

# Transpiration from subarctic deciduous woodlands: environmental controls and contribution to ecosystem evapotranspiration

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- 2 contribution to ecosystem evapotranspiration.
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#### 1 Abstract

Potential land-climate feedbacks in subarctic regions, where rapid warming is driving forest expansion into the tundra, may be mediated by differences in transpiration of different plant functional types. Here we assess the environmental controls of overstorey transpiration and its relevance for ecosystem evapotranspiration in subarctic deciduous woodlands. We measured overstorey transpiration of mountain birch canopies and ecosystem evapotranspiration in two locations in northern Fennoscandia, having dense (Abisko) and sparse (Kevo) overstories. For Kevo, we also upscale chamber-measured understorey evapotranspiration from shrubs and lichen using a detailed land cover map. Sub-daily evaporative fluxes were not affected by soil moisture, and showed similar controls by vapour pressure deficit and radiation across sites. At the daily timescale, increases in evaporative demand led to proportionally higher contributions of overstorey transpiration to ecosystem evapotranspiration. For the entire growing season, the overstorey transpired 33% of ecosystem evapotranspiration in Abisko and only 16% in Kevo. At this latter site, the understorey had a higher leaf area index and contributed more to ecosystem evapotranspiration compared to the overstorey birch canopy. In Abisko, growing season evapotranspiration was 27% higher than precipitation, consistent with a gradual soil moisture depletion over the summer. Our results show that overstorey canopy transpiration in subarctic deciduous woodlands is not the dominant evaporative flux. However, given the observed environmental sensitivity of evapotranspiration components, the role of deciduous trees in driving ecosystem evapotranspiration may increase with the predicted increases in tree cover and evaporative demand across subarctic regions. 

## 54 Keywords

- 55 Arctic, branch cuvettes, eddy covariance, evapotranspiration partitioning, mountain
- 56 birch, tundra, understorey

## 8 Introduction

Northern high latitudes (boreal and arctic biomes) exert an important influence in global biosphere-atmosphere interactions involving water, energy and composition. These interactions are globally relevant because of the large extent of these biomes (arctic tundra and boreal forest cover ca. 1.24·10<sup>8</sup> km<sup>2</sup>) and the intense and rapid warming occurring at northern high latitudes (0.5 K/decade since 1979; IPCC, 2013), which is partly driven by regional positive feedbacks (Chapin et al., 2000). Warmer temperatures and longer growing seasons are already inducing poleward and altitudinal treeline migration and shrub expansion in the tundra zone, which may in turn drive considerable land-atmosphere feedbacks in these latitudes (Kattsov et al., 2005; Swann, Fung, Levis, Bonan, & Doney, 2010; Zhang et al., 2013) Treelines across the subarctic vegetation belt are largely dominated by conifers, although deciduous broadleaves occupy 18% of the forest area at latitudes above 60° across Eurasia (Krankina et al., 2010) and can form the tundra-to-forest transition in many subarctic regions with oceanic influence (Callaghan et al., 2005). The area of deciduous broadleaf woodlands is increasing throughout the subarctic region (Hofgaard, Tømmervik, Rees, & Hanssen, 2013; Rundqvist et al., 2011; Tømmervik et al., 2004; Wang et al., 2019), following a general trend of increasing deciduous vegetation at northern high latitudes (Myers-Smith et al., 2011). These vegetation changes are predicted to continue in the future (Mekonnen, Riley, Randerson, Grant, & Rogers, 2019) and may cause substantial land-climate feedbacks mediated by changes in albedo, in carbon sequestration and in evaporative fluxes (Bonan, 2008; Bonfils et al., 2012). Higher transpiration rates by deciduous broadleaf forests could lead to stronger evaporative cooling locally (Chapin et al., 2000), although, in a regional context, the effects of the expansion of deciduous broadleaf trees into the tundra zone can be more complex and actually enhance Arctic warming (Swann et al., 2010). Moreover, increased soil moisture uptake by deciduous trees could lead to faster depletion of snowmelt water during the shoulder season, triggering further hydrological changes (Young-Robertson, Bolton, Bhatt, Cristóbal, & Thoman, 2016). Therefore, a greater understanding of the magnitudes and controls of evapotranspiration in deciduous woodlands is needed to predict future changes in land-atmosphere interactions in subarctic forest-tundra ecotones.

Syntheses addressing magnitudes and drivers of ecosystem evapotranspiration ( $ET_{eco}$ ) at northern high latitudes show a paucity of data for deciduous broadleaf forests from subarctic locations (Brümmer et al., 2011; Kasurinen et al., 2014; McFadden, Eugster, & Chapin III, 2003). These syntheses show that leaf area index (LAI), meteorological conditions and physiological regulation by vegetation are the three major factors affecting  $ET_{eco}$  in northern high-latitude ecosystems. In these ecosystems, evapotranspiration is largely driven by vapour pressure deficit (VPD), radiation and temperature, with soil moisture often playing a minor role (Beringer, Chapin, Thompson, & McGuire, 2005; Brümmer et al., 2011). In deciduous forests, growing season duration also affects seasonal evapotranspiration through the influence on LAI phenology (Brümmer et al., 2012). Deciduous broadleaf forests from northern high latitudes show higher evapotranspiration rates compared to conifer forests in the same region (Brümmer et al., 2011; Kasurinen et al., 2014), but they may also display a stronger stomatal control with increasing VPD (Welp, Randerson, & Liu, 2007). However, to what extent do these patterns in the drivers of  $ET_{eco}$  from northern high-latitude deciduous forests reflect the transpiration regulation by the main canopy? 

The partitioning of  $ET_{eco}$  into transpiration and evaporation and the factors controlling this partitioning are still poorly known (Schlesinger & Jasechko, 2014). Subarctic and northern boreal woodlands typically show a low LAI of the dominant canopy species, meaning that the contribution of understorey and soil evaporation to ecosystem evapotranspiration may be moderate to high (Blanken et al., 2001; Iida et al., 2009; Lafleur, 1992), although it will depend on vegetation structure (Beringer et al., 2005). This substantial contribution of the soil and understorey to  $ET_{eco}$  implies that eddy flux-based estimates of  $ET_{eco}$  in these forests may well represent the mix of physical and biological controls on evaporative fluxes and will only partially capture the physiological regulation exerted by the main canopy (Ikawa et al., 2015; Kasurinen et al., 2014). Evaporative fluxes of overstorey, understorey and the forest floor may have contrasting hydroclimatic responses (Iida et al., 2009) and a strong seasonal variation (Blanken et al., 2001). Although several studies have addressed the magnitudes and drivers of the different components of  $ET_{eco}$  in northern boreal and subarctic forests (Blanken et al., 2001; Grelle, Lundberg, Lindroth, Morén, & Cienciala, 1997; Iida et al., 

2009; Ikawa et al., 2015), we are not aware of any study of these characteristics from subarctic deciduous woodlands.

In this article, we quantify the magnitude and seasonal controls on  $ET_{eco}$  and on the transpiration of the main canopy in two deciduous broadleaf woodlands dominated by mountain birch (Betula pubescens spp. czerepanovii (Orlova) Hamet- Ahti). This is a representative species of subarctic woodlands covering 600000 ha throughout northern Fennoscandia (Haapanala et al., 2009). The Abisko site (N Sweden) displays a denser birch woodland compared to the sparser Kevo site (N Finland), which is also slightly colder and wetter. Therefore, the Abisko woodland would be representative of denser canopies which are becoming common across the subarctic in response to warming and reduced browsing (Callaghan et al., 2013). In both sites, we measured  $ET_{eco}$  and birch transpiration per leaf area  $(T_{leaf})$ , which was upscaled to the birch canopy level  $(T_{birch})$ . Our main goals were: (1) to identify the drivers of  $ET_{eco}$  and  $T_{leaf}$ , to understand the environmental controls between the two scales (ecosystem vs branch) and at sites, which differed substantially in stand structure (denser in Abisko, sparser in Kevo); and (2) to investigate how variation in canopy structure affects growing season values of  $ET_{eco}$  relative to growing season precipitation and to quantify the contribution of  $T_{birch}$ to  $ET_{eco}$ . To further understand this evapotranspiration partitioning in subarctic deciduous woodlands, at Kevo we also upscaled evaporative fluxes from birch and understorey ( $ET_{upscaled}$ ) to explore how this variable compares to  $ET_{eco}$ .

# 142 2. Methodology

*2.1. Study sites* 

Two mountain birch (*Betula pubescens* ssp. *czerepanovii*) forest sites within the northern Fennoscandia sub-Artic vegetation belt were chosen for this study: Abisko (northern Sweden) and Kevo (northern Finland). Both sites were located near the mountain birch/tundra ecotone, where mountain birches are polycormic because of the harsh environmental conditions and the frequent defoliation by autumn and winter moths (*Epirrita autumnata* and *Operophtera brumata*). At both sites, we measured transpiration of mountain birch branches, ecosystem evapotranspiration and other

environmental drivers during the mountain birch leaf-on period, hereby abbreviated as 'growing season', of 2007 (Abisko, DOY 153-241) and of 2008 (Kevo, DOY 171-257).

In Abisko (Figure 1a), measurements were undertaken at a location (68.326°N, 18.833°E, 519 m.a.s.l) ca. 3.2 km south-east of the Abisko Research Station. At the study site, mean annual temperature is -0.9°C and mean annual precipitation is 335 mm (1980-2010, temperature corrected assuming a lapse rate of 0.55 °C per 100 m of elevation). The predominant substrate is coarse glacial till and soils are typically micro-podzols, with no permafrost present (Hartley, Hopkins, Sommerkorn, & Wookey, 2010). The landscape presents a relatively complex topography, which results in highly variable forest cover (Nyström, Holmgren, & Olsson, 2012) and stand structures (Table 1). Understorey vegetation is dominated by the dwarf shrubs *Empetrum nigrum* ssp hermaphroditum, Vaccinium myrtillus and Vaccinium uliginosum (Hartley et al., 2010; Poyatos, Gornall, Mencuccini, Huntley, & Baxter, 2012). 

In Kevo (Figure 1b), measurements were undertaken at a location (69.492°N, 27.234°E, 260 m.a.s.l.) ca. 40 km south of the Kevo Subartic Research Institute. Climate at the site (1978-2007, data from the Kevo Institute station, corrected for lapse rate) is colder and wetter than in Abisko (-2.4°C and 422 mm mean annual temperature and precipitation, respectively) and the substrate is composed of gneiss covered by glacial till, and no permafrost is present at the forest site. Mountain birch forests in Kevo, located upon gentle slopes/ridges and surrounded by mires in topographically depressed areas, were sparser and showed a more homogeneous structure compared to Abisko (Table 1). Understorey vegetation showed a higher LAI compared to Abisko (Table 1); it consisted of E. nigrum below mountain birch canopies and distinct patches covered by Betula nana L. and Cladonia spp, lichens in the open areas (Poyatos et al., 2012). 

One forest inventory was established in the vicinity of each of the branch bags sites to quantify stand structure at the plot level (a 10-m circular plot in Abisko and a 30 x 30 m plot in Kevo). Another set of 30 x 30 m plots was measured in Abisko (N = 5) and Kevo (N = 7) to quantify ecosystem-level stand structure and maximum leaf area index, LAI<sub>max</sub> ( $m^2$  leaf  $m^2$ ground). Forest inventory plots were at an average distance from the eddy flux tower of 105 m in Abisko and 450 m in Kevo. Diameters and heights of all

stems with diameter at breast height DBH>12 mm within the plots were measured in 2007 at Abisko and in 2008 at Kevo. For Abisko, we used published allometric equations predicting leaf biomass from stem basal area and height (Dahlberg, Berge, Petersson, & Vencatasawmy, 2004) to convert leaf biomass supported by each stem into leaf area using site-specific leaf mass per area. For Kevo, we harvested N = 15 stems during the peak growing season in 2008, to measure their leaf area and we obtained sitespecific allometries between stem diameter and leaf area (Table S1). Understorey  $LAI_{max}$  was obtained from 1 m<sup>2</sup> vegetation surveys (N = 5) in each of the sites, following Fletcher et al. (2012). 

# 191 2.2. Branch-level transpiration measurements

At both sites, we selected eight mountain birch branches representative of low and mid-canopy conditions for branch transpiration measurements. Branch transpiration was measured using a multiplexed branch bag device based on the closed system approach (Rayment & Jarvis, 1999; Wingate, Seibt, Moncrieff, Jarvis, & Lloyd, 2007). This system measures water vapour concentration changes within eight 0.11 m<sup>3</sup> ventilated cuvettes enclosing individual branches during 7.5 minutes. Branches were measured sequentially, and a measurement cycle of all eight branches was completed within an hour. During each measurement period, air temperature, T (°C), relative humidity, RH (%), and photosynthetically active radiation, PAR (µmols photons m<sup>-2</sup> s<sup>-1</sup>), were recorded every 5 seconds by a datalogger. The system also recorded the value of environmental variables at the beginning of each transpiration observation (i.e. hourly). The subscript 'branch' was used to refer to branch-level meteorological variables (PAR<sub>branch</sub>, VPD<sub>branch</sub>). Further technical details of the branch bags system and of the calculation of branch-level transpiration can be found in the Supporting Information S2. We quantified branch transpiration on a leaf area basis,  $T_{leaf}$  (1 m<sup>-2</sup> hour<sup>-1</sup>), by dividing whole-branch transpiration by the leaf area of the branch within the bag. To account for seasonal variation in branch leaf area, we periodically counted the number of leaves inside the bags during the growing season. We then multiplied the leaf counts by an estimation of the average leaf area obtained from a sample of leaves (N = 10) close to the measured branch, fitted a nonlinear response as a function of day of year and, if needed, corrected by differences in leaf size between inside and outside the bags

- 213 (Poyatos et al., 2012). We expressed the seasonal variation in leaf area in relative terms
- 214 between 0 and 1 (minimum and maximum leaf area, respectively) to use for the
- 215 upscaling of branch transpiration fluxes.
- 216 2.3. Ecosystem evapotranspiration and environmental monitoring
- 217 At both sites, half-hourly ecosystem evapotranspiration,  $ET_{eco}$  (mm h<sup>-1</sup>), was estimated
- 218 from latent heat measurements using the eddy covariance (EC) technique in flux towers
- 219 located above the mountain birch canopy (Aubinet, Vesala, & Papale, 2012). The three
- 220 components of wind speed were measured with a sonic anemometer (R3, Gill
- 221 Instruments, Lymington, UK) and water vapour concentrations were measured by an
- 222 open-path infrared gas analyser (LI-7500, LI-COR Biosciences, Lincoln, USA). Raw
- 223 data were logged at 20 Hz and processed to 30-minute statistics using FluxView (Centre
- 224 for Ecology and Hydrology, Wallingford, UK) and quality-controlled following
- 225 standard procedures. These include correcting sonic data for angle-of-attack (Gash &
- 226 Dolman, 2003), compensating for the lag time between sonic and gas analyser, rotating
- 227 the co-ordinate system (so that the horizontal wind vector is aligned with the 30-min
- 228 mean and the vertical component is forced to zero), correcting sonic temperature for
- 229 humidity (Schotanus, Nieuwstadt, & De Bruin, 1983), correcting the fluxes for high-
- 230 and low- frequency spectral losses and correcting gas fluxes for density effects (Webb,
- 231 Pearman, & Leuning, 1980). Quality control involved despiking and removal of data
- outside physically reasonable limits, when instruments malfunctioned, when the
- 233 windows of the gas analyser were wet or dirty, and during periods of heavy rain.
- 234 Filtering of data during low turbulence conditions based on a friction velocity threshold
- 235 was not applied. Energy balance closure was within the expected range (Stoy et al.,
- 236 2013) and did not differ much across sites (Supplementary Information S3).
- 237 Meteorological stations installed at the flux towers measured half-hourly values of
- 238 temperature, relative humidity, PAR and precipitation above the birch canopy and we
- 239 refer to them using the subscript 'eco' (PAR<sub>eco</sub>, VPD<sub>eco</sub>). Soil volumetric water content
- 240 in the upper 30 cm of the soil, SWC (cm<sup>3</sup> cm<sup>-3</sup>), was measured with 1 or 2 frequency
- 241 domain reflectometers (CS616, Campbell Scientific, UK) at each site. To account for
- 242 site-specific differences in maximum and minimum water-holding capacity, we

- 243 transformed SWC into soil moisture deficit (SMD), which ranged from 0 (maximum
- 244 soil moisture) to 1 (minimum soil moisture) (Granier & Loustau, 1994).
- 245 2.4. Modelling environmental controls of evaporative fluxes
- Firstly,  $T_{leaf}$  and  $ET_{eco}$  data were filtered (PAR > 50  $\mu$ mol photons m<sup>-2</sup> s<sup>-1</sup>) to avoid noisy
- 247 vapour concentration data in the branch bags and low turbulence conditions in the case
- of EC. For  $T_{leaf}$ , the values of the meteorological drivers were measured locally in each
- 249 individual branch (VPD<sub>branch</sub>, PAR<sub>branch</sub>) and for ET<sub>eco</sub> they were measured above the
- 250 canopy (VPD<sub>eco</sub>, PAR<sub>eco</sub>).
- 251 All models were fitted using the nlme package (Pinheiro, Bates, DebRoy, Sarkar, & R
- 252 Core Team, 2018) in R (R Core Team 2016). T<sub>leaf</sub> was modelled using a linear mixed
- effects model (lme), with VPD<sub>branch</sub>, PAR<sub>branch</sub> and SMD as fixed factors and  $ET_{eco}$  was
- 254 fitted as a function of VPD<sub>eco</sub>, PAR<sub>eco</sub> and SMD using a generalized least squares model
- 255 (gls). In view of the residual distributions after preliminary analyses, we log-
- transformed the response variables,  $T_{leaf}$  and  $ET_{eco}$ , and the explanatory variables, except
- 257 for the case of PAR in  $ET_{eco}$  modelling. All models included a first-order autoregressive
- 258 correlation structure for the residuals, specifying fractional day of year as a continuous
- 259 time covariate. We applied model selection to include those terms which minimised the
- 260 value of the Akaike Information Criterion (AIC) while checking that variance inflation
- 261 factors were below 10 (Zuur, Ieno, & Elphick, 2010). Model selection was carried out
- 262 with models fitted using maximum likelihood, but final models were fitted using
- 263 restricted maximum likelihood (Pinheiro & Bates, 2000). Normality, linearity and
- 264 homoscedasticity of residuals were visually inspected and temporal autocorrelation was
- analysed visually by autocorrelation plots using the acf function in R. We calculated
- 266 marginal and conditional  $R^2$ , the proportion of variance explained by fixed and by both
- 267 fixed and random factors, respectively (Nakagawa & Schielzeth, 2013).
- 268 2.5. Overstorey contributions to ecosystem evapotranspiration
- 269 Before upscaling, evaporative flux data were aggregated at the daily scale, using models
- 270 obtained in section 2.4 to gap-fill missing hourly data and fitting daily models when
- 271 meteorological data from the measurement systems were missing (Supporting
- 272 Information S4). We obtained transpiration of the mountain birch canopy,  $T_{birch}$  (mm
- 273 day<sup>-1</sup>), by multiplying  $T_{leaf}$  by the LAI of mountain birch in each stand (Table 1),

- 274 corrected for seasonal variation (see section 2.2). The calculation was done using mean
- 275 and ± standard error (SE) of the LAI values, to propagate the uncertainty of the LAI
- 276 values at each site into the upscaled estimates of  $T_{birch}$ .
- 277 At both sites we calculated the mountain birch contribution to daily ecosystem
- evapotranspiration,  $T_{birch}/ET_{eco}$  (%). We analysed  $T_{birch}/ET_{eco}$  as a separate linear model
- 279 of VPD, PAR (both log-transformed) and SMD, including a factor coding for site
- 280 (Abisko and Kevo) which interacted with each of the environmental drivers. Model
- 281 selection was carried out based on AIC, as described in section 2.4. We also tested for a
- 282 possible influence of interception and subsequent canopy evaporation on  $T_{birch}/ET_{eco}$  by
- 283 testing for differences between dry and wet days, using a gls model as described in the
- 284 previous paragraph. We considered wet days as those within 2 days after a precipitation
- 285 event > 1 mm, assuming all wet surfaces would have dried up during this period
- 286 (Knauer, Werner, & Zaehle, 2015).
- 287 Growing season values (mm) of precipitation (P),  $T_{birch}$ ,  $ET_{eco}$  and  $ET_{upscaled}$  were
- 288 calculated by aggregating daily values. We also quantified the overall growing season
- contribution of  $T_{birch}$  and  $ET_{upscaled}$  to  $ET_{eco}$  and expressed growing season evaporative
- 290 fluxes as a percentage of growing season precipitation.
- 291 2.6. Upscaling evapotranspiration components in Kevo
- 292 In Kevo, measurements of evapotranspiration were available for other ecosystem
- 293 components, i.e., understorey shrubs and lichen (Table S3, Figure 1). These
- 294 evapotranspiration measurements were representative of small patches and were
- obtained with an automated chamber system (Poyatos et al., 2014) operated during the
- 296 2008 growing season, in a forest-mire ecotone ca. 200 m from the flux tower (Figure 1).
- 297 Hourly evapotranspiration of 12 tundra plots was calculated similarly to branch bags
- 298 fluxes (Supporting Information S5). Because of microclimatic alterations, water vapour
- 299 sorption in the tubing system and imperfect chamber sealing the automated chamber
- 300 system used here has been reported to underestimates the evaporative fluxes (Cohen et
- al., 2015). Therefore, we applied a correction factor of 2.3, obtained in that study, which
- used a similar device under comparable environmental conditions (Cohen et al., 2015).
- 303 Shrub evapotranspiration ( $ET_{shrub}$ ) was estimated as the mean of N = 9 plots (mean
- $304 \text{ LAI}_{max} \pm \text{SE} = 0.77 \pm 0.2$ ) with dwarf tundra vegetation (mainly *Empetrum*

305 hermaphroditum, Calluna vulgaris and Vaccinium spp.) while lichen evaporation

306 ( $ET_{lichen}$ ) was calculated as the mean of N = 3 lichen heath plots (Poyatos et al., 2014).

307 We then combined evapotranspiration of the individual components with the fractional

308 covers (f) of each component within the footprint of the flux tower. Fractional covers

309 were obtained from aerial photography obtained in August 2008 and subsequent

vegetation classification (Hartley et al., 2015). We used a dynamic footprint approach

311 (Hartley et al., 2015) to obtain f values which varied with atmospheric conditions,

312 although results were comparable to those using a simpler, fixed footprint approach

313 (Figure S4). We calculated  $ET_{upscaled}$  (mm day-1) as the product of the time-variable f of

314 each component and its corresponding T or ET value:

315 
$$ET_{upscaled} = T_{birch} + f_{birch} \cdot ET_{shrub} + f_{shrub} \cdot ET_{shrub} + f_{lichen} \cdot ET_{lichen}$$
 (1)

- Where  $f_{birch}$ ,  $f_{shrub}$  and  $f_{lichen}$  represent the fractional covers of birch forest, understorey
- 317 shrubs and lichen, respectively. This equation assumes that shrubs were also typically
- present under the birch canopies (cf. section 2.1) and that components other than birch,
- 319 shrubs and lichen (around 5% of fractional cover, Table S3) behave similarly to shrubs.

# **3. Results**

- 322 3.1. Temporal variation of environmental variables and evaporative fluxes
- 323 Evaporative demand (Figure 2a-d) was higher in Abisko than in Kevo, as shown by
- 324 higher mean growing season values ( $\pm$  standard deviation, SD) of air temperatures (10.5
- 325 °C  $\pm$  3.8 and 9.5 °C  $\pm$  3.6, respectively), VPD<sub>eco</sub> (0.5  $\pm$  0.3 kPa and 0.3  $\pm$  0.2 kPa) and
- $PAR_{eco}$  (407.0  $\pm$  170.0  $\mu$ mol m<sup>-2</sup> s<sup>1</sup> and 260.4  $\pm$  130.3  $\mu$ mol m<sup>-2</sup> s<sup>1</sup>). Light transmission
- 327 through the birch canopy was higher in Kevo: PAR<sub>branch</sub>/PAR<sub>eco</sub> was 56% in Kevo
- 328 compared to 30% in Abisko (Figure 2a,b). This was associated with the larger
- 329 difference between VPD<sub>branch</sub> and VPD<sub>eco</sub> (Figure 2c,d) in Kevo (average VPD<sub>branch</sub> -
- $VPD_{eco} = 0.30 \text{ kPa}$ ) compared to Abisko (average  $VPD_{branch}$   $VPD_{eco} = 0.14 \text{ kPa}$ ). Kevo
- 331 also received heavier and more frequent precipitation (Figure 2e,f), resulting in higher
- total growing season precipitation (167.5 mm) compared to Abisko (126.6 mm).
- Both  $T_{leaf}$  and  $ET_{eco}$  tended to be higher in Abisko than in Kevo, on average 50% higher
- 334 for  $T_{leaf}$  and 62% higher for  $ET_{eco}$ . Their seasonal dynamics were similar and followed

- the course of evaporative demand (Figure 2g-j). However, some differences between
- $T_{leaf}$  and  $ET_{eco}$  during the early growing season (before DOY 160) were apparent for
- 337 Abisko. The diurnal cycles of evaporative fluxes and their drivers varied seasonally in
- 338 both sites (Figure S2, S3), as expected due to the changing daylight hours at these
- latitudes. Abisko typically presented higher  $ET_{eco}$  and  $T_{leaf}$  except during the late season,
- 340 when  $T_{leaf}$  was equal for the two sites.
- 341 3. 2. Modelling environmental controls of evaporative fluxes
- $ET_{eco}$  and  $T_{leaf}$  increased with PAR and VPD but the relationship with VPD showed
- 343 much less scatter (Figure 3). In general,  $T_{leaf}$  and  $ET_{eco}$  at a given value of PAR or VPD
- 344 were higher for Abisko. Models of  $ET_{eco}$  and  $T_{leaf}$  showed a good predictive ability, with
- marginal  $R^2$  values > 0.7 (Table 2,3). Model predictors included a negative interaction
- 346 between PAR and VPD but did not include SMD (Table 2,3). The environmental
- responses of  $ET_{eco}$  did not vary across sites and we only detected site differences for the
- 348 intercept and the PAR coefficient in the  $T_{leaf}$  model (Table 2,3). In both models, the
- 349 interaction between VPD and PAR resulted in complex patterns in the variation of  $T_{leaf}$
- and  $ET_{eco}$  (Figure 4). For example, for  $T_{leaf}$ , steeper relationships with VPD<sub>branch</sub> were
- predicted at low PAR<sub>branch</sub> in both sites. In Abisko, higher  $ET_{eco}$  was predicted under
- 352 conditions of high PAR<sub>eco</sub> and low VPD<sub>eco</sub> values (Figure 4).
- 353 3. 3. Overstorey and understorey contributions to ecosystem evapotranspiration
- 354 Higher spatial variability of LAI in Abisko (Table 1) translated into a much larger
- 355 variability in  $T_{birch}$ , while  $T_{birch}$  was lower and less variable in Kevo (Figure 5). On
- average, the daily contribution of mean  $T_{birch}$  to  $ET_{eco}$  reached peak values of ca. 65% in
- 357 Abisko and ca. 30% in Kevo. However, the highly variable LAI in Abisko (Table 1)
- 358 resulted in the upper bound of  $T_{birch}/ET_{eco}$  occasionally approaching 100% at this
- 359 location (Figure 5c).
- 360 The value of  $T_{birch}/ET_{eco}$  increased with VPD<sub>eco</sub> and PAR<sub>eco</sub> (both log-transformed;
- 361 Table S4, Figure 6a,b). In both cases, model selection retained the interaction between
- 362 site and the environmental variable, but it was not significant for either driver (Table
- 363 S4). We did not detect any effect of SMD on  $T_{birch}/ET_{eco}$  (Figure 6c; Table S4). We did
- 364 not find any difference in  $T_{birch}/ET_{eco}$  between dry and wet days (p = 0.27).

The mean growing season contribution of  $T_{birch}$  to  $ET_{eco}$  was relatively low in Abisko

366 (ca. 33%) but it was even lower in Kevo (16%, Table 4). Daily evapotranspiration by

367 understorey components in Kevo was generally lower compared to  $T_{birch}$  (Figure 5d).

368 For the whole of the growing season,  $ET_{upscaled}$  only amounted to ca. 40% of  $ET_{eco}$  in

369 Kevo (Table 4).

370 Daily  $ET_{eco}$  was higher in Abisko (Figure 5a,b), which also showed higher growing

371 season totals compared to Kevo (Table 4). Remarkably, in Abisko  $ET_{eco}$  was 27%

372 higher than the precipitation in the same period, while in Kevo the ecosystem returned

373 to the atmosphere only ca. 59% of precipitation ( $ET_{eco}/P$ , Table 4). Nevertheless, the

374 relative role of mountain birch transpiration in recycling precipitation was much higher

in Abisko than in Kevo ( $T_{birch}/P$ , Table 4).

#### 377 4. Discussion

378 4.1. Differences in seasonal and environmental controls on transpiration and

379 evapotranspiration between sites

Boreal and arctic regions are undergoing very rapid and pronounced climatic warming,

381 which is expected to modify water and energy fluxes across much of the terrestrial

382 biosphere of these northern regions. We find that controls of evaporative fluxes by

383 mixed birch-tundra communities of Northern Fennoscandia largely consist of controls

384 by VPD (which strongly depends on air and canopy temperature) and by PAR. The

385 relative importance of these effects depended partly on specific site conditions and the

386 scale (branch versus ecosystem) at which they were considered. Predicted increases in

air temperature can therefore be expected to increase the relative contribution of VPD

388 relative to PAR in controlling evaporative fluxes.

Conversely, we find that the evaporative fluxes are not affected by temporal changes in

390 soil moisture, suggesting that water supply is currently not a major limiting factor to

evapotranspiration. Thus, there were no edaphic drought stress effects in  $T_{leaf}$  regulation

392 by mountain birch, confirming results observed for other birch species (Gartner,

393 Nadezhdina, Englisch, Čermak, & Leitgeb, 2009; Yan et al., 2018). Our results at the

394 ecosystem level are consistent with field studies in forest-tundra systems (Beringer et

al., 2005) and with a recent data synthesis, where no effect of soil moisture was reported

396 for evapotranspiration at high latitudes (Kasurinen et al., 2014). Nevertheless,

397 evaporative fluxes in boreal forests in more continental climates, with higher

evaporative demands, may be influenced by soil moisture (Ohta et al., 2008).

399 At the seasonal time scale, fluxes were primarily controlled by LAI dynamics at both

400 sites (cf. Poyatos et al., 2012). Seasonal courses of  $T_{leaf}$  and  $ET_{eco}$  mirrored each other,

401 except during the start of the growing season in Abisko, when the discrepancy between

 $T_{leaf}$  and  $ET_{eco}$  may have been caused by combined errors in the quantification of low

403 fluxes and leaf area during early leaf development. Alternatively, this temporal

404 mismatch between  $T_{leaf}$  and  $ET_{eco}$  may have been driven by substantial evaporation from

405 moist soils after snowmelt and/or spatial variability in the phenology of greening up

406 between the measured branches and the rest of the forest.

Both  $T_{leaf}$  and  $ET_{eco}$  were higher in Abisko than in Kevo because of the generally higher

408 evaporative demand in Abisko (Figure 2). Environmental controls on  $T_{leaf}$  were very

similar across sites. The only significant difference in the response of  $T_{leaf}$  to PAR may

410 be due to differences in stand structure at the plot level (Table 1). The responses of

evaporative fluxes to PAR and VPD differed between the two sites more clearly for  $T_{leaf}$ 

than for  $ET_{eco}$ , suggesting a higher sensitivity to VPD of the birch canopy compared to

other ecosystem components (see also section 4.2). The negative interaction between

414 VPD and PAR produced complex response surfaces of evaporative fluxes to

415 environmental conditions. Model responses during conditions of high evaporative

416 demand were reasonable, apart from those by  $ET_{eco}$  at Abisko, where the model showed

a decrease of  $ET_{eco}$  with VPD at high PAR. The more extreme responses were found for

418 unrealistic combinations of environmental conditions, which are not usually found in

419 the field. (i.e. high VPD and low PAR), and when the model's predictions of the

420 interaction effects are less reliable.

422 4.2. Contribution of mountain birch transpiration to ecosystem evapotranspiration

423 across sites and environmental conditions

424 The mean daily contribution of birch transpiration to ecosystem evapotranspiration (i.e.

 $T_{birch}/ET_{eco}$ ) was much higher in Abisko than in Kevo. In Abisko, the higher variability

in LAI at the landscape level propagates to a larger range of  $T_{birch}/ET_{eco}$  values compared to Kevo. When explaining seasonal variability in  $T_{birch}/ET_{eco}$ , we found that  $T_{birch}/ET_{eco}$  strongly depended on VPD and PAR, with  $T_{birch}/ET_{eco}$  saturating at high VPD, but this environmental control on  $T_{birch}/ET_{eco}$  was stronger in Abisko. Therefore, our results show an increased relative role of mountain birch in controlling ecosystem evapotranspiration as evaporative demand increases, especially in denser forests, in contrast with studies on waterlogged peatlands where understorey contribution increases with VPD (Ikawa et al., 2015). In our sites, mountain birch roots possibly access soil moisture at greater depths (Hunziker, Sigurdsson, Halldorsson, Schwanghart, & Kuhn, 2014), supplying water to meet the increasing evaporative demand and causing the increase in  $T_{birch}/ET_{eco}$ . 

At the growing season level, birch transpiration contributed ca. 33% of total ecosystem evapotranspiration in Abisko but the contribution was only ca. 16% in Kevo (Table 4). These differences were attributable not only to a higher birch LAI in Abisko (Table 1), but also to the higher  $T_{leaf}$  values at this site (Figure 2). Lower  $T_{birch}/ET_{eco}$  values in Kevo could also result from a disproportionately higher contribution from the understorey in a sparser woodland (i.e. higher below-canopy incident radiation compared to Abisko). The values of of  $T_{birch}/ET_{eco}$  at the two sites are consistent with the generally low contribution of overstorey to total evapotranspiration in subarctic and northern boreal forests (Iida et al., 2009; Ikawa et al., 2015; Kelliher et al., 1997; Lafleur, 1992; Warren et al., 2019). However, in Kevo, our estimates of upscaled evapotranspiration from individual ecosystem components (i.e. mountain birch, understorey shrubs and lichen heath) yielded growing season values, which were still far from total ecosystem evapotranspiration measured by eddy covariance (Table 4, cf. section 4.3). In the following section, we discuss potential methodological artefacts and unmeasured processes that could explain this discrepancy. 

# 452 4.3. Methodological considerations

This study jointly analyses a multi-scale dataset of evaporative fluxes from subarctic forest communities. Comparing evaporative fluxes across scales is hindered by the numerous potential errors associated with measurement techniques and upscaling procedures. Transpiration measurements from closed chambers could have been

457 affected by radiation-driven overheating (Poyatos et al., 2012), by raising VPD<sub>branch</sub>

458 above VPD<sub>eco</sub> and causing an overestimation of  $T_{leaf}$ . However, the relatively low values

of  $T_{birch}$  and  $ET_{upscaled}$ , both based on closed chamber measurements, do not suggest that

the conclusions of this study could have been affected by this artefact.

461 The upscaling procedure also has a number of potential limitations that warrant 462 consideration. Due to the sparseness of the forest in Kevo (i.e. little shading effects on

463 understorey vegetation), we assumed that the magnitude and regulation of understorey

464 evapotranspiration was similar to that shown by patches with similar composition in the

465 forest-tundra transition (Poyatos et al., 2014). However, LAI of the patches measured

with automated chambers in the forest-mire transition (see section 2.5) was ca. 50% of

467 the LAI actually measured in survey plots located within the forest (Table 1). Rescaling

468 the understorey fluxes according to this understorey LAI, evapotranspiration from

469 understorey components at the ecosystem level would be amount to 23.6 mm, an

470 evaporative flux 55% larger than  $T_{birch}$ . Scaling-up evapotranspiration estimated from

471 canopy and understorey components, accounting for their land cover fractions and

472 applying the LAI correction outlined above to understorey measurements would

473 increase growing season  $ET_{upscaled}$  values to 44.4 mm, or ca. 45% of  $ET_{eco}$ .

# 474 4.4. Differences in growing season water balance across sites

475 Even accounting for this likely underestimation of  $ET_{shrub}$  and  $ET_{lichen}$ , there is still a

fraction of  $ET_{eco}$  that cannot be explained by upscaled gas exchange measurements from

477 individual ecosystem components. Taking into account that  $T_{birch}$  obtained from branch-

478 bag measurements excludes evaporation of intercepted water, we showed that

 $T_{birch}/ET_{eco}$  does not vary between dry and wet days. This may suggest that evaporation

480 of intercepted water may not be captured by eddy covariance measurements, otherwise

 $T_{birch}/ET_{eco}$  would have been lower on wet days than dry days. Potentially high

482 evaporation rates after precipitation may be partially missed from  $ET_{eco}$  and  $ET_{upscaled}$ ,

483 because data from open-path gas analysers are removed when the sensor windows are

484 wet and subsequent gap-filling would not account for the missed evaporation of

485 intercepted water (Oishi, Oren, & Stoy, 2008). Combined interception by overstorey

486 canopies and mosses in northern boreal forests may amount up to 40% of bulk

precipitation (Price, Dunham, Carleton, & Band, 1997), and we are not currently accounting for this substantial contribution.

We found stark differences between sites in the percentage of precipitation returned to the atmosphere as evapotranspiration; the mountain birch woodland in Abisko evaporated more water than it received during the growing season, as observed in other deciduous boreal forests (Blanken et al., 2001; Kelliher et al., 1997). In contrast, Kevo showed a substantial water surplus (Table 4). Our measurements did not include the snowmelt period, but these sites can reach snowpack depths of > 1 m (data for Kevo, 2009) and tree water uptake during this period, especially from deciduous species, can progressively deplete soil water sources (Young-Robertson et al., 2016). This decline in soil water content after snowmelt is very clear in the seasonal course of SMD measured in Abisko in 2008 and 2009 (outside our measurement period in Abisko, Fig S5). Therefore, these differences in the role of the mountain birch canopy between Abisko and Kevo, mediated by their different stand structure, can illustrate the potential changes in the hydrological regime that can result from the expansion and densification of subarctic deciduous woodlands. 

# 503 4.5. Concluding remarks

We have shown that the dominant mountain birch canopy plays only a partial role in driving ecosystem evapotranspiration in both subarctic sites, and this may be a general feature of low-LAI subarctic and northern boreal forests (Saugier, Granier, Pontailler, Dufrene, & Baldocchi, 1997). Our results also show that both increased woodland cover and increased woodland density under climate change conditions (Rundqvist et al., 2011) will result in larger controls of the water fluxes by the canopies of deciduous trees as opposed to the understorey vegetation. However, our upscaling exercise also shows that adequately accounting for understorey components (and transpiration vs evaporation processes; Stoy et al., 2019) may be necessary to constrain future hydrological changes in these areas. The highly variable and patchy nature of subarctic vegetation may require flux upscaling approaches considering spatial variation not only of land cover (Hartley et al., 2015), but also of LAI (Stoy et al., 2013).

In the longer term, shifts towards deciduous-dominated communities in subarctic regions and an increased land cover by forest as opposed to tundra are expected to

- 518 induce large hydro-climatic effects. These effects are expected to be mediated by higher
- 519 transpiration rates, inducing complex land-climate feedbacks (Bonfils et al., 2012;
- 520 Swann et al., 2010), which need to be considered together with carbon- and energy-
- 521 related feedbacks (Wit et al., 2014). Overall, combining several flux datasets and land
- 522 cover information we provide, for the poorly studied subarctic deciduous woodlands,
- 523 highly valuable results that will help to calibrate and validate evapotranspiration
- 524 processes in ecosystem models.

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References

- Aubinet, M., Vesala, T., & Papale, D. (2012). Eddy Covariance: A Practical Guide to Measurement and Data Analysis. Springer Science & Business Media.
- Beringer, J., Chapin, F. S., Thompson, C. C., & McGuire, A. D. (2005). Surface energy exchanges along a tundra-forest transition and feedbacks to climate. *Agricultural and Forest Meteorology*, 131(3–4), 143–161.
- Blanken, P. D., Black, T. A., Neumann, H. H., den Hartog, G., Yang, P. C., Nesic, Z., & Lee, X. (2001). The seasonal water and energy exchange above and within a boreal aspen forest. *Journal of Hydrology*, 245(1–4), 118–136. https://doi.org/10.1016/S0022-1694(01)00343-2
- Bonan, G. B. (2008). Forests and Climate Change: Forcings, Feedbacks, and the Climate Benefits of Forests. *Science*, *320*, 1444–1449. https://doi.org/10.1126/science.1155121
- Bonfils, C. J. W., Phillips, T. J., Lawrence, D. M., Cameron-Smith, P., Riley, W. J., & Subin, Z. M. (2012). On the influence of shrub height and expansion on northern high latitude climate. *Environmental Research Letters*, 7(1), 015503. https://doi.org/10.1088/1748-9326/7/1/015503

- Brümmer, C., Black, T. A., Jassal, R. S., Grant, N. J., Spittlehouse, D. L., Chen, B., ... Wofsy, S. C. (2011). How climate and vegetation type influence evapotranspiration and water use efficiency in Canadian forest, peatland and grassland ecosystems. *Agricultural and Forest Meteorology*. https://doi.org/10.1016/j.agrformet.2011.04.008
- Callaghan, T. V., Björn, L. O., Chapin Iii, F., Chernov, Y., Christensen, T. R., Huntley, B., ... Shaver, G. R. (2005). Arctic tundra and polar desert ecosystems. In *Arctic climate impact assessment* (Vol. 1, pp. 243–352).
- Callaghan, T. V., Jonasson, C., Thierfelder, T., Yang, Z., Hedenås, H., Johansson, M., ... Sloan, V. L. (2013). Ecosystem change and stability over multiple decades in the Swedish subarctic: Complex processes and multiple drivers. *Philosophical Transactions of the Royal Society B: Biological Sciences*, 368(1624), 20120488. https://doi.org/10.1098/rstb.2012.0488
- Chapin, F. S., Mcguire, A. D., Randerson, J., Pielke, R., Baldocchi, D., Hobbie, S. E., ... Running, S. W. (2000). Arctic and boreal ecosystems of western North America as components of the climate system. *Global Change Biology*, *6*(S1), 211–223. https://doi.org/10.1046/j.1365-2486.2000.06022.x
- Cohen, L. R., Raz-Yaseef, N., Curtis, J. B., Young, J. M., Rahn, T. A., Wilson, C. J., ... Newman, B. D. (2015). Measuring diurnal cycles of evapotranspiration in the Arctic with an automated chamber system. *Ecohydrology*, 8(4), 652–659. https://doi.org/10.1002/eco.1532
- Dahlberg, U., Berge, T. W., Petersson, H., & Vencatasawmy, C. P. (2004). Modelling biomass and leaf area index in a sub-arctic Scandinavian mountain area. Scandinavian Journal of Forest Research, 19, 60–71. https://doi.org/10.1080/02827580310019266
- Fletcher, B. J., Gornall, J. L., Poyatos, R., Press, M. C., Stoy, P. C., Huntley, B., ... Phoenix, G. K. (2012). Photosynthesis and productivity in heterogeneous arctic tundra: Consequences for ecosystem function of mixing vegetation types at stand edges. *Journal of Ecology*, 100(2), 441–451. https://doi.org/10.1111/j.1365-2745.2011.01913.x
- Gartner, K., Nadezhdina, N., Englisch, M., Čermak, J., & Leitgeb, E. (2009). Sap flow of birch and Norway spruce during the European heat and drought in summer 2003. *Forest Ecology and Management*, 258(5), 590–599. https://doi.org/10.1016/j.foreco.2009.04.028
- Gash, J. H. C., & Dolman, A. J. (2003). Sonic anemometer (co)sine response and flux measurement: I. The potential for (co)sine error to affect sonic anemometer-based flux measurements. *Agricultural and Forest Meteorology*, 119(3), 195–207. https://doi.org/10.1016/S0168-1923(03)00137-0
- Granier, A., & Loustau, D. (1994). Measuring and modelling the transpiration of a maritime pine canopy from sap-flow data. *Agricultural and Forest Meteorology*, 71, 61–81.
- Grelle, A., Lundberg, A., Lindroth, A., Morén, A.-S., & Cienciala, E. (1997). Evaporation components of a boreal forest: Variations during the growing season. *Journal of Hydrology*, *197*(1), 70–87. https://doi.org/10.1016/S0022-1694(96)03267-2
- Haapanala, S., Ekberg, A., Hakola, H., Tarvainen, V., Rinne, J., Hellén, H., & Arneth, A. (2009). Mountain birch–potentially large source of sesquiterpenes into high latitude atmosphere. *Biogeosciences*, 6, 2709–2718.

- Hartley, I. P., Hopkins, D. W., Sommerkorn, M., & Wookey, P. A. (2010). The response of organic matter mineralisation to nutrient and substrate additions in sub-arctic soils. *Soil Biology and Biochemistry*, 42(1), 92–100.
- Hartley, Iain. P., Hill, Timothy. C., Wade, Thomas. J., Clement, Robert. J., Moncrieff, John. B., Prieto-Blanco, Ana., ... Baxter, Robert. (2015). Quantifying landscape-level methane fluxes in subarctic Finland using a multiscale approach. *Global Change Biology*, 21(10), 3712–3725. https://doi.org/10.1111/gcb.12975
- Hofgaard, A., Tømmervik, H., Rees, G., & Hanssen, F. (2013). Latitudinal forest advance in northernmost Norway since the early 20th century. *Journal of Biogeography*, 40(5), 938–949. https://doi.org/10.1111/jbi.12053
- Hunziker, M., Sigurdsson, B. D., Halldorsson, G., Schwanghart, W., & Kuhn, N. (2014). Biomass allometries and coarse root biomass distribution of mountain birch in southern Iceland. *Icelandic Agricultural Sciences*, 27, 111–125.
- Iida, S., Ohta, T., Matsumoto, K., Nakai, T., Kuwada, T., Kononov, A. V., ... Yabuki, H. (2009). Evapotranspiration from understory vegetation in an eastern Siberian boreal larch forest. *Agricultural and Forest Meteorology*, 149(6), 1129–1139. https://doi.org/10.1016/j.agrformet.2009.02.003
- Ikawa, H., Nakai, T., Busey, R. C., Kim, Y., Kobayashi, H., Nagai, S., ... Hinzman, L. (2015). Understory CO2, sensible heat, and latent heat fluxes in a black spruce forest in interior Alaska. *Agricultural and Forest Meteorology*, 214–215, 80–90. https://doi.org/10.1016/j.agrformet.2015.08.247
- IPCC. (2013). Climate Change 2013: The Physical Science Basis. Contribution of Working Group I to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change. https://doi.org/10.1017/CBO9781107415324
- Kasurinen, V., Alfredsen, K., Kolari, P., Mammarella, I., Alekseychik, P., Rinne, J., ... Berninger, F. (2014). Latent heat exchange in the boreal and arctic biomes. *Global Change Biology*, 20(11), 3439–3456. https://doi.org/10.1111/gcb.12640
- Kattsov, V. M., Källén, E., Cattle, H. P., Christensen, J., Drange, H., Hanssen-Bauer, I., ... others. (2005). Future climate change: Modeling and scenarios for the Arctic.
- Kelliher, F. M., Hollinger, D. Y., Schulze, E.-D., Vygodskaya, N. N., Byers, J. N., Hunt, J. E., ... Bauer, G. (1997). Evaporation from an eastern Siberian larch forest. *Agricultural and Forest Meteorology*, 85(3), 135–147. https://doi.org/10.1016/S0168-1923(96)02424-0
- Knauer, J., Werner, C., & Zaehle, S. (2015). Evaluating stomatal models and their atmospheric drought response in a land surface scheme: A multibiome analysis. *Journal of Geophysical Research: Biogeosciences*, *120*(10), 2015JG003114. https://doi.org/10.1002/2015JG003114
- Krankina, O. N., Pflugmacher, D., Hayes, D. J., McGuire, A. D., Hansen, M. C., Häme, T., ... Nelson, P. (2010). Vegetation cover in the eurasian arctic: Distribution, monitoring, and role in carbon cycling. In *Eurasian arctic land cover and land use in a changing climate* (pp. 79–108). Springer.
- Lafleur, P. M. (1992). Energy balance and evapotranspiration from a subarctic forest. *Agricultural and Forest Meteorology*, 58(3–4), 163–175. https://doi.org/10.1016/0168-1923(92)90059-D
- McFadden, J. P., Eugster, W., & Chapin III, F. S. (2003). A regional study of the controls on water vapor and CO2 exchange in arctic tundra. *Ecology*, 84(10), 2762–2776.

- Mekonnen, Z. A., Riley, W. J., Randerson, J. T., Grant, R. F., & Rogers, B. M. (2019). Expansion of high-latitude deciduous forests driven by interactions between climate warming and fire. *Nature Plants*, 1–7. https://doi.org/10.1038/s41477-019-0495-8
- Myers-Smith, I. H., Forbes, B. C., Wilmking, M., Hallinger, M., Lantz, T., Blok, D., ... Hik, D. S. (2011). Shrub expansion in tundra ecosystems: Dynamics, impacts and research priorities. *Environmental Research Letters*, *6*(4), 045509. https://doi.org/10.1088/1748-9326/6/4/045509
- Nakagawa, S., & Schielzeth, H. (2013). A general and simple method for obtaining R2 from generalized linear mixed-effects models. *Methods in Ecology and Evolution*, 4(2), 133–142. https://doi.org/10.1111/j.2041-210x.2012.00261.x
- Nyström, M., Holmgren, J., & Olsson, H. (2012). Prediction of tree biomass in the forest–tundra ecotone using airborne laser scanning. *Remote Sensing of Environment*, 123, 271–279. https://doi.org/10.1016/j.rse.2012.03.008
- Ohta, T., Maximov, T. C., Dolman, A. J., Nakai, T., van der Molen, M. K., Kononov, A. V., ... Yabuki, H. (2008). Interannual variation of water balance and summer evapotranspiration in an eastern Siberian larch forest over a 7-year period (1998–2006). *Agricultural and Forest Meteorology*, 148(12), 1941–1953. https://doi.org/10.1016/j.agrformet.2008.04.012
- Oishi, A. C., Oren, R., & Stoy, P. C. (2008). Estimating components of forest evapotranspiration: A footprint approach for scaling sap flux measurements. *Agricultural and Forest Meteorology*, *148*(11), 1719–1732.
- Pinheiro, J., Bates, D., DebRoy, S., Sarkar, D., & R Core Team. (2018). *nlme: Linear and Nonlinear Mixed Effects Models*. Retrieved from https://CRAN.R-project.org/package=nlme
- Poyatos, R., Gornall, J., Mencuccini, M., Huntley, B., & Baxter, R. (2012). Seasonal controls on net branch CO2 assimilation in sub-Arctic Mountain Birch (Betula pubescens ssp. Czerepanovii (Orlova) Hamet-Ahti). *Agricultural and Forest Meteorology*, *158–159*, 90–100. https://doi.org/10.1016/j.agrformet.2012.02.009
- Poyatos, R., Heinemeyer, A., Ineson, P., Evans, J. G., Ward, H. C., Huntley, B., & Baxter, R. (2014). Environmental and Vegetation Drivers of Seasonal CO2 Fluxes in a Sub-arctic Forest–Mire Ecotone. *Ecosystems*, 17(3), 377–393. https://doi.org/10.1007/s10021-013-9728-2
- Price, A. G., Dunham, K., Carleton, T., & Band, L. (1997). Variability of water fluxes through the black spruce (Picea mariana) canopy and feather moss (Pleurozium schreberi) carpet in the boreal forest of Northern Manitoba. *Journal of Hydrology*, 196(1), 310–323. https://doi.org/10.1016/S0022-1694(96)03233-7
- Rayment, M. B., & Jarvis, P. G. (1999). Seasonal gas exchange of black spruce using an automatic branch bag system. *Canadian Journal of Forest Research*, 29, 1528–1538.
- Rundqvist, S., Hedenås, H., Sandström, A., Emanuelsson, U., Eriksson, H., Jonasson, C., & Callaghan, T. V. (2011). Tree and Shrub Expansion Over the Past 34 Years at the Tree-Line Near Abisko, Sweden. *AMBIO: A Journal of the Human Environment*, 40(6), 683–692. https://doi.org/10.1007/s13280-011-0174-0
- Saugier, B., Granier, A., Pontailler, J. Y., Dufrene, E., & Baldocchi, D. D. (1997). Transpiration of a boreal pine forest measured by branch bag, sap flow and micrometeorological methods. *Tree Physiology*, 17, 511–519.

- Schlesinger, W. H., & Jasechko, S. (2014). Transpiration in the global water cycle. *Agricultural and Forest Meteorology*, 189–190, 115–117. https://doi.org/10.1016/j.agrformet.2014.01.011
- Schotanus, P., Nieuwstadt, F. T. M., & De Bruin, H. A. R. (1983). Temperature measurement with a sonic anemometer and its application to heat and moisture fluxes. *Boundary-Layer Meteorology*, 26(1), 81–93. https://doi.org/10.1007/BF00164332
- Stoy, P. C., Williams, M., Evans, J. G., Prieto-Blanco, A., Disney, M., Hill, T. C., ... Street, L. E. (2013). Upscaling tundra CO<sub>2</sub> exchange from chamber to eddy covariance tower. *Arctic, Antarctic, and Alpine Research*, 45(2), 275–284.
- Stoy, Paul C., El-Madany, T. S., Fisher, J. B., Gentine, P., Gerken, T., Good, S. P., ... Wolf, S. (2019). Reviews and syntheses: Turning the challenges of partitioning ecosystem evaporation and transpiration into opportunities. *Biogeosciences*, *16*(19), 3747–3775. https://doi.org/10.5194/bg-16-3747-2019
- Stoy, Paul C., Mauder, M., Foken, T., Marcolla, B., Boegh, E., Ibrom, A., ... Varlagin, A. (2013). A data-driven analysis of energy balance closure across FLUXNET research sites: The role of landscape scale heterogeneity. *Agricultural and Forest Meteorology*, 171–172, 137–152. https://doi.org/10.1016/j.agrformet.2012.11.004
- Swann, A. L., Fung, I. Y., Levis, S., Bonan, G. B., & Doney, S. C. (2010). Changes in Arctic vegetation amplify high-latitude warming through the greenhouse effect. *Proceedings of the National Academy of Sciences*, 107(4), 1295–1300. https://doi.org/10.1073/pnas.0913846107
- Tømmervik, H., Johansen, B., Tombre, I., Thannheiser, D., Høgda, K. A., Gaare, E., & Wielgolaski, F. E. (2004). Vegetation Changes in the Nordic Mountain Birch Forest: The Influence of Grazing and Climate Change. *Arctic, Antarctic, and Alpine Research*, 36(3), 323–332.
- Wang, J. A., Sulla-Menashe, D., Woodcock, C. E., Sonnentag, O., Keeling, R. F., & Friedl, M. A. (2019). Extensive land cover change across Arctic–Boreal Northwestern North America from disturbance and climate forcing. *Global Change Biology*, *0*(0). https://doi.org/10.1111/gcb.14804
- Warren, R. K., Pappas, C., Helbig, M., Chasmer, L. E., Berg, A. A., Baltzer, J. L., ... Sonnentag, O. (2019). Minor contribution of overstorey transpiration to landscape evapotranspiration in boreal permafrost peatlands. *Ecohydrology*, 11(5), e1975. https://doi.org/10.1002/eco.1975
- Webb, E. K., Pearman, G. I., & Leuning, R. (1980). Correction of flux measurements for density effects due to heat and water vapour transfer. *Quarterly Journal of the Royal Meteorological Society*, 106(447), 85–100. https://doi.org/10.1002/qj.49710644707
- Welp, L. R., Randerson, J. T., & Liu, H. P. (2007). The sensitivity of carbon fluxes to spring warming and summer drought depends on plant functional type in boreal forest ecosystems. *Agricultural & Forest Meteorology*, 147(3–4), 172–185.
- Wingate, L., Seibt, U., Moncrieff, J. B., Jarvis, P. G., & Lloyd, J. (2007). Variations in 13C discrimination during CO2 exchange by Picea sitchensis branches in the field. *Plant, Cell & Environment, 30*(5), 600–616. https://doi.org/10.1111/j.1365-3040.2007.01647.x
- Wit, H. A. de, Bryn, A., Hofgaard, A., Karstensen, J., Kvalevåg, M. M., & Peters, G. P. (2014). Climate warming feedback from mountain birch forest expansion:

- Reduced albedo dominates carbon uptake. *Global Change Biology*, 20(7), 2344–2355. https://doi.org/10.1111/gcb.12483
- Yan, C., Wang, B., Zhang, Y., Zhang, X., Takeuchi, S., & Qiu, G. (2018). Responses of Sap Flow of Deciduous and Conifer Trees to Soil Drying in a Subalpine Forest. *Forests*, *9*(1), 32. https://doi.org/10.3390/f9010032
- Young-Robertson, J. M., Bolton, W. R., Bhatt, U. S., Cristóbal, J., & Thoman, R. (2016). Deciduous trees are a large and overlooked sink for snowmelt water in the boreal forest. *Scientific Reports*, 6, srep29504. https://doi.org/10.1038/srep29504
- Zhang, W., Miller, P. A., Smith, B., Wania, R., Koenigk, T., & Döscher, R. (2013). Tundra shrubification and tree-line advance amplify arctic climate warming: Results from an individual-based dynamic vegetation model. *Environmental Research Letters*, 8(3), 034023. https://doi.org/10.1088/1748-9326/8/3/034023
- Zuur, A. F., Ieno, E. N., & Elphick, C. S. (2010). A protocol for data exploration to avoid common statistical problems. *Methods in Ecology and Evolution*, *I*(1), 3–14. https://doi.org/10.1111/j.2041-210X.2009.00001.x

## 537 Tables

Table 1. Stand characteristics of mountain birch forests in Abisko and Kevo. Values labelled as 'Site' represent the site mean (±SE) of all inventory plots in Abisko (N=6) and Kevo (N=8). Values labelled as 'BB' are the values of the plots in the vicinity of the branch bags measuring sites. Tree density refers to polycormic individuals, with multiple stems per tree.

	Tree density (trees ha <sup>-1</sup> )	Stems per tree	Basal area (m²ha-1)	DBH (mm)	Height (m)	Overstorey LAI <sub>max</sub> (m <sup>2</sup> m <sup>-2</sup> )	Understorey LAI <sub>max</sub> (m <sup>2</sup> m <sup>-2</sup> )
Abisko							
Site	$1260 \pm 80$	$3.7 \pm 0.2$	$6.5 \pm 0.2$	$36.9 \pm 1.0$	$3.9 \pm 0.2$	$1.2 \pm 0.3$	$1.0 \pm 0.2$
BB	1146	4.4	6.7	35.7	-	1.8	-
Kevo							
Site	$876 \pm 85$	$3.3 \pm 0.2$	$3.8 \pm 0.4$	$37.2 \pm 2.0$	$3.8 \pm 0.0$	$0.7 \pm 0.1$	$1.5 \pm 0.1$
ВВ	833	3.8	3.0	30.6	3.8	0.6	-
54.	3				7.04		

Table 2. Summary statistics of the linear mixed model of log-transformed  $T_{leaf}$  as a function of environmental variables (VPD<sub>branch</sub>, PAR<sub>branch</sub> and SMD) for Abisko and Kevo. Asterisks denote significant differences from zero (\*p<0.05, \*\*p<0.01, \*\*\*\*p<0.001). Statistical differences in model coefficients (p<0.05) between Abisko and Kevo were marked in bold. SD: Standard deviation. Interactions between variables are denoted by colon (:) and variables not included after model selection are denoted by 'n.i'.

	Abisko	Kevo
Fixed effects		
Intercept	$-2.98 \pm 0.09^*$	$-4.00 \pm 0.07^{***}$
$log(VPD_{branch})$	$1.26 \pm 0.01^{***}$	$1.27 \pm 0.01^{***}$
PAR <sub>branch</sub>	$4.1 \cdot 10^{-4} \pm 0.4 \cdot 10^{-4***}$	$7.5 \cdot 10^{-4} \pm 0.4 \cdot 10^{-4***}$
log(VPD <sub>branch</sub> ): PAR <sub>branch</sub>	$-8.4 \cdot 10^{-4} \pm 0.5 \cdot 10^{-4***}$	$-9.4 \cdot 10^{-4} \pm 0.4 \cdot 10^{-4***}$
SMD	n.i.	n.i.
Random effects (branch)		
SD (Intercept)	0.26	0.20
Residual error	0.40	0.48
Correlation structure ( $\phi$ )	4.40·10 <sup>-8</sup>	7.23 · 10-7
R <sup>2</sup> marginal (R <sup>2</sup> conditional)	0.78 (0.84)	0.77 (0.80)

Table 3. Summary statistics of the generalised least squares model of  $ET_{eco}$  as a function of environmental variables (VPD<sub>eco</sub>, PAR<sub>eco</sub> and SMD) for Abisko and Kevo. Asterisks denote significant differences from zero (\*p<0.05, \*\*p<0.01, \*\*\*p<0.001). No significant differences (p < 0.05) were found between model coefficients between Abisko and Kevo. Interactions between variables are denoted by colon (:) and variables not included after model selection are denoted by 'n.i'.

	Abisko	Kevo
Intercept	-4.93 ± 0.38***	-4.84 ± 0.40***
log(VPD)	$2.58 \pm 0.30^{***}$	$2.13 \pm 0.27^{***}$
log(PAR <sub>eco</sub> )	$0.47 \pm 0.06^{***}$	$0.47 \pm 0.06^{***}$
log(VPD <sub>eco</sub> ):log(PAR <sub>eco</sub> )	-0.39 ±0.05***	$-0.26 \pm 0.05^{***}$
SMD	n.i.	n.i.
Correlation structure (φ )	$8.23 \cdot 10^{-3}$	1.84·10 <sup>-2</sup>
R <sup>2</sup> marginal	0.71	0.69
9		

Table 4. Growing season values of precipitation (P), birch transpiration ( $T_{birch}$ ) and ecosystem evapotranspiration ( $ET_{eco}$ ) in Abisko and Kevo. Percentage of evaporative fluxes as a fraction of  $ET_{eco}$  and P are also shown for growing season values. Values with an uncertainty measure represent means  $\pm$  standard error.

$T_{birch}$ (mm) $ET_{eco}$ (mm )	$52.5 \pm 13.0$	$15.2 \pm 1.5$
$ET_{eco}(mm)$		
/	160.5	98.5
$ET_{upscaled}(mm)$	-	$39.4 \pm 1.5$
$T_{birch}$ / $ET_{eco}$ (%)	$32.7 \pm 8.1$	$15.5 \pm 1.5$
$T_{birch}/P(\%)$	$41.4 \pm 10.2$	$9.1 \pm 0.9$
$ET_{eco}/P$ (% )	126.6	58.8
$ET_{upscaled} / ET_{eco} (\%)$	<del>-</del>	$40.0 \pm 1.5$

# 567 Figure captions

- **Figure 1.** Study sites at Abisko (a) and Kevo (b), showing the locations of the branch
- 569 bags systems, the eddy flux towers and the understorey automated chambers at Kevo.
- 570 Panel (a) shows the aerial photography obtained in Abisko and (b) shows the land
- 571 classification at Kevo obtained from aerial photography (cf. Hartley et al., 2015). Birch
- 572 : mountain birch woodland; *Understorey*: low- and dwarf-shrubs; *Lichen*: lichen heath;
- 573 Mire: organic hummocks and interhummocks with shrubs and Spahgnum; Water: open
- water; Lawns: graminoid lawns; Board: boardwalks; Other: other land cover.
- 575 Figure 2. Seasonal course of environmental variables and evaporative fluxes (daily
- 576 means) in Abisko and Kevo. Environmental variables include photosynthetically active
- 577 radiation (a-b, PAR), vapour pressure deficit (c-d, VPD) and rainfall (e, f).
- 578 Environmental variables were measured at the ecosystem (black lines) and at the branch
- 579 level (red lines). Mountain birch transpiration per unit leaf area (g-h,  $T_{eaf}$ ) and
- 580 ecosystem evapotranspiration (i-j,  $ET_{eco}$ ) are also shown. Standard error is shown as
- 581 shaded grey.
- Figure 3. Sub-daily responses of ecosystem evapotranspiration ( $ET_{eco}$ ) and mountain
- birch transpiration per unit leaf area ( $T_{eaf}$ ) to PAR (panels a,c) and VPD (panels b,d),
- measured at the corresponding ecological scale (i.e. 'branch' for  $T_{leaf}$  and 'eco' for
- $ET_{eco}$ ) in Abisko (red) and Kevo (blue).
- Figure 4. Response surfaces of modelled  $T_{leaf}$  (panels a, b) and  $ET_{eco}$  (panels c, d) as a
- 587 function of VPD and PAR, in Abisko (panels a, c) and Kevo (panels b, d). Please note
- the different scales in the VPD axes in panels a and b compared to panels c and d.
- Figure 5. Seasonal course of daily ecosystem evapotranspiration ( $ET_{eco}$ , black lines)
- and upscaled birch transpiration ( $T_{birch}$ , grey lines), for Abisko (a) and Kevo (b). The
- shaded regions in panels a and b depict upscaled  $T_{birch}$  using mean±SE values of LAI
- 592 (Table 1). Daily percentage of  $T_{birch}/ET_{eco}$  for Abisko (c) and Kevo (d). Panel (f) shows
- 593 evapotranspiration components and their upscaled values for Kevo only:  $ET_{eco}$  (black
- 594 line),  $T_{birch}$  (grey line),  $ET_{shrub}$  (purple line),  $ET_{lichen}$  (green line),  $ET_{upscaled}$  (asterisk).
- **Figure 6.** Variation of daily  $T_{birch} / ET_{eco}$  in response to VPD<sub>eco</sub> (a), PAR<sub>eco</sub> (b) and
- 596 SMD (c), for Abisko (red) and Kevo (blue). Models summary are shown in Table S3.

- 597 Significant interaction between site and environmental value is shown in solid line and
- 598 no-significant interaction in dashed line.



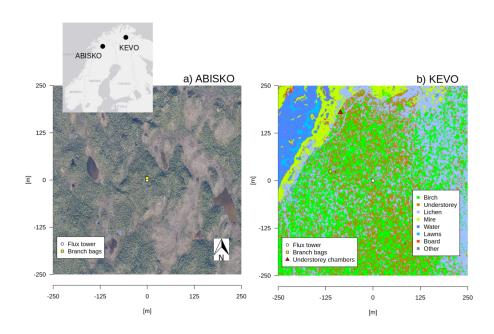


Figure 1. Study sites at Abisko (a) and Kevo (b), showing the locations of the branch bags systems, the eddy flux towers and the understorey automated chambers at Kevo. Panel (a) shows the aerial photography obtained in Abisko and (b) shows the land classification at Kevo obtained from aerial photography (cf. Hartley et al., 2015). Birch: mountain birch woodland; Understorey: low- and dwarf-shrubs; Lichen: lichen heath; Mire: organic hummocks and interhummocks with shrubs and Spahgnum; Water: open water; Lawns: graminoid lawns; Board: boardwalks; Other: other land cover.

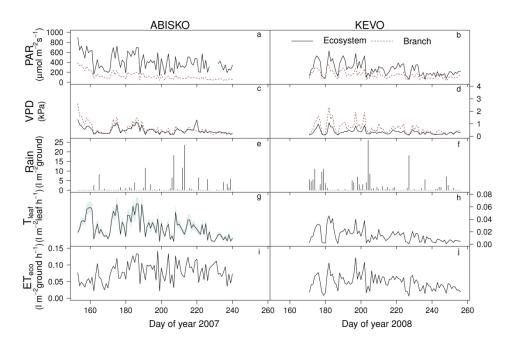


Figure 2. Seasonal course of environmental variables and evaporative fluxes (daily means) in Abisko and Kevo. Environmental variables include photosynthetically active radiation (a-b, PAR), vapour pressure deficit (c-d, VPD) and rainfall (e, f). Environmental variables were measured at the ecosystem (black lines) and at the branch level (red lines). Mountain birch transpiration per unit leaf area (g-h, Teaf) and ecosystem evapotranspiration (i-j, ETeco) are also shown. Standard error is shown as shaded grey.

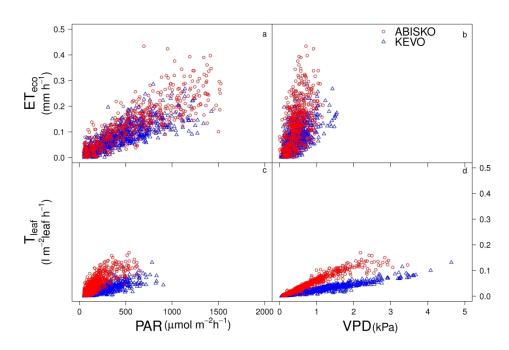


Figure 3. Sub-daily responses of ecosystem evapotranspiration (ETeco) and mountain birch transpiration per unit leaf area (Teaf) to PAR (panels a,c) and VPD (panels b,d), measured at the corresponding ecological scale (i.e. 'branch' for Tleaf and 'eco' for ETeco) in Abisko (red) and Kevo (blue).

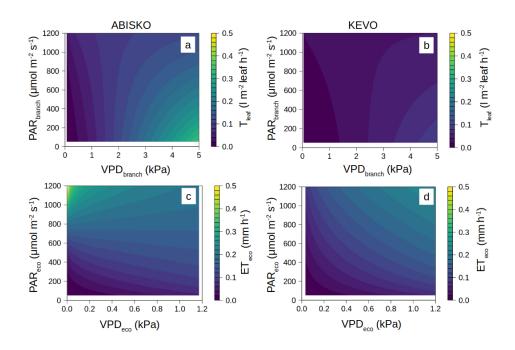


Figure 4. Response surfaces of modelled Tleaf (panels a, b) and ETeco (panels c, d) as a function of VPD and PAR, in Abisko (panels a, c) and Kevo (panels b, d). Please note the different scales in the VPD axes in panels a and b compared to panels c and d.

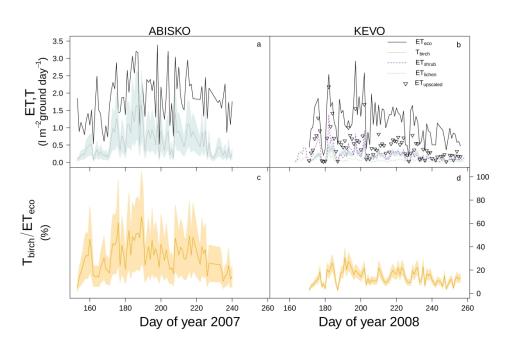


Figure 5. Seasonal course of daily ecosystem evapotranspiration (ETeco, black lines) and upscaled birch transpiration (Tbirch, grey lines), for Abisko (a) and Kevo (b). The shaded regions in panels a and b depict upscaled Tbirch using mean±SE values of LAI (Table 1). Daily percentage of Tbirch/ETeco for Abisko (c) and Kevo (d). Panel (f) shows evapotranspiration components and their upscaled values for Kevo only: ETeco (black line), Tbirch (grey line), ETshrub (purple line), ETlichen (green line), ETupscaled (asterisk).

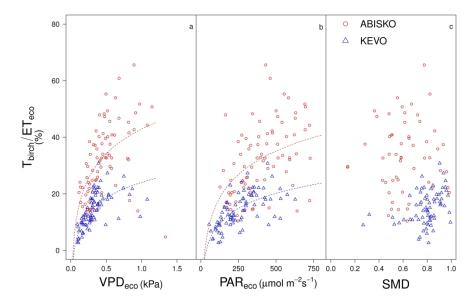


Figure 6. Variation of daily Tbirch / ETeco in response to VPDeco (a), PAReco (b) and SMD (c), for Abisko (red) and Kevo (blue). Models summary are shown in Table S3. Significant interaction between site and environmental value is shown in solid line and no-significant interaction in dashed line.

## 1 Supporting Information

### 2 S1. Allometric relationships

**Table S1**. Summary statistics for the allometric relationships used to predict leaf area supported by individual stems in Abisko and Kevo. Intercepts are labelled as a and coefficients associated to the different predictors are labelled as  $b_i$ .

Site	Response a	SE	Predictor	b <sub>i</sub>	SE	R <sup>2</sup> adj
	[units]		[units]			
Abisko	ln(leaf biomass) -8.11	0.71	ln (stem basal area)	0.43	0.11	0.86
	[kg]		$[mm^2]$			
			In (stem height)	1.06	0.37	
			[dm]			
Kevo	ln (leaf biomass) -4.95	0.25	ln (stem diameter)	1.56	0.07	0.97
	[g]		[mm]			

- 7 In Kevo, trees frequently presented a number of small stems (DBH<12mm), whose
- 8 number was recorded in the forest inventories; there were ca. 876 small stems ha<sup>-1</sup>. To
- 9 account for the leaf area supported by these stems, we assumed a typical diameter of 6
- 10 mm (half the value of the DBH threshold in the inventory) for these stems and applied
- 11 the allometric relationship above. Leaf area of small stems amounted, on average, 5% of
- 12 total plot leaf area.

## 13 S2. Technical description of the branch bags system

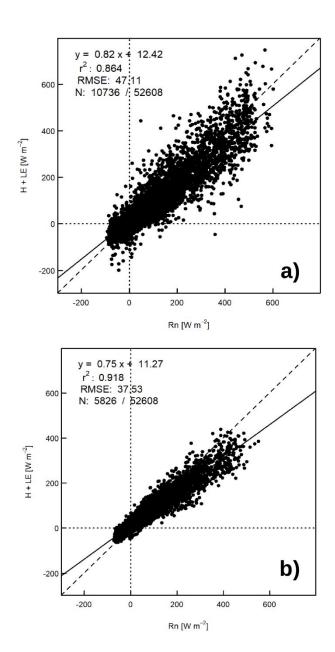
- 14 Branch bags (average volume=0.11 m<sup>3</sup>) were hung from tripods and poles to enclose the
- 15 sampled branches. A data-logger (CR10X, Campbell Scientific UK, Shepshed,
- Leicester, UK) and a control interface (SDM-CD16AC, Campbell Scientific UK)
- 17 controlled which bag was being sampled by delivering power to the electromagnetic
- 18 catch that sealed the bag, the internal mixing fan and the corresponding diaphragm
- 19 pump. This pump sampled air from the bag (5 dm<sup>3</sup> min<sup>-1</sup>) through polyethylene-lined
- 20 tubing, 5 mm in diameter, to an enclosure where a solenoid and another pump diverted
- 21 the air sample (0.2 dm<sup>3</sup> min<sup>-1</sup>) to an IRGA (LI-6262, LICOR Inc., Lincoln NE, USA).

- All the sampled branches were within ca. 10 m from this control box. Outputs from all
- 23 sensors and IRGA were transmitted through a relay multiplexer (AM416, Campbell
- 24 Scientific UK) and stored in the data-logger.
- 25 Each bag remained closed for 7.5 minutes (completing a measurement cycle of the eight
- 26 branches in one hour), during which H<sub>2</sub>O concentrations were measured every 5 s. with
- 27 the IRGA operating in absolute mode. Air temperature, relative humidity (HMP45C,
- 28 Vaisala, Vantaa, Finland), and air pressure (LI-6262-03, LI-COR Inc.) were also
- 29 recorded every 5 s. Average Photosynthetically Active Radiation (PAR) was also
- 30 measured inside the branch bags (SD101QV/SD201QV, Macam Ltd., Livingston, UK).
- 31 We calculated branch transpiration within the branch bag,  $T_{branch}$  (mmol s<sup>-1</sup>), as:

32 
$$E_{branch} = \frac{P_{bag} V_{bag} d}{RT_{bag,0} dt} [H_2 O]_{bag}$$
 (Eq. S1)

- Where  $P_{bag}$  (Pa) is air pressure,  $V_{bag}$  (m<sup>3</sup>) is bag volume, R (J mol<sup>-1</sup> K<sup>-1</sup>) is the ideal gas
- 34 constant,  $T_{bag,0}$  (K) is absolute temperature inside the bag and d[H<sub>2</sub>O]<sub>bag</sub>/dt (mmol s<sup>-1</sup>)is
- 35 the rate of change in H<sub>2</sub>O vapour concentration (mmol)during each observation. The
- 36 d[H<sub>2</sub>O]<sub>bag</sub>/dt (mmol s<sup>-1</sup>) was estimated from the first order term of a quadratic fit
- 37 between H<sub>2</sub>O vapour concentration (mmol) and time since chamber closure (Poyatos,
- 38 Gornall, Mencuccini, Huntley, & Baxter, 2012). Nonlinear fits describe better the
- 39 concentration dynamics, and do not systematically underestimate the fluxes (Wagner,
- 40 Reicosky, & Alessi, 1997). However, in the presence of noisy concentration data under
- 41 low flux conditions, we opted for the more stable linear fit. We achieved this by
- 42 selecting the linear regression whenever the slope for the linear fit and the linear term of
- 43 the quadratic fit had opposite signs. Flux calculations were implemented in a R script (v
- 44 2.9, R Development Core Team, Vienna, Austria), which also included tests for
- 45 autocorrelation (Durbin-Watson) and normality (Shapiro-Wilk) of residuals for each
- 46 flux observation (Kutzbach et al., 2007), and produced diagnostic plots of
- 47 instantaneous  $H_2O$  concentrations during each observation. The values of  $T_{branch}$  were
- 48 transformed from molar to volume using the molar volume of water and converted to
- 49 hourly rates (1 hour<sup>-1</sup>), prior to their conversion to transpiration per unit leaf area  $T_{leaf}$
- 50 (cf. main text).

# **S3.**Energy balance closure in eddy covariance measurements



54 Figure S1. The sum of sensible and latent heat fluxes versus net radiation (ground heat

55 flux and heat stored in the canopy are neglected) in Abisko (a) and Kevo (b).



## 57 S4. Data temporal aggregation and gap-filling

Due to technical problems of the measurement systems, gaps in temporal series were present. Not accounting for these gaps could bias the quantification of daily and seasonal aggregates of evaporative fluxes. To address this, hourly  $T_{leaf}$  and  $ET_{eco}$  were gap-filled with values predicted by the models presented in the main text (Table 2, 3). Gap-filling of  $ET_{shrub}$  and  $ET_{lichen}$  were performed using a similar modelling strategy, using PAR and VPD as predictors. The models for these two hourly fluxes showed a marginal and conditional R<sup>2</sup> of 0.54 and 0.56, respectively. However, we could not make predictions when meteorological variables were also missing. Therefore, daily aggregates of  $T_{leaf}$  were calculated when data for at least three branches and 50% of hours in each day were present; the same criterion for the minimum number of timesteps was applied for  $ET_{eco}$  and for environmental drivers such as PAR and VPD. Precipitation was calculated as the daily summation of hourly values. After this hourly gap-filling, days with missing data were imputed using daily models of the corresponding evaporative flux and its drivers. When drivers (VPD, PAR) were also missing, we gap-filledthese data using data from nearby sensors deployed within the ABACUS measuring campaigns. The overall number of hourly gaps across the growing season was overall very low for the branch bags system, and higher for the eddy covariance and the understorey chambers and the R2 of the daily models ranged between 0.7 and 0.8 (Table S2). 

**Table S2.** Percentage of hourly gaps that had to be imputed using hourly or daily models of evaporative fluxes as a function of VPD and PAR. The R<sup>2</sup> of daily models is also shown.

	% of hourly gaps	Daily model R <sup>2</sup>
Abisko		
$T_{\text{birch}}$	6%	Not needed
$\mathrm{ET}_{\mathrm{eco}}$	37%	0.70
Kevo		
$T_{birch}$	6%	0.79

$\mathrm{ET}_{\mathrm{eco}}$	25%	0.80
$\mathrm{ET}_{\mathrm{lichen}}$	44%	0.69
$\mathrm{ET}_{\mathrm{shrub}}$	42%	0.69

## 81 S5. Technical description of the automated chamber system for measuring tundra

## 32 <u>evapotranspirati</u>on

- 83 In Kevo, at ca. 150 m from the eddy flux tower, we deployed 12 PVC collars (19.9 cm
- states internal diameter and 4.5 cm height) in early June 2008 to measure four microsite types
- 85 of tundra communities, with three replicates for each type (cf. Table 1 in Poyatos et al.
- 86 2014). Three microsite types were dominated by tundra shrubs and differed in their
- 87 spatial location, both in terms of microtopography and position along the mire to forest
- 88 ecotone (Poyatos et al., 2014).
- 89 We used a closed dynamic gas exchange system for measuring H<sub>2</sub>O flux rates (mmol
- 90 H<sub>2</sub>O m<sup>-2</sup> s<sup>-1</sup>). The system comprised an infra-red gas analyser (Li-Cor 8100, Li-Cor Inc.,
- 91 Lincoln, Nebraska, USA), a custom-built multiplexed gas handler unit (Electronics
- 92 Workshop, Biology Department, University of York, UK) and 12 clear, Perspex
- 93 chambers based on a commercial soil respiration model (LiCor 8100-101; 20 cm
- 94 diameter). Chambers closed and opened sequentially, allowing hourly measurement
- 95 cycles of 12 vegetation patches at a maximum radial distance of 20 m from the
- 96 multiplexer. The chamber bases had rims with a rubber gasket, which ensured a tight fit
- 97 with PVC collars. These collars were deployed on the 12 selected patches and gently
- 98 sealed to the ground, without cutting or inserting into the substrate, using non-setting
- 99 plumber's putty (Plumber's Mait, Bostik Ltd., Leicester, UK). We took this precaution
- 100 to avoid damaging the prostrate stems and the roots of dwarf-shrub tundra species,
- 101 which could potentially affect measured fluxes. The system operated from the 11th of
- June (DOY 163) until the 14th of September (DOY 258) of 2008.
- Evapotranspiration from the chamber  $ET_{chamber}$  (mmol H<sub>2</sub>O m<sup>-2</sup> s<sup>-1</sup>) was calculated as:

104 
$$ET_{chamber} = \frac{P_{chamber} V_{chamber} d}{RT_{chamber} 0} [H_2 O]_{chamber}$$
 (Eq.S2)

where  $P_{chamber}$  is air pressure inside the chamber (Pa),  $V_{chamber}$  (m<sup>3</sup>) is the system volume (chamber, irga/multiplexer and tubing),  $T_{chamber,0}$  (°C) is air temperature at chamber closure, R (J K<sup>-1</sup> mmol<sup>-1</sup>) is the ideal gas constant,  $A_{chamber}$  is chamber surface area (m<sup>2</sup>) and d[H<sub>2</sub>O]/dt (mmol mol<sup>-2</sup>s<sup>-1</sup>) is the rate of change in water vapour in the chamber headspace. We calculated this rate from water vapour concentrations measured every 2 seconds, over the 150 s period when the chamber remained closed. We estimated d[H<sub>2</sub>O]/dt from the first order term of a quadratic fit between [H<sub>2</sub>O] and time since chamber closure. Nonlinear fits describe better the concentration dynamics in the closed chamber, and do not systematically underestimate the fluxes. However, in the presence of noisy concentration data under low flux conditions, we opted for the more stable linear fit; we selected the linear regression whenever the slope for the linear fit and the linear term of the quadratic fit had opposite signs (cf. Supporting Information S2). Because of microclimatic alterations, water vapour sorption in the tubing system and imperfect chamber sealing the automated chamber system used here has been reported to underestimates the evaporative fluxes (Cohen et al., 2015). Therefore, we applied a correction factor of ca. 2.3, obtained in this latter study, which used a similar device under comparable environmental conditions (Cohen et al., 2015).

#### 123 S6. Seasonal variation in the daily patterns of evaporative fluxes

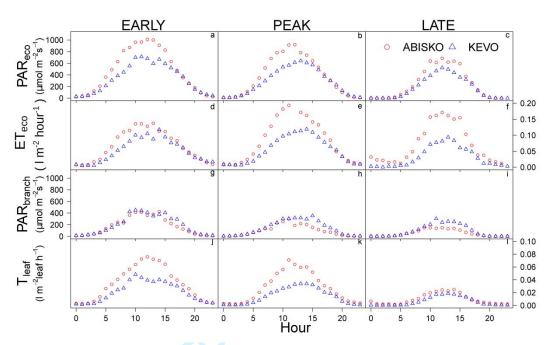


Figure S2. Seasonal variation in the mean daily patterns of ecosystem evapotranspiration ( $ET_{eco}$ ) and birch transpiration per unit leaf area ( $T_{leaf}$ ) compared to mean daily variation in PAR at the corresponding measurement scale. The growing season was split into three distinct periods according to leaf phenology (Early: 153-185, 171-185; Peak: 186-225, 186-230; Late: 226-241, 231-257, DOY in Abisko and Kevo, respectively).

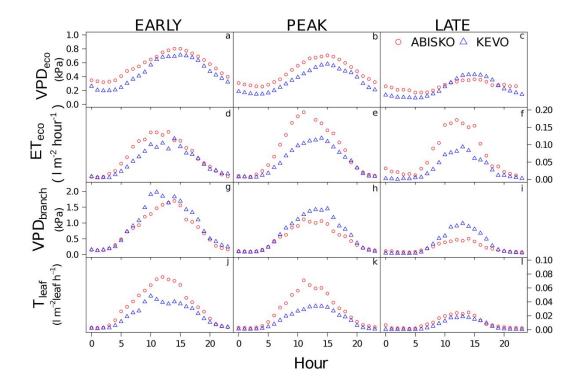


Figure S3. Seasonal variation in the mean daily patterns of ecosystem evapotranspiration ( $ET_{eco}$ ) and birch transpiration per unit leaf area ( $T_{leaf}$ ) compared to mean daily variation in VPD at the corresponding measurement scale. The growing season was split into three distinct periods according to leaf phenology (Early: 153-185, 171-185; Peak: 186-225, 186-230; Late: 226-241, 231-257, DOY in Abisko and Kevo, respectively).

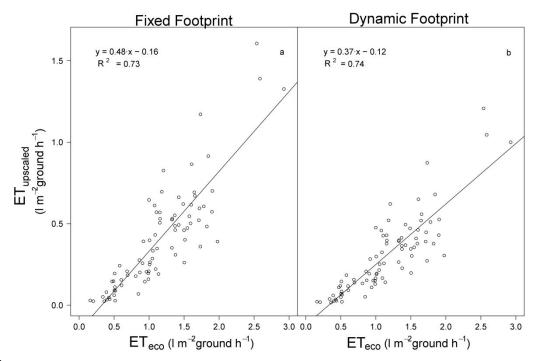
#### **S7. Footprint modeling**

- 138 The contribution of different land cover types to the eddy covariance source area was
- estimated using an analytical footprint model (Hsieh, Katul, & Chi, 2000) assuming
- 140 lateral dispersion (Detto, Montaldo, Albertson, Mancini, & Katul, 2006; Schmid, 1994).
- 141 To save computation time a look-up-table approach was used (Crawford, Grimmond,
- 142 Ward, Morrison, & Kotthaus, 2017), where the observed meteorological conditons at
- each 30-min timestep were matched to the pre-calculated land cover composition.
- 144 Intervals of 15°, 0.2 m s<sup>-1</sup> and 0.5 m s<sup>-1</sup> were used for wind direction, friction velocity,
- and standard deviation of lateral wind, respectively; and three stability classes
- accounted for stable, unstable and neutral conditions. Typically, more than 80% of the
- source area was located within 500 m of the tower, with the peak contribution at a
- 148 distance of about 23 m (36 m) in unstable (stable) conditions. As the land cover around
- the tower is a mixture of lichen, shrubs and birch, the variation in footprint composition
- with atmospheric conditions is small. (e.g. there is a slightly larger contribution from
- 151 lichen for northeasterly winds and from trees for southwesterly winds, and the
- 152 contribution of graminoid lawns and vascular plants near water is greater under stable
- conditions than unstable conditions).
- 154 The average footprint composition for the study period (DOY 160-260) is shown in
- 155 Table S2 (percentage contributions have been scaled to give a total of 100%). Trees
- 156 form the largest contribution (40%), followed by understorey (29%), followed closely
- by lichen (25%), with other land cover types contributing only a few percent at most.

**Table S3:** Average footprint composition for the study period in Kevo.

Land cover type	Average contribution to footprint [%]
Lichen	25.1
Vascular plants near water	3.4
Graminoid lawns	2.3
Understorey	29.1
Sphagnum	< 0.1
Trees	39.6
Road, boardwalk, powerlines	0.1
Open water	0.5

## 162 S8. Evapotranspiration upscaling using fixed and dynamic footprint



**Figure S4.** Linear regressions between  $ET_{eco}$  and  $ET_{upscaled}$  using a fixed (a) or (b) dynamic footprint approach.

#### S9. Variation of daily $T_{birch}/ET_{eco}$ in response to $VPD_{eco}$ , $PAR_{eco}$ and SMD.

**Table S4**. Linear models of  $T_{birch}/ET_{eco}$  in response to VPD, PAR and SMD after applying AIC-based model selection (cf. Methods). The reference level in this model is Abisko and significance codes are: \*p<0.05, \*\*p<0.01, \*\*\*p<0.001.

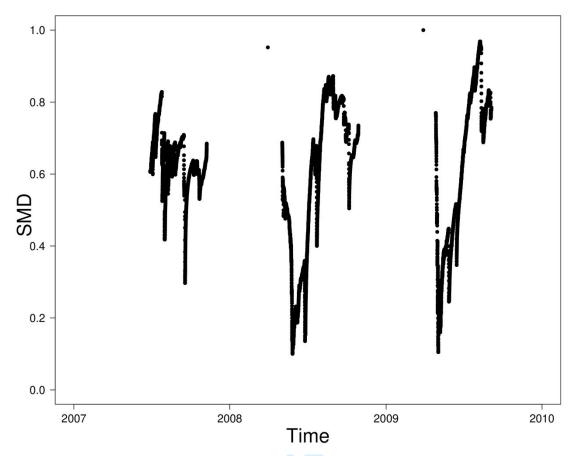
	C	1	, 1		
•	VPD model				
		Estimate	SE	t-value	p
•	(Intercept)	42.896	1.781	24.079	< 2.10-16 ***
	log(VPD)	12.554	1.689	7.432	6.19·10 <sup>-12</sup> ***
	siteKevo	-18.863	2.890	-6.527	8.56.10-10 ***
	log(VPD) : siteKevo	-4.661	2.390	-1.950	0.053
171					
•	PAR model				
		Estimate	SE	t-value	p
•	(Intercept)	-39.354	13.191	-2.984	0.0033 **
	log(PAR)	11.981	2.223	5.391	2.5·10 <sup>-7</sup> ***
	siteKevo	12.023	16.555	0.726	0.469
	Log(PAR) : siteKevo	-4.340	2.887	-1.508	0.133
172					
•	SMD model				
		Estimate	SE	t-value	p
-	(Intercept)	33.435	1.256	26.62	< 2.10-16 ***

-18.454

1.679

siteKevo

-10.99 < 2·10<sup>-16</sup> \*\*\*



75 S10. Soil moisture deficit dynamics in Abisko in 2007 - 2009

Figure. S5. Soil moisture deficit (SMD) in Abisko in 2007 – 2009 showing the continuous increase after snowmelt in 2008 and 2009.

#### 179 References

- Cohen, L. R., Raz-Yaseef, N., Curtis, J. B., Young, J. M., Rahn, T. A., Wilson, C. J., ... Newman, B. D. (2015). Measuring diurnal cycles of evapotranspiration in the Arctic with an automated chamber system. *Ecohydrology*, 8(4), 652–659. https://doi.org/10.1002/eco.1532
- Crawford, B., Grimmond, C. S. B., Ward, H. C., Morrison, W., & Kotthaus, S. (2017). Spatial and temporal patterns of surface—atmosphere energy exchange in a dense urban environment using scintillometry. *Quarterly Journal of the Royal Meteorological Society*, *143*(703), 817–833. https://doi.org/10.1002/qj.2967
- Detto, M., Montaldo, N., Albertson, J. D., Mancini, M., & Katul, G. (2006). Soil moisture and vegetation controls on evapotranspiration in a heterogeneous Mediterranean ecosystem on Sardinia, Italy. *Water Resources Research*, 42(8). https://doi.org/10.1029/2005WR004693
- Hsieh, C.-I., Katul, G., & Chi, T. (2000). An approximate analytical model for footprint estimation of scalar fluxes in thermally stratified atmospheric flows. *Advances in Water Resources*, 23(7), 765–772. https://doi.org/10.1016/S0309-1708(99)00042-1
- Kutzbach, L., Schneider, J., Sachs, T., Giebels, M., Nykanen, H., Shurpali, N. J., ... Wilmking, M. (2007). CO<sub>2</sub> flux determination by closed-chamber methods can be seriously biased by inappropriate application of linear regression. *Biogeosciences*, 4(6), 1005–1025.
- Poyatos, R., Gornall, J., Mencuccini, M., Huntley, B., & Baxter, R. (2012). Seasonal controls on net branch CO2 assimilation in sub-Arctic Mountain Birch (Betula pubescens ssp. Czerepanovii (Orlova) Hamet-Ahti). *Agricultural and Forest Meteorology*, 158–159, 90–100. https://doi.org/10.1016/j.agrformet.2012.02.009
- Poyatos, R., Heinemeyer, A., Ineson, P., Evans, J. G., Ward, H. C., Huntley, B., & Baxter, R. (2014). Environmental and Vegetation Drivers of Seasonal CO2 Fluxes in a Sub-arctic Forest–Mire Ecotone. *Ecosystems*, *17*(3), 377–393. https://doi.org/10.1007/s10021-013-9728-2
- Schmid, H. P. (1994). Source areas for scalars and scalar fluxes. *Boundary-Layer Meteorology*, 67(3), 293–318. https://doi.org/10.1007/BF00713146
- Wagner, S. W., Reicosky, D. C., & Alessi, R. S. (1997). Regression models for calculating gas fluxes measured with a closed chamber. *Agronomy Journal*, 89(2), 279–284.