Reliance on exotic plants by two groups of threatened samango monkeys, *Cercopithecus albogularis labiatus*, at their southern range limit

# Kirsten Wimberger<sup>ab\*</sup>, Katarzyna Nowak<sup>c,d</sup>, Russell A. Hill<sup>c</sup>

<sup>a</sup>Biological Sciences, University of Cape Town, Cape Town, South Africa <sup>b</sup>Zoology and Entomology, University of Fort Hare, Alice, South Africa <sup>c</sup>Evolutionary Anthropology Research Group, Durham University, Durham, UK <sup>d</sup>Zoology and Entomology, University of the Free State, Owaqwa, South Africa

\*Corresponding author: wimbergerk@gmail.com

# Abstract

Understanding how threatened species adapt their behavior to landscapes shaped by humans is increasingly important to ensuring they persist in a changing world. Matrix habitats can be shared spaces where human and non-human primates co-exist. We set out to determine how an endemic, nationally threatened forest specialist, the frugivorous, arboreal samango monkey (*Cercopithecus albogularis labiatus*), has responded to a matrix habitat made up of residential gardens and commercial plantations in Eastern Cape province, South Africa. We followed two groups from dawn to dusk for a mean of three days per month for 12 months (1 February 2011 to 31 January 2012) using scan sampling to collect data on their diet, activity and ranging patterns. We used resource abundance transects to describe the groups' home ranges and monitored tree phenology to calculate fruit and seed availability indices.

accounting for over 50% of their overall annual diet, with seeds of the invasive black wattle (*Acacia mearnsii*) the most commonly consumed exotic species followed by acorns of two oak species (*Quercus robur* and *palustris*.). However, monkeys responded to the availability of indigenous rather than exotic fruits and seeds and increased their consumption of exotics when indigenous fruits were less available. While monkeys spent less time moving when feeding on exotic species compared to indigenous species, eating exotics did not free up monkeys' time to rest or socialize as additional time was required to process exotic foods. To offset the possible negative consequences of the monkeys' reliance on exotic seeds, including escalating conflict between monkeys and people in gardens, we suggest gradual removal of exotic plant species in the habitat and replacement with indigenous species as one mitigation strategy.

**Keywords**: ecological flexibility, *Acacia mearnsii*, black wattle, human-modified habitat, fallback foods, *Cercopithecus mitis* 

## Introduction

Wild animals are increasingly compressed within small patches of native vegetation surrounded by matrices of human-modified land (Chapman et al. 2007; Estrada 2009; Estrada et al. 2012; Fischer and Lindenmayer 2007). Some of these matrices consist of exotic plants (non-native species found outside their normal distributional range: van Wilgen 2011), which may provide an additional resource for native wildlife (Vitule et al. 2012), including threatened species (Schlaepfer et al. 2011). Exotic plants, because they lack natural predators and disease risks (Iponga 2011; van Wilgen and Impson 2011), can alter the quality and quantity of nutrients available to local wildlife by, for example, yielding an increased availability and abundance of certain plant parts such as seeds (Richardson and van Wilgen 2004; van Wilgen 2011; Vitousek 1990) or fruit (Davis 2011).

Opportunistic taxa like some primates can thrive on exotic vegetation given their dietary flexibility and ability to exploit newly available resources (McKinney et al. 2015). The incorporation of exotic species into the diet can influence primates' patterns of activity. For example, vervet monkeys (*Chlorocebus pygerythrus*) (Saj et al 1999) and baboons (*Papio* spp.) (Strum 2010, van Doorn et al 2010) that consume cultivated plant parts (crops) and human food waste spent less time foraging and traveling and more time resting and socializing because metabolic demands were easily and quickly satisfied. Ring-tailed lemurs (*Lemur catta*) that ate white cedar (*Melia azedarach*) and cattley guava (*Psidium cattleianum*) devoted less time to moving and feeding than a group without access to such exotics (Gabriel 2013). In contrast, chacma baboons (*Papio ursinus*) living in exotic pine (*Pinus* spp.) plantations had similar annual activity budgets to those living in natural habitats although their relatively smaller home ranges suggested that pine nuts are spatially clumped but require similar processing times to native foods (Hoffman and O'Riain 2011).

According to optimal foraging theory, primates should choose to eat a specific food item that maximizes their energy returns, by weighing up the benefits (e.g. high energy food source) and the costs/constraints (e.g. high secondary compounds, increased visibility to predators) (Krebs and Davies 1993, Began et al. 1996). Fruits, which are easily processed and are a high-energy resource (Hemingway and Bynum 2005; Kaplin and Moermond 2000; Wilson and Downs 2012), are a classic "preferred food" of frugivorous primates, constituting the majority of the diet, and are eaten when available (Marshall and Wrangham 2007). Leaves are a common "fallback food" for frugivorous primates (e.g. blue monkeys, *Cercopithecus mitis*, Foerster et al. 2012), and are eaten when preferred species and parts are seasonally unavailable (Marshall and Wrangham 2007). Fallback foods are usually abundant and easily located but hard to process due to high fibre and plant secondary compounds (Lambert 2002, Hemingway and Bynum 2005, Rosenberger 2013). An important question in matrix habitats is therefore whether primate species targeting exotic plant species consume them as preferred or fallback foods.

Exotic fruits and crops are often targeted by primates (Strum 2010). These plant parts are generally more abundant and have higher sugar content than indigenous fruit (Mokotjomela et al. 2013), and/or they are more available during seasonal shortages of endemic fruits. For instance, a population of howler monkeys (*Aloutta caraya*) targeted exotic orange trees (*Citrus sinensis*), where >60% (and as much as 97%) of fruits eaten were oranges (Bicca-Marques and Calegaro-Marques 1994). A group of chimpanzees (*Pan troglodytes verus*) increased their reliance on cultivated plants when experiencing a dearth in the availability of wild fruits

(Hockings et al. 2009). Primate populations may even come to depend on or develop preferences for exotic plant species, which help them achieve higher population densities and improved survival probabilities in fragmented habitats (Singh et al. 2001; Pozo-Montuy et al. 2013). For example, Zanzibar red colobus monkeys (*Procolobus kirkii*) that rely heavily on exotic Indian almond (*Terminalia catappa*) outside of Jozani National Park exist at the highest population densities recorded for any nonhuman primate (Siex and Struhsaker 1999).

The samango monkey (*Cercopithecus albogularis*, subspecies *labiatus*, *erythrarchus* and *schwarzi*: Dalton et al. 2015), a forest specialist, is nationally threatened in South Africa (Linden et al. 2016) where it is becoming increasingly isolated and detrimentally affected by forest fragmentation and deforestation from historic logging and ongoing forestry plantations (Friedman and Daly 2004; Kingdon et al. 2008; Lawes 1992). Groups are reluctant to disperse between forest patches or to inhabit small and isolated forest fragments (Friedman and Daly 2004; Lawes 1992, 2002). Matrix habitat, planted with exotic species, might help to increase connectivity and encourage movement between fragments, as it does for the Angola black-andwhite colobus (*Colobus angolensis palliatus*, Anderson et al. 2007). Some samango monkey populations have been observed to occupy matrix habitat and eat plant parts from exotic species including loquat (*Eriobotrya japonica*), pine (*Pinus* spp.), oak (e.g. *Quercus robur*), black wattle (*Acacia mearnsii*) and bugweed (*Solanum mauritanum*) (Breytenbach 1988; Lawes et al. 1990; Linden et al 2015; McMahon 1977; von dem Bussche and van der Zee 1985).

Samango monkeys are arboreal and frugivorous, with 50% of their diet consisting of fruit (range: 52 - 92 %); followed by leaves (range: 2 - 44%), with flowers, insects, seeds and fungi eaten in smaller quantities (reviewed in Coleman and

Hill 2014). During seasonal fruit scarcity, these monkeys consume more mature leaves, flowers and unripe fruits (Lawes et al. 1990; Lawes 1991). Samango monkeys are able to subsist on a folivorous diet when fruit is scarce largely as a consequence of being hindgut fermentors, such that they are able to extract protein from the usually non-digestible cellulose in plant cell walls (Bruorton and Perrin 1988).

We studied two groups of samango monkeys in and around the village of Hogsback, Amathole Mountains, Eastern Cape, South Africa, where both groups had access to a matrix of exotic plants as well as indigenous forest. One group's home range included the village, and thus had more people, dogs, shops, houses, roads and electrical lines compared to the other group's range. Both groups' ranges included people's small holdings which have been planted with exotic "ornamental" plant species (e.g. Azalea), and supplementary food for the growers such as vegetables (e.g. maize, Zea mays), nut and fruit trees (e.g. cherry plum, Prunus cerasifera and pecan trees, Carya illinoinensis, see here: http://hogsbackgardens.blogspot.co.za/). These food items provide an incentive for monkeys to enter residential properties and have possibly resulted in the monkeys' home ranges increasingly overlapping with people's gardens. Starting about 10 years ago, human-monkey conflict escalated in Hogsback with people complaining that monkeys did not previously occur outside of the adjoining indigenous forest, that they are wasteful with food obtained from planted fruit and nut trees, are becoming bold and aggressive, are targeting rubbish bins and houses for food, and are also increasing in numbers (Nowak et al., this issue, Wimberger and Bidner 2012).

We predicted that samango monkeys would target the abundant exotic species, particularly black wattle and oak trees. These species occur in large stands in Hogsback and the seeds are consumed readily by non-human and human primates where they are native as they are a good source of oils and protein (Seigler 2003, Guo et al. 2007). We set out to understand the extent of monkeys' reliance on exotic plants and the seasonal and habitat drivers of this consumption. Hogsback is a highly seasonal site at the most southerly extent of this taxon's range, with seasonal variation in day length representing an important ecological constraint as short winter days restrict the length of time available for foraging and other activities (Hill et al. 2003). We use multiple linear regression models to determine the factors that best account for the consumption of indigenous and exotic species by samango monkeys at Hogsback, exploring the role of seasonal variations in day length, rainfall, group ID, and indices of indigenous fruit and seeds and exotic fruit and seed availability. Since consumption of exotic fruit and seeds would provide monkeys with high protein and fat content, we examine if time feeding on indigenous fruits and exotic species has associated with activity patterns and day journey lengths. Finally, we expected that the group with greater access to the village and gardens would exploit exotic species relatively more and so incorporate group size into all of our analyses.

# Methods

#### Study site and study groups

The study site lies within the Amathole mountain range in the Eastern Cape province of South Africa, at an altitude of roughly 1270 m and includes the village of Hogsback (32°35' S, 26°56' E). Our behavioral study ran from 1 February 2011 to 31 January 2012, where we define seasons as autumn (March-May), winter (June-August), spring (September-November) and summer (December-February). Although typically a summer rainfall area (mean annual rainfall = 1029 mm), during the study period there was high winter rainfall (measured daily on site by a resident using a rain gauge; Webster unpubl data). Temperatures ranged from a mean minimum of 5.7  $(\pm 1.2)$  °C in winter (June-Aug), when it often snows, to a maximum of 29.6  $(\pm 2.2)$  °C in summer (Dec-Feb) (South African Weather Service: SAWS unpubl data). Day length ranges from 10.0 hrs in winter to 14.3 hrs in summer which we derived based on the time difference between the sunrise and sunset times provided by a hand-held Global Positioning System (GPS, Garmin Dakota 20, Garmin Ltd., Kansus, USA) for our specific location.

The natural vegetation (indigenous forests) at the site is characterized as southern mistbelt forest, typically found between 1000 and 1400 m ("Afromontane") and dominated by large trees including yellowwood (*Podocarpus/Afrocarpus* spp., white stinkwood (*Celtis africana*), cape chestnut (*Calodendrum capense*) and forest knobwood (*Zanthozylum davyi*, Mucina and Geldenhuys 2006). However, these indigenous forests are degraded as a result of the historical harvesting of large yellowwood trees, the introduction of exotic pine (*Pinus* spp.) within commercial plantations and the invasion of exotic species, particularly Australian black wattle (Webster 2009).

We selected two samango monkey groups that use indigenous forest and a matrix of residential gardens. The range of G1 (approximately 27 individuals) overlapped with Hogsback village, while we chose G2 (approximately 18 individuals) on the basis of its range being 1.5 km outside the center of the village and at least 1 km from the other study group. G2's home range contained a very small portion (<1%) of commercially planted exotic pine. Both groups' home ranges were outside any formally protected area.

We followed each group from dawn to dusk (summer  $\sim 05:00$  to 19:00, winter  $\sim 07:00$  to 17:00) for a mean of three days per month for 12 months to collect data on

diet, activity and ranging patterns. We followed G1 for 35 days (6 summer, 9 autumn, 11 winter, 9 spring) and G2 for 34 days (6 summer; 10 autumn; 10 winter; 8 spring). Our locating of groups was facilitated with the use of VHF (Telenax, Mexico) radio collars, fitted by a veterinarian in September 2010 on the group male, and 2 (G2) or 3 (G1) adult females in each group. Prior to data collection we habituated both groups to close observation (~5 m) over a period of six months.

#### Diet and activity budgets

We conducted instantaneous scans sampling every 15 minutes (4 samples per hour) with each scan lasting 5 minutes (Guo et al. 2007; Kaplin and Moermond 2000). While there is some debate on the accuracy of scan sampling methods for examining foraging and food intake rates (Reynoso-Cruz et al 2016, Amato et al 2013, Zinner 1999, Harcourt and Stewart 1984), it was the most appropriate method for this study given relatively large group sizes and difficulty identifying individuals. We collected data on adult females (physically large, long nipples), sub-adults (physically smaller, females have short nipples, males have visible testes) and juveniles (physically smaller, females have no visible nipples, males have hardly visible testes; but not on infants (physically small, not independent, suckle) or adult males (one male per group, large testes, physically larger than adult females). We estimated that both groups had a similar ratio of individuals within each age class (roughly 5% adult male, 30% adult females, 65% sub-adults and juveniles), but cannot rule out some small variation that may influence our results. In total, we collected 387 observation hours for G1 (13,132 individual scan samples; mean =  $25.0 (\pm 17.3 \text{ SD})$  individuals per hour), and 374 observation hours for G2 (12,027 individual scan samples; mean = 23.5 ( $\pm 15.2$  SD) individuals per hour).

We classified activity into five broad categories: 1) feeding (including 'eating' i.e. ingesting food and 'foraging' i.e. searching for and processing food), 2) moving (e.g. walking, running), 3) socializing (affiliative interactions such as grooming), 4) resting (e.g. sitting, lying), and 5) other (Kaplin and Moermond 2000; van Doorn et al. 2010). We categorized all plant parts consumed as either indigenous or exotic, and grouped them into types: fruit (including unripe, ripe), seed pod/nut (hereafter termed "seed", including unripe, ripe), leaves (including young, mature), stems, buds (including buds, shoots), exudate, invertebrates, human food waste and "other" (including fungus, lichen, and clay). We combined unripe and ripe fruit and seed and also young and mature leaves for some species (e.g. fruit of Celtis africana, vine leaves) when it was difficult for us to visually distinguish between different stages that were eaten by monkeys in the tall canopy (as was done by Tesfaye et al. 2013 who studied Boutourlini's blue monkeys C. m. boutourlini). We combined plant parts eaten from the two most common oak species (Quercus robur and Q. palustris) in Hogsback and in the groups' diet, as species distinction was not important in determining the role of this exotic food in the diet of these monkeys.

We derived a measure of consumption of each food item by summing the number of individuals observed to be eating a particular food item during each of the four 5 minute scans within each hour of observation, then dividing by the total number of individuals observed in that hour. Similarly, we quantified activity budgets by summing the number of individuals observed to be engaged in a particular behavior during each of the four scans within each hour of observation, then dividing by the total number of individuals observed in that hour. To control for seasonal differences in day length, we calculated mean dietary and activity proportions for each day and multiplied by the day length to determine hours per day devoted to feeding on a specific item or devoted to an activity (following Hill et al. 2003, 2004; Hoffman and O'Riain 2011). We then excluded the "other" activity category from further analysis. To allow for dietary comparisons with other studies, we also present mean annual percentages of each dietary item consumed out of total eaten.

## Home range characteristics

We recorded the location of the approximate group center every 30 minutes during group follows with a hand-held GPS (Dakota 20, Garmin Inc., USA). We imported GPS data points (G1, n=954, G2 n=947) into ESRI<sup>®</sup> ArcMap<sup>TM</sup> 9.3.1 (ESRI, Inc. 2009) and projected them in UTM Zone 35S, spheroid WGS 1984, central meridian 27. We calculated annual home ranges using 90% fixed kernel density estimator, a common method for calculating home range (e.g. Gabriel 2013) from Hawth's Analysis Tools 3.27 (Beyer 2004) extension with a smoothing factor (H) of 50 and raster cell size of 10. We used Hawth's Analysis Tools to overlay a grid on the GPS data points, with a grid cell size of 50 x 50 m chosen based on an estimate of group spread (similar to that reported for blue monkey, Kaplin 2001).

We labelled each grid cell as either "indigenous" or "matrix" by visually assessing whether native or exotic plants were dominant (>50%) in each cell of the grid, placed over a recent (2010) aerial photograph of the area (Department of Rural Development and Land Reform, South Africa: Cape Town). We visually estimated boundaries between habitat types by on the ground observations. Within each of these two types of habitats, and for each group, we placed three fixed strip-width transects (100 x 10 m, 0.1 ha; modified from Boyes and Perrin 2010). The size of the sampled area should capture the diversity of the habitat within the home range and cover the key habitat, with "adequate" sample size considered to be between <1% (Chaves et al

2012) and 10% (Snaith and Chapman 2008). We sampled 0.6 ha of each group's home range, which constituted approximately 3% of each group's home range (~12% of the groups' core home ranges), and due to homogeneity of indigenous habitat, we deemed this adequately sampled. Not all species consumed by the monkeys were detected along our matrix habitat transects due to the clumping of exotic species, but we did not want to bias our sampling by specifically targeting certain patches. We recorded the species, height, and diameter at breast height (DBH) of all trees with >5 cm DBH along each transect and with a width of 5 m either side (following Renton 2001, Boyes and Perrin 2010). DBH has been established as a good proxy of fruit abundance (Boyes and Perrin 2010; Chapman et al. 1992; Renton 2001). We calculated tree densities (# / ha) and tree species richness for each habitat type within each group's home range by dividing the total number of trees or species by 0.3 ha. We also calculated the percentage of large trees (> 40cm, Lee et al. 2014) for each habitat.

As the samango monkey is largely frugivorous (Coleman and Hill 2014), we focused on fleshy fruit availability, but also monitored the availability of black wattle seed pods and oak acorns given their consumption observed during the habituation phase (~6 months). We calculated a fruit and seed availability index (FSAI) by combining stand basal areas of trees used by the monkeys with their fruit and seed phenology (adapted from Agostini et al. 2010). We measured the fruiting and seed phenology at the end of each month (April 2011 to February 2012), calculating separate FSAI measures for indigenous (I-FSAI) and exotic plant species (E-FSAI) for each group. We monitored between one and seven individual trees from most of the fruiting species targeted by the monkeys including 9 indigenous and 13 exotic species (see species in bold in Appendix S1). These species were within the groups'

ranges (not limited to those encountered on transects) and we visually scored the fruits and seed (ripe or unripe) between 0 (none present) to 1 (full canopy) in increments of 0.1. We included measures of seeds contained in dry pods/nuts still in the tree and not measures of dry fruit. We did not measure fallen fruit and seeds. For each month, we calculated a fruiting and seed phenology score for each species weighted by the number of individuals of that species that were monitored (using the total number of individual trees monitored that month, range 25-40). We multiplied this weighted score by the stand basal areas for the species, calculated from the sum of the basal areas of individuals encountered on transects divided by the total area sampled for each group (0.6 ha). To estimate fruit and seed availability for March, during a hiatus in data collection, we calculated the mean for February and April; the results from the analyses reported here are unaffected if these data are excluded.

#### Statistical analyses

We conducted all statistical analyses in SPSS v. 22.0.0.1 (IBM, International Business Machines Corp., New York). We set out to understand the extent of monkeys' reliance on exotic plants and the seasonal and habitat drivers of this consumption using multiple linear regression models with backward removal of non-significant terms. As dependent variables, we used hours per day spent feeding (incorporating eating and foraging) on indigenous species, exotic species, indigenous fruit, exotic seeds, indigenous leaves, black wattle seeds and oak seeds. Whilst stepwise multiple regression may inflate the probability of Type I errors (false positive results) (Wilkinson 1979), particularly where there are large numbers of correlated parameters, we restricted the variables incorporated in the full model to analyse the relative effects of a small realistic set of candidate predictors. We used the following

factors in the initial models: day length, group ID, index of indigenous fruit and seed availability (I-FSAI) and index of exotic fruit and seed availability (E-FSAI). We also included rainfall (cumulative monthly values), since it is an important correlate of plant flushing and flowering (van Schaik et al. 1993), and because rainfall patterns were atypical in this study year. We did not include ambient temperature as it was linearly correlated with day length. We checked various diagnostics of model validity and stability and none indicated obvious influential cases, nor significant deviations from the assumptions of normality and homogeneity of residuals (Quinn & Keough 2002, Field 2005). Using partial correlations, we examined the similarity in the IFSAI and EFSAI profiles for both groups. Controlling for day length and group, we used Pearson's correlations to examine if feeding (hours per day) on indigenous fruits and exotic seeds was associated with activity patterns and day journey lengths. We incorporated group effects into all of our analyses to explore or account for differences between groups.

## **Ethical Note**

The project was approved by the National Zoological Gardens of South Africa's Research and Ethics Committee and the University of Fort Hare. We obtained permits from the Department of Economic Development and Environmental Affairs, and Department of Agriculture, Forestry and Fisheries, Eastern Cape Province. A registered veterinarian, A.T. Tordiffe, led the capture of monkeys for placement of radio-collars, and we obtained a National Threatened or Protected Species (TOPS) permit. Our research did not involve direct contact with monkeys during observational follows. The monkeys were habituated to the presence of one observer (KW). Since the monkeys were already entering gardens when we commenced this

study, the process of habituation did not encourage this behavior, although we note that habituation can have implications for human-monkey conflict and monkey risk-taking behavior (Williamson and Feistner 2011; Nowak et al. 2014).

### Results

# Determining the extent of seasonal variation in exotic species consumption

Over 50% of the total diets of both groups were exotic species (G1:  $58.7 \pm 27.5\%$  SD, G2:  $52.1 \pm 32.0\%$  SD), followed by indigenous species (G1:  $26.6 \pm 23.2\%$ , G2:  $38.6 \pm 29.5$ ). Other items (e.g., invertebrates, food waste) each contributed less than 5% to the monkeys' diets (Table 1). The most consumed dietary item out of the annual total was exotic seeds, followed by indigenous fruits and leaves (Table 1).

G1 ate fruit and seeds from 29 species, 19 of which were exotic, while G2 ate fruit and seeds from 27 species, 12 of which were exotic (Appendix S1). The top species targeted for fruits and seeds were exotic black wattle (G1: annual mean 22.3  $\pm$  26.3%; G2: 27.6  $\pm$  25.8%), exotic oak (combination of *Q.robur* and *Q.palustris*, G1: 13.6  $\pm$  27.0%; G2: 5.4  $\pm$  19.3%), and indigenous red currant (*Searsia chirindensis*) (G1: 8.4  $\pm$  13.7%; G2: 13.0  $\pm$  17.4%). The monkeys processed the fruit and seeds of each of those species differently (described in Appendix S1). The other dietary items each contributed less than 5% to the diet (Appendix S1).

The two groups were comparable in their seasonal food choices, eating exotic seeds year-round with peaks in indigenous fruit consumption in autumn (February – May) (Fig.1). However, in winter, G2, unlike G1, had a pronounced peak in feeding on indigenous leaves.

**Table 1**. The contribution of each food item to the diet (mean % eating scans) of two samango groups (G1, G2) in Hogsback, South Africa, (February 2011 - January 2012) with plant parts split into exotic, indigenous and unknown categories (mean  $\pm$  SD %).

Dietary item	Plant parts	G1	G2
Exotic plant species	Fruit	$4.3 \pm 5.1$	$2.2\pm4.2$
	Seed	$37.4 \pm 24.2$	$35.0\pm25.3$
	Leaf	$6.5 \pm 8.4$	$4.6\pm7.4$
	Stem	$1.6 \pm 2.0$	$1.3\pm2.5$
	Bud	$3.9\pm7.4$	$2.9\pm8.3$
	Flower	$4.9 \pm 11.7$	$5.8 \pm 15.4$
	Exudate	$0.2\pm0.6$	$0.5\pm0.7$
Indigenous plant species	Fruit	$12.5 \pm 18.7$	$18.1\pm20.8$
	Seed	$3.4\pm7.7$	$2.3\pm4.2$
	Leaf	$5.7\pm7.5$	$11.0\pm12.7$
	Stem	$1.2 \pm 1.9$	$1.1 \pm 2.01$
	Bud	$2.6 \pm 7.7$	$4.3\pm9.4$
	Flower	$1.1 \pm 2.4$	$1.8 \pm 3.1$
Unknown plant species	Total	< 0.1	<0.1
Invertebrates		$4.8\pm6.9$	$2.8\pm2.7$
Other: clay, fungus, lichen		-	$0.7 \pm 2.8$
Human food waste		$1.5 \pm 3.0$	<0.1
Unidentified items		$4.7\pm4.6$	3.7 ± 4.5

## Determining the extent of seasonal changes in exotic plant availability

Each group's home range contained over two-thirds matrix habitat (Table 2). The areas of indigenous habitat in both groups' home ranges had similar tree densities and species richness, but G2 had larger trees (>40 cm DBH) and more red currant trees (Table 2) than G1. Meanwhile, the matrix component was markedly different for the two groups. In G1's home range, the matrix had a low tree density and species richness, and many large trees compared with the matrix available to G2 (Table 2). The gardens within Hogsback generally had large, open grassy areas and stands of old, emergent exotic trees to which only G1 had access. Both groups had large stands

of black wattle in their home ranges, but G2 had a relatively higher density of wattle. Meanwhile, G2 had relatively few oaks in its home range compared with G1.



**Figure 1**. Monthly variation in time spent feeding (eating and foraging) on indigenous fruit, exotic seeds and indigenous leaves by two groups of samango monkeys in Hogsback, South Africa: a) G1 and b) G2 (February 2011 - January 2012).

The seasonal pattern of availability was similar for both indigenous and exotic plant species, but exotics tended to have lower availability year round (Fig. 2). Both groups experienced similar extreme seasonal changes in food availability (Fig. 2) with highly correlated I-FSAI profiles for the two groups (r=0.86, N=12, P<0.001). I-FSAI peaked in summer-autumn (January-April), with a trough in indigenous fruit over winter (June-August). E-FSAI also peaked in late summer (February-March), and declined during winter (June-August) (Fig. 2). However, from September to December, E-FSAI in G2 habitat increased more than in G1 habitat, such that the E-FSAI profiles for the two groups were not correlated (r=0.38, N=12, P=0.23). These results reflect the differences in tree density and species richness between the groups' ranges (particularly black wattle and oak).

**Table 2.** Home range size and environmental characteristics of indigenous and matrix habitat types for two groups of samango monkeys (G1, G2) in Hogsback, South Africa, (February 2011 - January 2012) including densities of important indigenous (red currant) and exotic (black wattle and oak) food species.

Home range characteristics	G1	G2
Home range size	27.9 ha	22.0 ha
Percentage matrix habitat	67 %	73 %
Indigenous habitat:		
Tree density	1593 trees/ha	1467 trees/ha
Species richness	34 species	26 species
% trees of >40 cm DBH	7.3%	10.0%
Red currant	33 trees/ha	50 trees/ha
Matrix habitat:		
Tree density	480 trees/ha	1053 trees/ha
Species richness	23 species	32 species
% trees of >40 cm DBH	29.2%	9.2%
Black wattle tree density	30 trees/ha	283 trees/ha
Oak tree density	53 trees/ha	0 trees/ha*

\* No oak trees were encountered along transects as there was only one small grove of oaks planted within this group's home range.

# Determining the drivers of exotic plant consumption

Despite the prevalence of exotic species in the diet, indigenous food availability was the primary determinant of feeding (Table 3). Both groups spent more time feeding on indigenous plant species as more indigenous fruit and seeds were available. As fewer indigenous fruits and seeds became available and days became longer, monkeys switched to exotic species (Table 3), but exotic food availability did not drive overall food choices (Appendix S2). G2 fed significantly more on indigenous species than G1. Rainfall was not significant in either model.



**Figure 2.** Monthly variation in the fruit and seed availability index for indigenous (I-FSAI; dotted) and exotic (E-FSAI; solid) plant species for two groups of samango monkeys in Hogsback, South Africa, G1 (black) and G2 (grey), from February 2011 to January 2012.

Time spent feeding on indigenous fruit increased with I-FSAI, even as day length decreased (in autumn and winter) (Table 4). Indigenous leaves were eaten when exotic fruit and seed availability was low and when days were shorter in winter (Table 4). However, G2 ate more indigenous leaves than G1 (Table 4, Fig.1). Monkeys from both groups ate more exotic seeds when days were longer (in summer), independent of when exotic or indigenous fruits and seeds were most available.

**Table 3.** Linear regression models of time spent feeding (eating and foraging) on indigenous and exotic species each day by samango monkeys in Hogsback, South Africa, (Feb 2011- Jan 2012), with day length, indigenous fruit and seed availability (I-FSAI) and group as variables. Only variables that contributed significantly to the model are shown, with the full model provided in Appendix S2.

Indigenous Species Feeding (hours per day)					
r <sup>2</sup>	df	F	Р		
0.42	3,65	15.69	< 0.001		
Model	В	Т	р		
Constant	6.02	4.76	< 0.001		
Day length	-0.54	-5.32	< 0.001		
I-FSAI	2.33	4.31	< 0.001		
Group	0.77	2.76	0.008		
	<b>Exotic Species Fee</b>	ding (hours per	day)		
$\mathbf{r}^2$	df	F	Р		
0.57	2,66	43.05	< 0.001		
Model	В	Т	Р		
Constant	-6.72	-4.74	< 0.001		
Day length	1.01	8.18	< 0.001		
I-FSAI	-4.19	-6.42	< 0.001		

When we considered black wattle and oak seed consumption separately, different patterns emerged for exotic seeds. Monkeys' diets consisted of over 95% black wattle seeds on some days (mostly during November-December), while oak acorns accounted for over 50% of the diet on some days in winter. The time that monkeys devoted to feeding on black wattle seeds increased with day length but declined with increasing I-FSAI (Table 5), suggesting consumption increased when preferred natural fruits were unavailable. In contrast, consumption of oak acorns (especially by G1) increased as day lengths became shorter. Oak seed consumption was not influenced by I-FSAI but increased with E-FSAI (Table 5). Interestingly, this result

appears to be driven by G2 that had only one small stand of oak trees in its range and acorn availability was thus not reflected in the E-FSAI estimate for this group.

**Table 4.** Linear regression models of time spent feeding (eating and foraging) on indigenous fruit, exotic seed and indigenous leaves by samango monkeys in Hogsback, South Africa, (Feb 2011- Jan 2012) with day length, indigenous fruit and seed availability (I-FSAI), exotic fruit and seed availability (E-FSAI) and group as variables. Only variables that contributed significantly to the model are shown, with full models in Appendix S3.

Indigenous Fruit Feeding (hours per day)					
$\mathbf{r}^2$	df	F	Р		
0.76	2,66	46.11	< 0.001		
Model	В	Т	Р		
Constant	1.67	3.04	0.003		
I-FSAI	2.40	9.51	< 0.001		
Day length	-0.18	-3.87	< 0.001		
	<b>Exotic Seed Feedin</b>	ng (hours per da	ay)		
$\mathbf{r}^2$	df	F	Р		
0.70	1,67	64.75	< 0.001		
Model	В	Т	Р		
Constant	-7.52	-6.58	< 0.001		
Day length	0.77	8.05	< 0.001		
•	Leaf Feeding (hours	per day)			
$\mathbf{r}^2$	df	F	Р		
0.32	3,65	10.24	< 0.001		
Model	В	Т	Р		
Constant	1.28	1.99	0.051		
Day length	-0.11	-2.04	0.046		
Group	0.30	3.43	0.001		
E-FSAI	-0.88	-2.31	0.024		

**Table 5.** Linear regression models of time spent feeding (eating and foraging) on exotic black wattle (*Acacia mearnsii*) and oak (*Quercus robur* and *palustris*). seeds by samango monkeys in Hogsback, South Africa, (Feb 2011- Jan 2012) with day length, indigenous fruit and seed availability (I-FSAI), exotic fruit and seed availability (E-FSAI) and group as variables. Only variables that contributed significantly to the model are shown, with full models in Appendix S4.

Black Wattle Seed Feeding (hours per day)					
$\mathbf{r}^2$	df	F	Р		
0.50	2,66	33.47	< 0.001		
Model	В	Т	Р		
Constant	-9.25	-6.49	< 0.001		
Day length	1.00	8.09	< 0.001		
I-FSAI	-2.19	-3.34	0.001		
Oak Seed Feed	ing (hours per d	ay)			
$\mathbf{r}^2$	df	$\mathbf{F}$	Р		
0.28	3,65	8.39	< 0.001		
Model	В	Τ	Р		
Constant	3.19	5.30	< 0.001		
Group	-0.48	-3.96	< 0.001		
Day length	-0.18	-3.74	< 0.001		
E-FSAI	0.95	2.69	0.009		

## Effects of exotic plant consumption on activity budgets

The time monkeys spent both eating and foraging increased with exotic seed consumption, such that feeding overall also increased significantly (Table 6). In contrast, monkeys spent less time moving when feeding on exotic seeds and this was also reflected in a reduction in their day journey lengths. There was no relationship between feeding on exotics and resting time but increased time spent feeding on exotic seeds significantly decreased the time monkeys had for socializing.

Monkeys' feeding time was unaffected by indigenous fruit consumption but both groups spent more time moving and had longer day journeys the more they fed on indigenous fruits (Table 6). Resting time decreased significantly as a result, but had no effect on social time.

**Table 6.** Partial correlations (controlling for day length and group; n = 65 days) of time in activity (feeding is a combination of eating and foraging) and day journey length (DJL) and time spent feeding on indigenous and exotic plant species by samango monkeys in Hogsback, South Africa (Feb 2011- Jan 2012). Significant relationships are highlighted in bold.

Activity		Time feeding on indigenous fruit (hours per day)	Time feeding on exotic seeds (hours per day)
Feeding (hrs)	Correlation	-0.19	0.63
	Р	0.129	< 0.001
Eating (hrs)	Correlation	0.23	0.38
	Р	0.061	0.002
Forage (hrs)	Correlation	-0.59	0.39
	Р	< 0.001	0.001
Move (hrs)	Correlation	0.75	-0.34
	Р	< 0.001	0.005
Rest (hrs)	Correlation	-0.57	-0.23
	Р	< 0.001	0.061
Social (hrs)	Correlation	0.04	-0.35
	Р	0.754	0.004
DJL (m)	Correlation	0.53	-0.26
	Р	< 0.001	0.031

# Discussion

Two groups of samango monkeys in Hogsback consumed large amounts of exotic plant species. Contrary to an expectation that consumption of exotic fruit and seeds would vary according to exotic fruit and seed availability, the availability of native foods inside indigenous forest and not availability of exotic foods was selected as the primary driver of samango monkeys' foraging decisions in our models. The monkeys ate indigenous species as they became seasonally available, while they targeted exotic plant species when indigenous food availability declined. Day length was a significant factor influencing exotic fruit and seed consumption. Furthermore, the group with greater access to the village (G1) also had greater access and made more use of exotic species. Although previous studies have suggested that monkeys spend less time foraging and moving and more time resting and socializing the more they fed on exotic fruits and seeds, this was not the case in our analysis. Instead, the more the monkeys fed on exotic seeds, the more time they spent feeding and the less time socializing; therefore, a reliance on exotic plants does not appear to free up monkeys' time. While the drivers of the monkeys' use of the matrix are complex, our analysis suggests that exotic plants represent potential "fallback" foods in this humanmodified and seasonal environment.

The availability of native foods inside indigenous forest was the primary driver of samango monkeys' foraging decisions in Hogsback. Interestingly, however, indigenous fruits contributed a small proportion to the groups' diets (max 18% of total diet), which is at the lower spectrum of what has been observed in other samango and blue monkey populations (17.0 - 54.6% reviewed by Tesfaye et al 2013, 26.3 – 91.1% reviewed by Coleman and Hill 2014). However, because native fruits were targeted as soon as they became available, they appeared to be the monkeys' preferred food item. When indigenous fruit availability declined in winter, both Hogsback groups ate indigenous leaves to extents similar to other populations of samango monkeys in South Africa (Coleman and Hill 2014; Lawes 1991). The annual mean amount of leaves eaten by Hogsback monkeys (~ 5 - 11 % of total diet) was again at the lower spectrum of the wide range reported for other populations (e.g. 1.6 - 51.6 % reviewed by Coleman and Hill 2014; 6.7 - 51.3 %, reviewed by Tesfaye et al 2013). This low to

moderate reliance on leaves may be explained by the access that these monkeys have to a matrix planted with exotic species including ones with high-quality (high protein and calorie) seeds.

Compared with other populations, the Hogsback groups have a much higher intake of seeds (~ 40 % of total diet) relative to a previously observed range of 0.6 -16.5 % (reviewed by Tesfaye et al. 2013) and can be described as "granivorous". Seed predation by frugivorous primates is viewed as uncommon and possibly a last resort fallback food item (Gautier-Hion et al. 1993; Kaplin et al. 1998). In fact, indigenous seed consumption by the Hogsback samango monkeys was low (~ 3 % of total diet). Given that consumption of exotic seeds did not increase with decreasing availability of indigenous fruits, their role as potential "fallback foods" is more complicated and best deciphered by examining monkeys' consumption of oak and black wattle seeds separately.

Black wattle seeds appear to be an important fallback food, with both groups eating more of these seeds with decreasing availability of indigenous fruit. These seeds seem to fulfill the definition of "staple" fallback foods, defined by Marshall and Wrangham (2007) as foods available year round, usually eaten throughout the entire year, and in some seasons, constituting the whole diet. Although *Acacia* sp. can have high protein (20 %) and fat content (e.g. 68 % linoleic oil), black wattle in particular has high tannin content (~ 28.8 %, reviewed by Siegler 2003) and was planted in South Africa for its high total phenolics content for the tannin industry (de Beer 1986). These tannins likely impede the digestive process, and their consumption may explain why Hogsback monkeys exhibit geophagy, which counteracts possible ill effects on digestion, as seen in chacma baboons (Pebsworth et al. 2012). Black colobus (*Colobus satanas*, McKey 1978) and Wolf's mona monkey (*Cercopithecus*  *wolfi*, Gautier-Hion et al 1993) have also been observed to eat seeds with high levels of secondary chemicals instead of mature leaves with the same chemicals because of the much higher nutritional value of seeds (more digestible, higher in fats, higher in protein). In line with Lambert's (2007) definition of two types of fallback strategies, Hogsback samango monkeys are thus choosing a high quality fallback food (seeds) that is less abundant but mechanically protected (hard covering) rather than low quality (nutrient-poor) fallback foods (mature leaves) that are abundant, but which sometimes require anatomical adaptations (e.g. samango monkeys have hindgut fermentation, Bruorton and Perrin 1988).

As the short winter days coincide with the lowest mean monthly temperatures, and lowest indigenous and exotic fruit-seed availability in Hogsback, this energetic bottleneck places a greater premium on fallback acorns at this time. In fact, day length was an important determinant of diet and behavior in virtually all analyses, confirming its importance as an ecological constraint (Hill et al. 2003, 2004; Hill 2005; van Doorn et al. 2010). During winter, the two groups appear to use different strategies to survive this period, with G1 relying heavily on acorns (fat and protein rich: Shimada and Saitoh 2006), and G2 relying more on leaves (generally low levels of soluble carbohydrate energy: Lambert 2002, Hemingway and Bynum 2005, Rosenberger 2013). G1 did not increase their consumption of acorns with decreasing availability of indigenous fruit but day length was the key determinant of acorn foraging, with consumption greatest in mid-winter when there was little other food around. Without a measure of fallen food availability, we are unable to state whether the relationship with E-FSAI, largely driven by G2, suggests there could be a preference for acorns in certain situations. As it stands, therefore, consumption of acorns is only partially consistent with Marshall and Wrangham's (2007) definition of fallback food and a more detailed analysis of seasonal acorn availability should be explored further.

Monkeys in G1 significantly increased their time foraging on the ground in gardens in winter as they were eating fallen acorns, despite increased risks from humans and domestic dogs, suggesting these seeds are high quality food items (energetic gains outweigh risks; see Nowak et al., this issue). G1 thus seems to employ a "risk-prone" strategy (Krebs and Davies 1993; Nowak et al., this volume) by searching for patchily distributed clumps of high quality foods (acorns), which are nutritionally dense but require increased handling time (Lambert 2007). G2 also ate fallen acorns but not to the extent that G1 did, and so seemed to employ a relatively "risk-averse" strategy switching to leaves which are more abundant and evenly distributed (compared to oak trees) and are generally less protein rich and higher in secondary compounds than acorns (Lawes et al. 1990; van Schaik et al. 1993). The differences in strategy employed by the two groups may be a consequence of the differences in the availability of oak trees (G2 had relatively few oaks in its home range: only one small grove) and possibly vine leaves (G2 had a higher density of large exotic trees) in the two home ranges. More ecological work, including measuring fallen fruit, would be needed to better understand the role of species availability and distribution within these two areas and how it relates to possible behavioral differences between the groups. A further possible explanation of our observed differences in groups' strategies is that G1, whose home range encompasses the village, has learned how to manage anthropogenic risk to a better extent than G2 (no village access) and as a result, exploits exotics in people's gardens in winter while G2 falls back on leaves.

Animals are expected to invest more time in foraging and eating during periods of low food diversity and availability, to feed at the expense of rest time and thus to trade-off between energy acquisition and rest (Lawes 1991; Lawes et al. 1990; van Schaik et al. 1993). Having to spend less time feeding is often associated with increased availability of nutrient dense food sources (Guo et al. 2007; Riley 2007). We expected that monkeys would spend less time foraging and moving and more time resting and socializing the more they fed on exotic fruit and seeds. Both indigenous and exotic fruiting species are patchily distributed throughout the habitat, but exotic tree species (esp. black wattle, oak) occur in large clumps which provide monkeys with high food biomass that can support a group for long periods. Monkeys spent less time moving when consuming exotics compared to indigenous species. However, the time required to extract seeds from the wattle pods and oak acorns meant more overall time spent foraging. Similar patterns have been observed in vervet monkeys eating indigenous umbrella thorn tree (Acacia tortillas) pods (Isbell and Young 1993) and chacma baboons eating exotic pine nuts (Hoffman and O'Rian 2011). The large amount of time spent foraging could also be due to time spent having to search for the seeds in the leaf litter if they have fallen to the ground. We thus found that monkeys did not have extra time for resting and actually decreased their social time when concentrating their feeding on exotics.

Despite several limitations in our study design namely 1) our vegetation sampling missing clumps of some exotic species used by monkeys, 2) our lack of foresight to measure availability of fallen fruits (mostly acorns), 3) our lack of precise data on group demography and 4) our lumping of different stages of plant parts, which may have different nutritional values and digestibility (e.g. Lambert and Rothman 2015), as well as the relatively short (one year) duration of our study, our data strongly suggest that these samango monkeys have become well-adapted to a human-modified habitat through their use of exotic plant species. We suggest that people's gradual removal of exotic plant species in the areas of monkeys' known home ranges may facilitate their return to a greater relative use of native fallback foods, as in other groups in the Amathole region which have no access to exotic plant species and still survive the winter.

One consequence of samango monkeys using matrix habitats is that they are brought into closer contact with potential risks from humans and human infrastructure. These monkeys suffer from unquantified morbidity and mortality from electrocutions, cars, pellet-gun wounds, snares and domestic dogs. Besides the known negative health consequences of living commensally with humans (Estrada et al. 2012) and eating food waste (Kemnitz et al. 2002), consumption of exotics may have detrimental effects on their health, such as increased wear on teeth, and thus affect longevity (Tordiffe et al. unpublished data). Another further possible consequence of matrix use is a compromise of monkeys' role as seed dispersers in indigenous forests (Linden et al. 2015) if they are predominately eating and dispersing exotics. Reproductive rates might become artificially high as a result of exotic species intake, as found elsewhere (Singh et al. 2001, Siex and Struhsaker 1999), possibly increasing human-primate conflict.

To manage these matrix-using samango monkey groups, we recommend 1) phasing out the planting of exotic species, 2) active planting of indigenous species in gardens and 3) removal of fallen black wattle and oak seeds where and when possible before and during the winter season. We also recommend that groups without access to exotic plant species be studied to enable comparisons with the groups we studied to

help inform management of this southernmost, behaviorally distinct and seasonally limited population of African guenons.

#### Acknowledgements

KW would like to thank Dr. A. Tordiffe, Prof J. O'Riain and Dr. S. Boyes for tremendous support during and following fieldwork. Thanks to N. Amstrong for help in conducting plant transects and plant identification, where T. Dold and T. Trinder-Smith also helped with the latter. For GIS support thanks to N. Lindenberg and T. Slingsby at UCT. Thanks to the Hogsback residents who allowed KW access to their properties to follow the monkeys. We thank Hogsback resident T. Webster and South African Weather Service (SAWS) for access and use of their ambient and rainfall data. UKZN is thanked for the lending of capture cages. Without the support of the following funding sources this project would not have been possible: Claude Leon Foundation, UCT URC Fellowship, JMasters NRF Fund, Novartis SAVF Wildlife Research Fund, Primate Conservation Inc., Mazda Wildlife Fund. KN was supported by a Durham University CO-FUND Junior Research Fellowship during the preparation of this paper. We thank Shane A. Richards for his valuable advice on analyses. Finally, we would like to thank the guest editors Noemi Spagnoletti, Kimberley Hockings, and Matt McLennan, and Joanna Setchell, and also two anonymous reviewers for their helpful comments on our manuscript.

#### References

- Agostini, I., Holzmann, I., & Di Bitetti, M.S. (2010). Are howler monkey species ecologically equivalent? Trophic niche overlap in syntopic *Alouatta guariba clamitans* and *Alouatta caraya*. *American Journal of Primatology*, 72, 173–186.
- Amato, K.R., Van Belle, S. & Wilkinson, B. (2013) A Comparison of Scan and Focal Sampling for the Description of Wild Primate Activity, Diet and Intragroup Spatial Relationships. *Folia Primatologica*, 84, 87-101.
- Anderson, J., Rowcliffe, J., & Cowlishaw, G. (2007). Does the matrix matter? A forest primate in a complex agricultural landscape. *Biological Conservation*, *135*(2), 212–222.
- Beyer, H.L. (2004). Hawth's Analysis Tools for ArcGIS. <a href="http://www.spatialecology.com/htools.>">http://www.spatialecology.com/htools.></a>.
- Bicca-Marques, J.C., & Calegaro-Marques, C. (1994). Exotic plant species can serve as staple food sources for wild howler populations. *Folia Primatologica*, 63, 209–211.
- Boyes, R.S., & Perrin, M.R. (2010). Aerial surveillance by a generalist seed predator: food resource tracking by Meyer's parrot (*Poicephalus meyeri*) in the Okavango Delta, Botswana. *Journal of Tropical Ecology*, 26, 1-12.
- Breytenbach, G.J. (1988). Diet of the samango monkey in the Northern Transvaal. South African Journal of Wildlife Research, 18, 113-119.
- Bruorton, M.R., & Perrin, M.R. (1988). The anatomy of the stomach and caecum of the samango monkey, *Cercopithecus mitis erythrarchus* Peters, 1852. Zeitschrift für Säugetierkunde, 53, 210-224.
- Chapman, C.A., Chapman, L.J., Wrangham, R., Hunt, K., Gebo, D., & Gardner, L. (1992). Estimators of fruit abundance of tropical trees. *Biotropica*, 24, 527-531.

- Chapman, C.A., Naughton-Treves, L., Lawes, M.J., Wasserman, M.D., & Gillespie, T.R. (2007). Population declines of colobus in Western Uganda and conservation value of forest fragments. *International Journal of Primatology* 28: 513–528.
- Chaves, Ó. M., Stoner, K. E., & Arroyo-Rodríguez, V. (2012). Differences in diet between spider monkey groups living in forest fragments and continuous forest in Mexico. *Biotropica*, 44(1), 105–113.
- Coleman, B.T. & Hill, R.A. (2015). Biogeographic variation in the diet and behaviour of *Cercopithecus mitis*. *Folia Primatologica*, 85, 319-334.
- Dalton, D.L., Linden, B., Wimberger, K., Nupen, L.J., Tordiffe, A.S.W., Taylor, P.J., Madisha, T.M. & Kotze, A. (2015). New insights into samango monkey speciation in South Africa. *PLoS ONE*, 10(3), e0117003. doi:10.1371/journal.pone.0117003.
- Davis, M.A. (2011). Do native birds care whether their berries are native or exotic? No. *Bioscience*, 61, 501-502.
- De Beer, H. (1986). Black Wattle. *Farming in South Africa, Weeds* A.2411986. Pretoria, South Africa: Department of Agriculture and Water Supply.
- Estrada, A. (2009). Primate conservation in South America: the human and ecological dimensions of the problem. In P.A. Garber, A. Estrada, J. Bicca-Marques, E. Heymann & K. Strier (Eds.), South American Primates: Comparative Perspectives in the Study of Behavior, Ecology, and Conservation (pp 463–505). New York: Springer Press.

- Estrada, A., Raboy, B.E., & Oliveira, L.C. (2012). Agroecosystems and primate conservation in the tropics: a review. *American Journal of Primatology*, 74, 696–711.
- Field, A. (2005). Discovering Statistics using SPSS. Sage Publications. London.
- Fischer, J., & Lindenmayer, D.B. (2007). Landscape modification and habitat fragmentation : a synthesis. *Global Ecology and Biogeography*, 16, 265–280.
- Foerster, S., Cords, M., & Monfort, S. L. (2012). Seasonal energetic stress in a tropical forest primate: proximate causes and evolutionary implications. *PloS* one, 7, e50108.
- Friedman, Y., & Daly, B. (Eds). (2004). Red Data Book of the Mammals of South Africa: a Conservation Assessment. South Africa: CBSG Southern Africa/Conservation Breeding Specialist Group (SSC/IUCN)/Endangered Wildlife Trust (EWT).
- Gabriel, D. N. (2013). Habitat use and activity patterns as an indication of fragment quality in a strepsirrhine primate. *International Journal of Primatology*, 34, 388–406.
- Gautier-Hion, A., Gautier, J.-P., & Maisels, F. (1993). Seed dispersal versus seed predation: An inter-site comparison of two related African monkeys. *Vegetatio* 107/108: 237–244.
- Guo, S., Li, B., & Watanabe, K. (2007). Diet and activity budget of *Rhinopithecus roxellana* in the Qinling Mountains, China. *Primates*, 48, 268-276.
- Harcourt, A.H. & Stewart, K.J. (1984) Gorillas time feeding aspect of methodology, body size, competition and diet. *African Journal of Ecology*, 22, 207-215.

- Hemingway, C.A., & Bynum, N. (2005). The influence of seasonality on primate diet and ranging. In D.K. Brockman & C.P. van Schaik (Eds.), *Seasonality in Primates* (pp. 57-104). Cambridge: Cambridge University Press.
- Hill, R.A. (2005). Daylength seasonality and the thermal environment. In: D.K.
  Brockman & C.P. van Schaik (Eds.), *Primate seasonality: Implications for Human Evolution* (pp. 197-213). Cambridge: Cambridge University Press.
- Hill, R.A., Barrett, L., Gaynor, D., Weingrill, T., Dixon, P., & Payne, H. (2003). Day length latitude and behavioural (in)flexibility in baboons (*Papio cynocephalus ursinus*). *Behavioral Ecology and Sociobiology*, 53, 278–286.
- Hill, R.A., Barrett, L., Gaynor, D., Weingrill, T., Dixon, P., Payne, H. & Henzi, S.P. (2004). Day length variation and seasonal analysis of behaviour. *South African Journal of Wildlife Research*, 34, 39–44.
- Hockings, K.J., Anderson, J.R., & Matsuzawa, T. (2009). Use of Wild and Cultivated
  Foods by Chimpanzees at Bossou, Republic of Guinea: Feeding Dynamics in a
  Human-Influenced Environment. *American Journal of Primatology*, 71, 636-646.
- Hoffman, T.S., & O'Riain, M.J. (2011). The spatial ecology of chacma baboons (*Papio ursinus*) in a human-modified environment. *International Journal of Primatology*, 32, 308–328.
- Iponga, D.M. (2011). Towards a predictive understanding of invasion success. In L. Zietsman (Ed.), Observations on Environmental Change in South Africa (Section 3, pp 127-130.). Stellenbosch: SUN MeDIA.
- Isbell, L.A., & Young, T.P. (1993). Social and ecological influences on activity budgets of vervet monkeys, and their implications for group living. *Behavioral Ecology and Sociobiology*, 32, 377-385.

- Kaplin, B. A. (2001). Ranging behavior of two species of guenons (Cercopithecus lhoesti and C.mitis doggetti) in the Nyungwe Forest Reserve, Rwanda. *International Journal of Primatology*, 22, 521–548.
- Kaplin, B.A., & Moermond, T.C. (2000). Foraging ecology of the mountain monkey (*Cercopithecus l'hoesti*): implications for its evolutionary history and use of disturbed forest. *American Journal of Primatology*, 50, 227–246.
- Kaplin, B.A., Munyaligoga, V., & Moermond, T.C. (1998). The influence of temporal changes in fruit availability on diet composition and seed handling in blue monkeys (*Cercopithecus mitis doggetti*). *Biotropica*, 30, 56-71.
- Kemnitz, J.W., Sapolsky, R.M., Altmann, J., Muruthi, P., Mott, G.E., & Stefanick, M.L. (2002). Effects of food availability on serum insulin and lipid concentrations in free-ranging baboons. *American Journal of Primatology*, 57, 13-19.
- Kingdon, J., Gippoliti, S., Butynski, T.M., Lawes, M.J., Eeley, H., Lehn, C. & De Jong, Y. (2008). *Cercopithecus mitis*. The IUCN Red List of Threatened Species 2008: e.T4221A10676022. Available at: www.iucnredlist.org
- Krebs, J.R. & Davies, N.B., Eds. (1993). Economic decisions and the individual. In:
   Introduction to Behavioural Ecology (3<sup>rd</sup> Edition, pp 48-76.). Oxford: Blackwell
   Scientific Publications.
- Lambert, J.E. (2002). Digestive retention times in forest guenons (*Cercopithecus* spp.) with reference to chimpanzees (*Pan troglodytes*). International Journal of *Primatology*, 23, 1169-1185.
- Lambert, J.E. (2007). Seasonality, fallback strategies, and natural selection: a chimpanzee and cercopithecoid model for interpreting the evolution of hominin diet. In: P.S. Ungar (Ed.), *Evolution of the Human Diet: the Known, the*

Unknown, and the Unknowable (pp 324–343). Oxford: Oxford University Press.

- Lambert, J. E., & Rothman, J. M. (2015). Fallback Foods, Optimal Diets, and Nutritional Targets: Primate Responses to Varying Food Availability and Quality. In D. Brenneis & K. B. Strier (Eds.), Annual Review of Anthropology, Vol 44 (Vol. 44, pp. 493-512). Palo Alto: Annual Reviews
- Lawes, M.J. (1991). Diet of samango monkeys (*Cercopithecus mitis erythrarchus*) in the Cape Vidal dune forest, South Africa. *Journal of Zoology*, 224, 149-173.
- Lawes, M.J. (1992). Estimates of population density and correlates of the status of the samango monkey *Cercopithecus mitis* in Natal, South Africa. *Biological Conservation*, 60, 197-210.
- Lawes, M.J. (2002). Conservation of fragmented populations of *Cercopithecus mitis* in South Africa: the role of reintroduction, corridors and metapopulation ecology In M.E. Glenn & M. Cords (Eds.), *The Guenons: Diversity and Adaptation in African Monkeys* (pp 375–392). New York: Kluwer Academic/ Plenum Publishers.
- Lawes, M.J., Henzi, S.P., & Perrin, M.R. (1990). Diet and feeding behaviour of samango monkeys (*Cercopithecus mitis labiatus*) in Ngoye Forest, South Africa. *Folia Primatologica*, 54, 57-69.
- Lawes, M.J., Mealin, P.E., Piper, S.E. (2000). Patch occupancy and potential metapopulation dynamics of three forest mammals in fragmented afromontane forest in South Africa. *Conservation Biology*, 14, 1088-1098.
- Lee, D.C., Powell, V.J., & Lindsell, J.A. (2014). The conservation value of degraded forests for agile gibbons *Hylobates agilis*. *American Journal of Primatology*, 77, 76–85.

- Linden, B., Linden, J., Fischer, F., & Linsenmair, K. E. (2015). Seed dispersal by South Africa's only forest-dwelling guenon, the samango monkey (*Cercopithecus mitis*). *African Journal of Wildlife Research* 45, 88–99.
- Linden, B., Wimberger, K., Ehlers-Smith, Y., Howlett, C., and Child, M. (2016). A Conservation Assessment of *Cercopithecus albogularis*. In M.F. Child, E. Do
  Linh San, D. Raimondo, H. Davies-Mostert & L. Roxburgh (Eds.), *The Red List* of Mammals of South Africa, Swaziland and Lesotho. South Africa: South African National Biodiversity Institute and Endangered Wildlife Trust.
- Marshall, A. J., & Wrangham, R. W. (2007). Evolutionary Consequences of Fallback Foods. *International Journal of Primatology*, 28, 1219–1235.
- McKey, D.B. (1978). Soils, vegetation, and seed-eating by black colobus monkeys. InG.G. Montgomery (Ed.), *The Ecology of Arboreal Folivores* (pp 423–437).Washington, DC: Smithsonian Institution Press.
- Mckinney, T., Westin, J. L., & Serio-Silva, J. C. (2015). Anthropogenic Habitat Modification, Tourist Interactions and Crop-Raiding in Howler Monkeys. In M.M. Kowalewski, P.A. Garber, L. Cortéz, B. Urbani, & D. Youlatos (Eds.), *Howler Monkeys, Developments in Primatology: Progress and Prospects* (pp 281-311). New York: Springer Science+Business Media.
- McMahon, R.P. (1977). Aspects of the Behaviour of Samango Monkeys (Cercopithecus (mitis) albogularis labiatus) [dissertation]. Durban: University of Natal
- Mokotjomela, T. M., Musil, C. F., & Esler, K. J. (2013). Do frugivorous birds concentrate their foraging activities on those alien plants with the most abundant and nutritious fruits in the South African Mediterranean-climate region? *Plant Ecology*, 214, 49–59.

- Mucina, L., & Geldenhuys, C.J. (Eds) (2006). Afrotemperate, subtropical and azonal forests. In: *The Vegetation of South Africa, Lesotho and Swaziland* (pp 585-614, Strelitzia 19). Pretoria: South African National Biodiversity Institute.
- Nowak, K., le Roux, A., Richards, S.A., Scheijen, C.P.J., Hill, R.A. (2014). Human observers impact habituated samango monkeys' perceived landscape of fear. *Behavioral Ecology* 25: 1199-1204.
- Pebsworth, P. A, Bardi, M., & Huffman, M. A. (2012). Geophagy in chacma baboons: patterns of soil consumption by age class, sex, and reproductive state. *American Journal of Primatology*, 74, 48–57.
- Pozo-Montuy, G., Serio-Silva, J. C., Chapman, C. A., & Bonilla-Sánchez, Y. M. (2013). Resource Use in a landscape matrix by an arboreal primate: evidence of supplementation in black howlers (Alouatta pigra). *International Journal of Primatology*, 34(4), 714–731.
- Quinn, G.P. & Keough, M.J. (2002). Experimental Designs and Data Analysis for Biologists. Cambridge University Press. Cambridge.
- Renton, K. (2001). Lilac-crowned parrot diet and food resource availability: resource tracking by a parrot seed predator. *The Condor*, 103, 62–69.
- Reynoso-Cruz, J.E., Rangel-Negrin, A., Coyohua-Fuentes, A., Canales-Espinosa, D.,
  & Dias, P.A. (2016) Measures of food intake in mantled howling monkeys. *Primates*, 57, 161-166.
- Richardson, D. M., & van Wilgen, B. W. (2004). Invasive alien plants in South Africa: how well do we understand the ecological impacts? *South African Journal Of Science*, 100, 45–52.

- Riley, E.P. (2007). Flexibility in diet and activity patterns of *Macaca tonkeana* in response to anthropogenic habitat alteration. *International Journal of Primatology*, 28, 107–133.
- Rosenberger, A.L. (2013). Fallback foods, preferred foods, adaptive zones, and primate origins. *American Journal of Primatology*, 75, 883-890
- Saj, T., Sicotte, P., & Patterson, J.D. (1999). Influence of human food consumption on the time budget of vervets. *International Journal of Primatology*, 20, 977-994.
- Schlaepfer, M. A., Sax, D. F., & Olden, J. D. (2011). The potential conservation value of non-native species. *Conservation Biology*, 25(3), 428–437.
- Seigler, D.S. (2003). Phytochemistry of Acacia—sensu lato. *Biochemical Systematics* and Ecology, 31: 845–873.
- Shimada, T., & Saitoh, T. (2006). Re-evaluation of the relationship between rodent populations and acorn masting: a review from the aspect of nutrients and defensive chemicals in acorns. *Population Ecology*, 48, 341–352.
- Snaith, T. V, & Chapman, C. A. (2008). Red colobus monkeys display alternative behavioral responses to the costs of scramble competition. *Behavioral Ecology*, *19*(6), 1289–1296.
- Siex, K.S. & Struhsaker, T.T. (1999). Ecology of the Zanzibar red colobus monkey: Demographic variability and habitat stability. *International Journal of Primatology*, 20, 163-192.
- Singh, M., Kumara, H. N., Kumar, M. A., & Sharma, A. K. (2001). Behavioural responses of Lion-tailed macaques (*Macaca silenus*) to a changing habitat in a tropical rain forest fragment in the Western Ghats, India. *Folia Primatologica*, 72(5), 278–291.

- Skinner, J.D., & Chimimba, C.T., eds. (2005). Sykes' monkey. In: *The Mammals of the Southern Africa Subregion* (3<sup>rd</sup> Edition, pp 227-230). Cape Town: Cambridge University Press.
- Strum, S.C. (2010). The development of primate raiding: implications for management and conservation. *International Journal of Primatology*, 31, 133-156.
- Tesfaye, D., Fashing, P.J., Bekele, A., Mekonnen, A., & Atickem, A. (2013). Ecological flexibility in Boutourlini's blue monkeys (*Cercopithecus mitis boutourlinii*) in Jibat Forest, Ethiopia: a comparison of habitat use, ranging behavior, and diet in intact and fragmented forest. *International Journal of Primatology*, 34, 615–640.
- Van Doorn, A.C., O'Riain, M.J., & Swedell, L. (2010). The effects of extreme seasonality of climate and day length on the activity budget and diet of semicommensal chacma baboons (*Papio ursinus*) in the Cape Peninsula of South Africa. *American Journal of Primatology*, 72, 104–112.
- Van Schaik, C.P., Terborgh, J.W., & Wright, S.J. (1993). The phenology of tropical forests: adaptive significance and consequences for primary consumers. *Annual Review of Ecology and Systematics*, 24, 353-377.
- Van Wilgen, B.W. (2011). Invasive alien species-introduction. In L. Zietsman (Ed.), *Observations on Environmental Change in South Africa* (Section 3, pp 125-126). Stellenbosch: SUN MeDIA.
- Van Wilgen, B.W. & Impson, M.D. (2011). Are invasive aliens a real threat to biodiversity in South Africa? In: L. Zietsman (Ed.), Observations on Environmental Change in South Africa (Section 3, pp 130-134). Stellenbosch: SUN MeDIA.

- Vitousek, P. M. (1990). Biological invasions and ecosystem processes : Towards an integration of population biology and ecosystems studies. *Oikos*, *57*(1), 7–13.
- Vitule, J.R.S., Freire, C.A., Vazquez, D.P., Nunez, M.A., & Simberloff, D. (2012).
  Revisiting the potential conservation value of non-native species. *Conservation Biology*, 26, 1153-1155.
- Von dem Bussche, G.H. & Van der Zee, D. (1985). Damage by Samango monkeys, *Cercopithecus (mitis) albogularis*, to pine trees in the Northern Transvaal. South African Forestry Journal, 134, 47-51.
- Webster, T. (2009). The Story of Hogsback. Hogsback: Trevor Webster.
- Williamson, E.A. & Feistner, A.T.C. (2011). Habituating primates: Processes, techniques, variables and ethics. In: Setchell J.M., Curtis D.J. (Eds.). *Field and Laboratory Methods in Primatology: A Practical Guide*. Cambridge (UK): Cambridge University Press. pp. 33–50.
- Wilkinson, L. (1979). Tests of significance in stepwise regression. Psychological Bulletin, 86, 168–174.
- Wilson, A.-L., & Downs, C. T. (2012). Fruit nutritional composition and non-nutritive traits of indigenous South African tree species. *South African Journal of Botany*, 78, 30–36.
- Wimberger, K., & Bidner, L. (2012). A short guide to living with samango monkeys and baboons in Hogsback. URL: www.imfene.org/sites/default/files/babooninterface-resources/living-with-samangos-and-baboons.pdf
- Zinner, D. (1999) Relationship between feeding time and food intake in hamadryas baboons (*Papio hamadryas*) and the value of feeding time as predictor of food intake. *Zoo Biology*, 18(6), 495-505.

# **Supplementary Material**

Appendix S1. Annual percentage consumption of fruit and seeds by species (excluding non-woody plants e.g. granadilla vine, Passiflora ligularis) and origin (exotic and indigenous) by two groups of samango monkeys (G1, G2) in Hogsback, South Africa, (Feb 2011-Jan 2012). Species superscripted with a star (\*) indicate that these plant parts are also eaten by humans. Species in bold are those included in fruitseed availability indices ("FSAI"; refer to methods). The three most consumed species are processed very differently: Black wattle (Acacia mearnsii) seeds are eaten from pods (autumn, spring, summer), which occur as clumps on the tree, and as fallen seeds on the ground (in winter, when monkeys have to search for them in the leaf litter). In the tree, the monkeys grab one pod off the tree and use their teeth to tear one side of the pod off. Monkeys then eat an exposed seed before opening the pod further to reveal more seeds. Oak (Quercus robur and palustris) acorns are eaten in the tree (autumn, summer) and on the ground (winter, spring). The back molars are used to crack open the acorn shell and the nut inside is eaten. Monkeys search for fallen acorns in the leaf litter. Red currant (Searsia chirindensis) fruit are eaten singly from an inflorescence on a tree.

Species	Origin	Fruit or Seed	G1	G2
			Annual % diet	Annual % diet
Acacia decurrens	Exotic	Seed	<0.1	-
Acacia mearnsii	Exotic	Seed	22.3 ± 26.3	27.6 ± 25.8
Acacia melanoxylon	Exotic	Seed	-	1.0 ± 3.8

Betula pendula	Exotic	Seed	0.4 ± 1.4	0.3 ± 1.1
Burchelia bubalina	Indigenous	Fruit	-	<0.1
Calodendrum capense	Indigenous	Seed	<0.1	<0.1
Carya illinoinensis *	Exotic	Seed	-	0.1 ± 0.4
Castanea sativa *	Exotic	Seed	0.5 ± 2.0	<0.1
Celtis africana	Indigenous	Fruit	2.6 ± 5.4	2.4 ± 5.3
Citrus limon *	Exotic	Fruit	0.4 ± 1.1	1.0 ± 4.0
Cornus kousa	Exotic	Fruit	0.3 ± 1.1	0.2 ± 0.6
Corylus sp. *	Exotic	Seed	0.1 ± 0.5	-
Cotoneaster sp.	Exotic	Fruit	0.1 ± 0.3	-
Crataegus monogyna	Exotic	Fruit	0.6 ± 2.8	<0.1
Diospyros whyteana	Indigenous	Seed	-	0.1 ± 0.3
Eugenia zeyheri	Indigenous	Fruit	0.1 ± 0.7	-
Grewia occidentalis	Indigenous	Fruit	<0.1	-
Gymnosporia buxifolia	Indigenous	Seed	<0.1	0.1 ± 0.3
Hakea salicifolia	Exotic	Seed	0.1 ± 0.3	0.3 ± 0.8
Halleria lucida	Indigenous	Fruit	$0.2 \pm 0.7$	0.3 ± 0.8
Kiggelaria africana	Indigenous	Seed	0.1 ± 0.5	0.1 ± 0.3
Malus floribunda *	Exotic	Fruit	0.4 ± 1.7	-
Persea americana *	Exotic	Fruit	0.1 ± 0.6	<0.1
Pinus patula	Exotic	Seed	<0.1	0.1 ± 0.8
Podocarpus latifolius	Indigenous	Fruit	0.5 ± 1.9	0.3 ± 1.9

Prunus armeniaca *	Exotic	Fruit	<0.1	-
Prunus cerasifera *	Exotic	Fruit	0.5 ± 2.1	0.2 ± 1.4
Prunus persica *	Exotic	Fruit	0.5 ± 1.8	0.1 ± 0.5
Prunus serotina	Exotic	Fruit	0.5 ± 1.9	-
Quercus sp (robur+palustris)	Exotic	Seed	13.6 ± 27.0	5.4 ± 19.3
Rapanea melanophloeos	Indigenous	Fruit	0.1 ± 0.7	1.3 ± 3.4
Searsia chirindensis	Indigenous	Fruit	8.4 ± 13.8	13.0 ± 17.4
Solanum mauritianum	Exotic	Fruit	0.2 ± 0.5	0.2 ± 0.5
Trichocladus ellipticus	Indigenous	Seed	2.6 ± 7.1	0.1 ± 0.4
Xymalos monospora	Indigenous	Fruit	-	0.2 ± 1.3

Appendix S2. Full linear regression models of time spent feeding (eating and foraging) on indigenous and exotic species each day by samango monkeys in Hogsback, South Africa, (Feb 2011- Jan 2012), with day length, indigenous fruit and seed availability (I-FSAI) and group as the only significant variables.

Indigenous Species Feeding (hours per day)						
$\mathbf{r}^2$	df	$\mathbf{F}$	р			
0.44	5.63	9.78	< 0.001			
Model	В	Т	р			
Constant	4.98	3.21	0.002			
Day length	-0.48	-4.00	< 0.001			
I-FSAI	2.85	3.74	< 0.001			
E-FSAI	-1.20	-1.00	0.319			
Group	0.86	2.90	0.005			
Rainfall	0.00	0.87	0.389			
Exoti	ic Species Feeding	g (hours per day	y)			
$\mathbf{r}^2$	df	$\mathbf{F}$	р			
0.58	5.63	17.28	< 0.001			
Model	В	Т	р			
Constant	-6.40	-3.39	0.001			
Day length	0.97	6.59	< 0.001			
I-FSAI	-4.60	-4.95	< 0.001			
E-FSAI	0.89	0.61	0.542			
Group	0.23	0.65	0.519			
Rainfall	-0.00	-0.78	0.437			

**Appendix S3**. Full linear regression models of time spent feeding (eating and foraging) on indigenous fruit, exotic seeds and indigenous leaves by samango monkeys in Hogsback, South Africa, (Feb 2011-Jan 2012) with day length, indigenous fruit and seed availability (I-FSAI), exotic fruit and seed availability (E-FSAI) and group as the only significant variables.

Indigenous Fruit Feeding (hours per day)				
$\mathbf{r}^2$	df	F	Р	
0.58	5.65	17.28	< 0.001	
Model	В	Т	Р	
Constant	1.97	2.10	0.040	
Day length	-0.26	-3.52	< 0.001	
Group	0.23	1.14	0.193	
E-FSAI	0.06	0.09	0.931	
I- <u>FSAI</u>	2.84	6.15	< 0.001	
Rainfall	0.00	1.46	0.148	
E	xotic Seed Feedi	ng (hours per da	y)	
$\mathbf{r}^2$	df	F	Р	
0.48	5,65	11.65	< 0.001	
Model	В	Т	Р	
Constant	-8.65	-4.66	< 0.001	
Day length	0.90	6.21	< 0.001	
I-FSAI	-1.53	-1.68	0.099	
Group	0.23	0.64	0.525	
E-FSAI	-0.40	-0.28	0.799	
Rainfall	0.00	2.05	0.044	
	igenous Leaf Fee	eding (hours per	day)	
$\mathbf{r}^2$	df	F	Р	
0.32	5,65	6.08	< 0.001	
Model	В	Т	Р	
Constant	1.30	1.85	0.068	
Day length	-0.11	-2.05	0.044	
Group	0.43	3.25	0.002	
E-FSAI	-0.78	-1.44	0.153	
I-FSAI	-0.08	-0.23	0.821	
Rainfall	0.00	0.61	0.546	

**Appendix S4.** Full linear regression models of time spent feeding (eating and foraging) on exotic black wattle (*Acacia mearnsii*) and oak (*Quercus robur* and *palustris*) seeds by samango monkeys in Hogsback, South Africa, (Feb 2011-Jan 2012) with day length, indigenous fruit and seed availability (I-FSAI), exotic fruit and seed availability (E-FSAI) and group as the only significant variables.

Black Wattle Seed Feeding (hours per day)				
$\mathbf{r}^2$	df	F	Р	
0.56	5,65	15.94	< 0.001	
Model	В	Т	Р	
Constant	-11.50	-6.34	< 0.001	
Day length	1.06	7.51	< 0.001	
I-FSAI	-1.71	-1.92	0.060	
Group	0.63	1.81	0.075	
E-FSAI	-1.27	-0.90	0.370	
Rainfall	0.00	1.99	0.051	
	Oak Seed Feedin	g (hours per day	r)	
<b>r</b> <sup>2</sup>	df	F	Р	
0.29	5,65	5.18	< 0.001	
Model	В	Т	Р	
Constant	2.96	4.56	< 0.001	
Group	-0.45	-3.64	0.001	
Day length	-0.17	-3.39	0.001	
I-FSAI	0.33	1.03	0.305	
E-FSAI	0.58	1.16	0.249	
Rainfall	0.00	-0.23	0.820	