of time spent in the hyporheic zone is the great  
477 potential for biogeochemical processing (e.g. Pinay et al., 2009). However, studies  
478 have struggled to show a relationship between exchange with transient storage and,  
479 for example, in-stream nutrient cycling (e.g. Hall et al., 2002). The role of transient  
480 storage is then either highly variable across time and space, or not as important as  
481 first thought. Wondzell (2011) compared exchange of water with the hyporheic zone  
482 ( $Q_{hz}$ ) with down-channel discharge ( $Q$ ) and found that the ratio of  $Q_{hz}/Q$  was  
483 maximum for the lowest order stream but even then it was 1.9%: at 60 km<sup>2</sup> the  $Q_{hz}/Q$   
484 was as low as 0.002%, i.e. negligible. The study showed that  $Q_{hz}$  was essentially  
485 constant with changing  $Q$  and so its importance decreased with increasing  $Q$ .  
486 Furthermore, potential hyporheic exchange would be lowest where the stream bed  
487 was composed of fine-grained sediments as opposed to gravels with the exchange  
488 being limited by the effective hydraulic conductivity of the stream-bed. Given the  
489 catchment scale used in this study, the result of Wondzell (2011) suggests that

490 transient storage has a near negligible effect on a method that was discharged-  
491 weighted. The result of Wondzell (2011) mirrors that of Robinson et al. (1995) who  
492 showed that transport properties in catchments greater than 10 km<sup>2</sup> were network-  
493 dominated as distinct from being hillslope-dominated. This is not to say that transient  
494 storage areas are not important for biogeochemical processing, because their ability  
495 to cycle nutrients or remove pollutants might be disproportionate to the volumes of  
496 water exchange, but the inclusion of biogeochemical rates would be a separate  
497 study. Equally, no method for estimating transit time in rivers, be it the method  
498 proposed here or other methods discussed, can allow for the presence of lakes and  
499 reservoirs. It is known that lakes and reservoirs act as large stores of  
500 biogeochemically important components and can have water residence times of  
501 years (e.g. Syvitski et al., 2005). Fortunately, the UK is relatively unimpounded and  
502 has few large lakes. The method proposed here is limited by its need for calibration  
503 data; in this study a minimum of 5 gauging stations per river was set as a minimum  
504 number so that the fit of equations such as equation (viii) is based only on a very  
505 small number of data points. However, the results from calibrated catchments could  
506 be used to generalise across flows and catchments and other approaches also  
507 require calibration often with more parameters to fit than required here.

508         Our motivation for modelling in-stream residence is to understand the time  
509 over which biogeochemical reactions can occur. For example, the measurement of  
510 BOD in the UK is based upon a 5-day measurement yet the in-stream residence time  
511 even at 95% exceedence flow is less than 3 days. When a 5-day in-stream  
512 residence time is considered, then even at 95% exceedence flow there are only 26  
513 out of 323 catchments that showed a in-stream residence time greater than 5 days:  
514 these catchments represent 18% of the land area, but represent only 2% of the

515 discharge. Therefore, for UK conditions a 5-day BOD measurement represents an  
516 extreme worse case and, in most cases, would represent impacts on estuaries and  
517 not on the river.

518 An improved method to estimate the in-stream residence time would be to use  
519 a tracer which starts changing the moment it enters the stream. One possibility is the  
520 excess dissolved CO<sub>2</sub> concentration: this is the concentration of CO<sub>2</sub> that is present  
521 in excess over and above that would be in equilibrium with the atmosphere. Soil- and  
522 ground-waters have dissolved CO<sub>2</sub> concentrations well in excess of that which would  
523 be present in equilibrium with the atmosphere. Worrall and Lancaster (2005)  
524 considered the excess dissolved CO<sub>2</sub> concentrations throughout the River Thames  
525 catchment over a 29-year period and showed the mean concentration of excess  
526 dissolved CO<sub>2</sub> in groundwater was 4.99 mg C/l, for clay soil catchment at source the  
527 mean was 4.46 mg C/l, while for surface water at the catchment outlet the average  
528 concentration was 0.79 mg C/l, i.e. groundwater and soil water had degassed on  
529 emergence at the surface. Jones and Mulholland (1998) suggested that excess  
530 dissolved CO<sub>2</sub> concentration at a catchment outlet was:

531

$$532 \quad p\text{CO}_{2\text{stream}} = p\text{CO}_{2\text{gw}} - p\text{CO}_{2\text{evasion}} + p\text{CO}_{2\text{metabol}} \quad (\text{xx})$$

533

534 where:  $p\text{CO}_{2\text{stream}}$  = dissolved CO<sub>2</sub> in stream at the catchment monitoring point;  
535  $p\text{CO}_{2\text{gw}}$  = the dissolved CO<sub>2</sub> from the soil-groundwater of the catchment;  $p\text{CO}_{2\text{evasion}}$   
536 = the dissolved CO<sub>2</sub> lost to the atmosphere between groundwater emergence and  
537 the catchment monitoring point; and,  $p\text{CO}_{2\text{metabol}}$  = the dissolved CO<sub>2</sub> produced by in-  
538 stream metabolism between the discharge of groundwater into the channel and the  
539 catchment monitoring point. It should be possible to reverse this equation, if the

540 concentration at source and outlet are known and the rates of evasion and metabolic  
541 production are known, then the in-stream residence time can be calculated. Neal et  
542 al. (1998) give a range of methods for calculating excess dissolved CO<sub>2</sub> from a  
543 range of often readily available monitoring data (combinations of pH, alkalinity, Ca  
544 and stream temperature). The evasion rate of CO<sub>2</sub> from the stream water can be  
545 estimated from the stagnant two-film model (Liss and Slater, 1974). The problem is  
546 the estimation of the metabolic production of CO<sub>2</sub> in stream from the turnover of  
547 organic matter. River flow gauging stations and catchment characteristics are widely  
548 available in many developed countries, but measures of organic matter turnover are  
549 rare and perhaps the only widespread measure of organic turnover is BOD and such  
550 a measure has already been criticised above.

551 Zarnetske et al. (2012) proposed that a bulk Damköhler number could be  
552 used for stream channels once a residence time is known. A bulk Damköhler number  
553 can be defined as:

554

$$555 \quad D_{river} = \frac{kl}{v} C^{n-1} = kt_r \quad (\text{xxi})$$

556

557 where: k = the first order removal rate ([M][L]<sup>-3</sup>[T]<sup>-1</sup>); l = the river length ([L]); v = water  
558 velocity ([L][T]<sup>-1</sup>); C = initial concentration ([M][L]<sup>-3</sup>); and n = reaction order. Worrall et  
559 al. (2013) have measured zero-order rate constants for DOC loss in the River Tees  
560 as between (0.19 and 2.15 mg C/l/hr). Moody et al. (2013) gave the average initial  
561 concentration of the DOC in the headwaters of the River Tees between 1993 and  
562 2008 as 17.6 ± 6 mg C/l, where n= 896 and the variation is difference between the  
563 25<sup>th</sup> and 75<sup>th</sup> percentiles. Applying the above method for in-stream residence time to  
564 the DOC sources of the River Tees over the period for which initial concentrations

565 were known gives values of  $50.3 \pm 22$  hours. Applying these ranges to Equation (xxi)  
566 gives a median Damköhler number of 2.9 with an inter-quartile range of 1.8 to 4.2,  
567 i.e. this would approach would suggest that for DOC in the River Tees the dominant  
568 process is removal of DOC over advection.

569 For wider application, the in-stream residence time to a point of interest could  
570 help target management intervention to relieve problems of water quality For any  
571 water quality component (e.g. dissolved organic carbon, DOC; nitrate) that is turned  
572 over and removed in stream water, then knowing the in-stream residence time can  
573 then target land management options in a catchment. If the rate of turnover is known  
574 and this compared to the in-stream residence time, then it would be possible to  
575 identify the region within which the river has not had time to reduce the  
576 concentration. For example, Moody et al. (2013) has shown that on average over a  
577 12-month period DOC concentrations decreased by an average of 70% in UK river  
578 water over a 24 hour period and within this time reached a new equilibrium  
579 concentration. Therefore, for areas of a catchment outside 24 hours travel time of a  
580 water treatment works, there is little point investing in land management as the river  
581 has sufficient time to process and limit the concentration; however, within a 24 hour  
582 travel time then the river will not have sufficient time to process the inputs and  
583 source control would be more effective.

584

## 585 **5. Conclusions**

586 The study has developed a method for calculating in-stream residence time  
587 applicable to catchments where there are 5 or more gauging stations. The method  
588 was applied to 323 catchments across the UK by comparison to catchment  
589 characteristics in order to give regional estimates of in-stream residence time. When

590 estimates of in-stream residence time were compared between catchments, it is  
591 shown that, for UK rivers as a whole, the in-stream residence is dominated by a  
592 small number of large, low-gradient rivers.

593

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597

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Figure 1. Location of gauging stations within the River Tees, northern England.

Figure 2. The bankfull width compared to catchment area for 124 catchments from Dangerfield (1997) and from 5 gauging stations on the River Tees.

Figure 3. The change in slope along the length of the River Tees from its source at the channel head with Putzinger equation fitted (Equation (v)).

Figure 4. The fit of equation (xiii) to the observed river depth at bankfull discharge for 5 gauging stations on the River Tees.

Figure 5. The downstream velocity profile (from channel head of the main channel) of the River Tees for varying exceedence flows as predicted by this study.

Figure 6. Observed transit times with varying exceedence flow for the River Tees between Middleton-in-Teesdale and Broken Scar in comparison to predicted in-stream residence times.

Figure 7. The location of the rivers and gauging stations used in the calculation of in-stream residence time for the UK. Where: 1 = Tees; 2 = Thames; 3 = Severn; 4 = Trent; 5 = Bedford Ouse; 6 = Tweed; 7 = Clyde; 8 = Spey; 9 = Wye; and 10 = Ely Ouse.

Figure 8. The variation of mainstream channel length with catchment area.

Figure 9. a) The percentage of the national in-stream residence time at 50% exceedence flow represented by each river in the study. b) The in-stream residence time at 50% exceedence flow for each river catchment studied.