

Journal Pre-proof

Development of a simulated lung fluid leaching method to assess the release of potentially toxic elements from volcanic ash

Ines TOMAŠEK, David E. DAMBY, Carol STEWART, Claire J. HORWELL, Geoff PLUMLEE, Christopher J. OTTLEY, Pierre DELMELLE, Suzette MORMAN, Sofian EL. YAZIDI, Philippe CLAEYS, Matthieu KERVYN, Marc ELSKENS, Martine LEERMAKERS



PII: S0045-6535(21)00773-6

DOI: <https://doi.org/10.1016/j.chemosphere.2021.130303>

Reference: CHEM 130303

To appear in: *ECSN*

Received Date: 24 November 2020

Revised Date: 9 March 2021

Accepted Date: 11 March 2021

Please cite this article as: TOMAŠEK, I., DAMBY, D.E., STEWART, C., HORWELL, C.J., PLUMLEE, G., OTTLEY, C.J., DELMELLE, P., MORMAN, S., YAZIDI, S.E., CLAEYS, P., KERVYN, M., ELSKENS, M., LEERMAKERS, M., Development of a simulated lung fluid leaching method to assess the release of potentially toxic elements from volcanic ash, *Chemosphere*, <https://doi.org/10.1016/j.chemosphere.2021.130303>.

This is a PDF file of an article that has undergone enhancements after acceptance, such as the addition of a cover page and metadata, and formatting for readability, but it is not yet the definitive version of record. This version will undergo additional copyediting, typesetting and review before it is published in its final form, but we are providing this version to give early visibility of the article. Please note that, during the production process, errors may be discovered which could affect the content, and all legal disclaimers that apply to the journal pertain.

© 2021 Elsevier Ltd. All rights reserved.

Ines Tomašek: *Conceptualization, Methodology, Investigation, Formal analysis, Visualization, Writing – original draft.* **David E. Damby:** *Conceptualization, Methodology, Resources, Writing – review & editing.* **Carol Stewart:** *Conceptualization, Resources, Writing – review & editing.* **Claire J. Horwell:** *Conceptualization, Writing – review & editing.* **Geoff Plumlee:** *Conceptualization, Writing – review & editing.* **Pierre Delmelle:** *Conceptualization, Writing – review & editing.* **Christopher J. Otley:** *Conceptualization, Writing – review & editing.* **Suzette Morman:** *Conceptualization, Writing – review & editing.* **Sofian El Yazidi:** *Investigation.* **Philippe Claeys:** *Funding acquisition, Resources, Supervision, Writing – review & editing.* **Mathieu Kervyn:** *Resources, Supervision, Writing – review & editing.* **Marc Elskens:** *Resources, Supervision, Writing – review & editing.* **Martine Leermakers:** *Methodology, Investigation, Writing – review & editing.*

Title

Development of a simulated lung fluid leaching method to assess the release of potentially toxic elements from volcanic ash

Authors

Ines TOMAŠEK^{1,2,*}, David E. DAMBY³, Carol STEWART⁴, Claire J. HORWELL⁵, Geoff PLUMLEE⁶, Christopher J. OTTLEY⁵, Pierre DELMELLE⁷, Suzette MORMAN⁸, Sofian EL YAZIDI¹, Philippe CLAEYS¹, Matthieu KERVYN², Marc ELSKENS¹ and Martine LEERMAKERS¹

Affiliations

¹Analytical, Environmental and Geochemistry (AMGC), Department of Chemistry, Vrije Universiteit Brussel, Brussels, Belgium.

²Physical Geography (FARD), Department of Geography, Vrije Universiteit Brussel, Brussels, Belgium.

³US Geological Survey, Volcano Science Center/California Volcano Observatory, Menlo Park, California, USA.

⁴School of Health Sciences, Massey University, Wellington, New Zealand.

⁵Department of Earth Sciences, Durham University, Durham, United Kingdom.

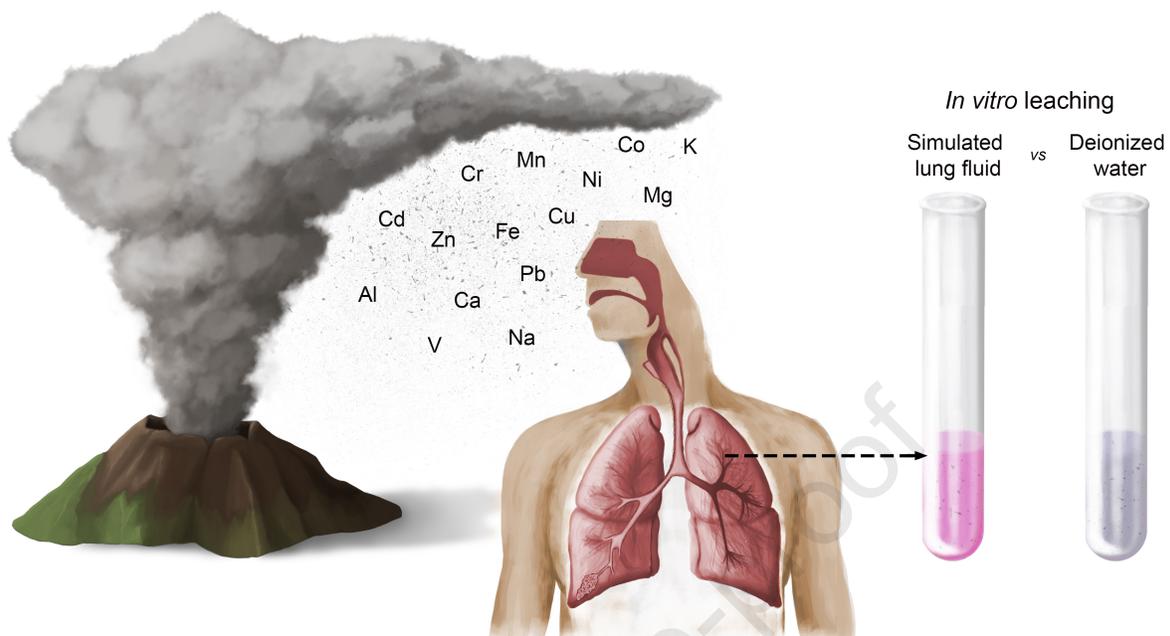
⁶US Geological Survey, Reston, Virginia, USA.

⁷Earth & Life Institute, Université Catholique de Louvain, Louvain-la-Neuve, Belgium.

⁸US Geological Survey, Denver Federal Center, Denver, Colorado, USA.

* **Correspondence:** *ines.tomasek@vub.be*

Estimation of the soluble-element hazard from inhaled ash



1 Abstract

2 Freshly erupted volcanic ash contains a range of soluble elements, some of which can generate
3 harmful effects in living cells and are considered potentially toxic elements (PTEs). This work
4 investigates the leaching dynamics of ash-associated PTEs in order to optimize a method for volcanic
5 ash respiratory hazard assessment. Using three pristine (unaffected by precipitation) ash samples, we
6 quantify the release of PTEs (Al, Cd, Co, Cr, Cu, Fe, Mn, Ni, Pb, V, Zn) and major cations typical of
7 ash leachates (Mg, Na, Ca, K) in multiple simulated lung fluid (SLF) preparations and under varying
8 experimental parameters (contact time and solid to liquid ratio). Data are compared to a standard
9 water leach (WL) to ascertain whether the WL can be used as a simple proxy for SLF leaching. The
10 main findings are: PTE concentrations reach steady-state dissolution by 24 h, and a relatively short
11 contact time (10 min) approximates maximum dissolution; PTE dissolution is comparatively stable at
12 low solid to liquid ratios (1:100 to 1:1000); inclusion of commonly used macromolecules has
13 element-specific effects, and addition of a lung surfactant has little impact on extraction efficiency.
14 These observations indicate that a WL can be used to approximate lung bioaccessible PTEs in an
15 eruption response situation. This is a useful step towards standardizing *in vitro* methods to determine
16 the soluble-element hazard from inhaled ash.

17 **Keywords:** volcanic ash, simulated lung fluid, leaching, potentially toxic elements, *in vitro* method,
18 hazard assessment

19 **1 Introduction**

20 Some elements present in particulate matter, such as Al, Cd, Fe, Ni, Pb and V, can generate harmful
21 effects in living cells (*e.g.*, oxidative stress) and, therefore, are considered potentially toxic elements
22 (PTEs) (Chen and Lippmann, 2009; Wallenborn et al., 2009). Release of PTEs in the lung
23 environment has been strongly linked with the toxicity of particles and associated adverse health
24 effects (Utembe et al., 2015; Misra et al., 2012). Characterising the presence of these soluble species
25 is thus a primary concern when assessing respiratory health hazards.

26 Freshly erupted volcanic ash contains a range of soluble compounds that are leached upon contact
27 with water or body fluids. They predominantly consist of mixed sulphate and halide salts, which are
28 replaced by gas-ash interactions in the volcanic plume and various processes throughout ash
29 transportation and deposition, leading to a variable element abundances on an ash surface (Stewart et
30 al., 2020; Ayris et al., 2015; Witham et al., 2005). The principal method used to quantify species
31 adsorbed onto ash particles is leaching (Stewart et al., 2020). Leachate analyses show that, although
32 an array of cations and anions are readily mobilised, the most abundant soluble elements are usually
33 Ca, Na and K, followed by Al, Mg, Fe and Cu, and the most common minor elements (defined as < 5
34 mg/kg ash) are Ni and Zn (Ayris and Delmelle, 2012). The release of these elements may result in the
35 contamination of water bodies and soils with potential impactsto human and animal health in ash-
36 affected areas (Stewart et al., 2020; Witham et al., 2005).

37 As a common hazard assessment strategy, leaching with simulated lung fluid (SLF) is used to
38 investigate the lung bioaccessibility of PTEs for a wide range of inhalable materials (*e.g.*, Martin et
39 al., 2018; Dean et al., 2017; Wiseman and Zereini, 2014; Wolf et al., 2011; Plumlee and Morman,
40 2011; Gray et al., 2010; Colombo et al., 2008; Twining et al., 2005). SLFs are solutions that comprise
41 of a mixture of physiologically relevant constituents (electrolytes and organic molecules) representing
42 the conditions in different compartments of the human respiratory system. Acellular *in vitro* studies
43 are easily implemented and can provide a quick and cost-effective alternative to cellular *in vitro* and
44 *in vivo* studies. Although leaching experiments do not reproduce the complex processes that occur in

45 the human body and, thus, the health relevance cannot be directly extrapolated from the results (Boisa
46 et al., 2014; Kastury et al., 2017), they provide a first-order understanding of the release of PTEs in
47 the lung environment.

48 The most commonly used SLF is known as Gamble's solution. It is a near-neutral (pH 7.4) solution
49 consisting of cations (Na^+ , K^+ , Ca^{2+} , Mg^{2+}) and anions (HCO_3^- , Cl^- , HPO_4^{2-} , SO_4^{2-}) at concentrations
50 representative of those measured in lung lining fluid (Gamble, 1967), with acetate ($\text{H}_3\text{C}_2\text{O}_2^-$) and
51 citrate ($\text{H}_5\text{C}_6\text{O}_7^{3-}$) substituting for macromolecules such as proteins and lipids, despite having different
52 biochemical functionalities. There are now varying formulations of the original Gamble's solution,
53 though modifications are often presented without a clear explanation (Kastury et al., 2017). These
54 modified solutions include organic compounds that are representative of anions and functional groups
55 in the lung lining fluid (*e.g.*, albumin, mucin, citrate, glycine, cysteine, glutathione, lactate, pyruvate,
56 etc.), which can act as chelating agents towards specific metals or metalloids of interest and may
57 promote dissolution of otherwise insoluble compounds (Caboche et al., 2011; Pelfrène et al., 2017), as
58 well as lung surfactants (*e.g.*, dipalmitoylphosphatidylcholine; DPPC), which can increase wettability
59 of particles, improve contact between leachant and metals, and prevent aggregation (Caboche et al.,
60 2011; Pelfrène et al., 2017). Such modifications add to the complexity of the solution and hinder
61 comparison among studies, particularly because the impacts of these modifications on overall
62 leaching dynamics are not understood.

63 The viability of an SLF method to assess volcanic ash has not yet been specifically tested. In contrast
64 to the highly polluted geological materials that are usually of interest for lung bioaccessibility studies
65 (*e.g.*, mine waste, soils, urban dust; Kastury et al., 2017; Plumlee and Morman, 2011), volcanic ash
66 contains very low concentrations of PTEs (Stewart et al., 2020), particularly where speciation is a
67 primary concern for toxicity (As, Cd, Cr, Hg, Pb, Se). However, ash has an abundance of generally
68 non-toxic elements like Ca, Na, S and Cl (Ayrís and Delmelle, 2012). Since some of these ions are
69 already present in SLF in substantial quantities, as either components of the SLF recipe or as
70 impurities in the reagents, this can cause signal reduction or poor precision during measurements (*i.e.*,
71 high background values). This can then cause difficulties in determining concentrations leached from

72 the ash, especially for minor elements, including PTEs, because of the necessity of making large
73 dilutions of the sample matrix. Thus, there is a clear need to test different parameters of a SLF method
74 for volcanic ash to establish their influence on apparent PTE bioaccessibility.

75 There is no consensus on leaching parameters (such as extraction time and solid to liquid ratio (S:L))
76 appropriate for the inhalation pathway on which to base a method for ash (Kastury et al., 2017).
77 Recommendations for a SLF leachate method applied to volcanic ash were an outcome of expert
78 discussions hosted by the International Volcanic Health Hazard Network (IVHHN) (Stewart et al.,
79 2013), but the method was not fully tested. Therefore, these recommendations were not included in
80 the IVHHN protocol for the rapid assessment of hazards from leachable elements in ashfall (Stewart
81 et al., 2020). This was mainly because testing of the 2013 protocol faced the aforementioned
82 difficulties associated with measurement of low PTE concentrations from ash in typical SLF
83 solutions. Additionally, there was the concern that an SLF method may not be readily spun up for
84 eruption response work because of the number and expense of analytical-grade reagents required for
85 SLF preparation. Rapid analysis and dissemination of results is the key intention of this method,
86 which sits within a broader IVHHN protocol to rapidly assess health-relevant physicochemical and
87 toxicological characteristics of volcanic ash (available at www.ivhhn.org). This leachate protocol
88 currently includes a general-purpose deionised (DI) water leach (WL) that is appropriate for assessing
89 the impacts of ashfall on water resources, such as drinking water supplies, and a ‘simple gastric’ leach
90 that is intended to estimate the bioaccessible fraction of PTEs in the event of ash ingestion by humans
91 or livestock (Stewart et al., 2020). An SLF method was not yet developed to a point where inclusion
92 would provide timely data during an eruption response.

93 To date, only three studies have addressed leaching of volcanic ash in SLF. In the first, Damby (2012)
94 investigated which minerals dissolve and the types of secondary minerals that might precipitate in the
95 lung following inhalation of volcanic ash using samples from five different volcanoes (Colima,
96 Merapi, Mt. St. Helens, Santiaguito and Unzen). After a four-week incubation of samples in SLF at 37
97 °C, a loss in mass, attributed to glass dissolution, was noted among all samples, but no new mineral
98 precipitation was observed using X-ray diffraction. In the second study, Tomašek et al. (2019) leached

99 synthetic ash laden with sulphate salts in water and determined that the majority of salts dissolved
100 within 10 minutes. Using the resulting water-leach concentrations, saturation in SLF was simulated by
101 reaction-path modelling. The SLF was under-saturated in sulphate salts (CaSO_4 , Na_2SO_4 and MgSO_4),
102 suggesting that no new phases were being formed and the predominant salt deposits found on ash
103 surfaces would dissolve in lung fluid, likely prior to cellular uptake. In a recent study, Barone et al.
104 (2020) quantified the soluble element burden of volcanic ash samples from Etna volcano in water and
105 SLF according to the initial recommendations by Stewart et al. (2013). They found that the
106 concentrations of elements released in SLF are lower than those measured in water.

107 The primary objective of this study is to develop a robust *in vitro* method to assess the release of PTEs
108 from ash in the lung environment to evaluate the hazard of volcanic eruptions to public health. Given
109 the scarcity of past ash leaching studies on which to base such a method, and the general lack of
110 studies reporting comparative efficiencies of existing SLF approaches, this work establishes the effect
111 of different SLF compositions and extraction parameters (ash to leachant ratio and extraction time) on
112 the leaching efficiency of PTEs. This is a critical first step in the development of a standardised
113 method for health hazard assessment and inclusion with other, existing IVHHN ash analysis protocols
114 as the lung lining fluid is the first interface that inhaled materials come into contact with in the
115 airways. A second objective of this work was to compare element leaching efficiency in SLF and DI
116 water (*i.e.*, IVHHN's general-purpose WL). Toxicology studies of particulate matter indicate that
117 water-soluble elements may be associated with toxic effects in the lungs (Oller et al., 2009; Costa and
118 Dreher, 1997), and that their release into water may differ from their release into an SLF (Caboche et
119 al., 2011; Pelfrène et al., 2017). Hence, the present experiments were designed to ascertain whether
120 WL could be used as a proxy for SLF leaching of volcanic ash. DI water is the most common leachant
121 for ash studies (Stewart et al., 2020) due to its wide availability. Its use for rapid respiratory hazard
122 assessment would bolster data comparability with previous assessments, and eliminate the need to
123 perform multiple leachate analyses on a sample, which are often difficult to obtain in sufficient
124 amounts.

125 2 Materials and Methods

126 2.1 Volcanic ash samples

127 For this study, three ash samples from recent volcanic eruptions were selected (**Table 1**). The samples
 128 were collected fresh (unaffected by precipitation or surface weathering) from ashfall deposits
 129 according to IVHHN ash collection recommendations (Stewart et al., 2020) and stored into self-
 130 sealing plastic bags. The SLF leaching experiments (*Section 2.3*) were carried out in batch conditions
 131 on bulk volcanic ash, which had not been oven dried. Each sample was homogenized by gently
 132 rotating in a sealed container before taking a sub-sample for leaching experiments.

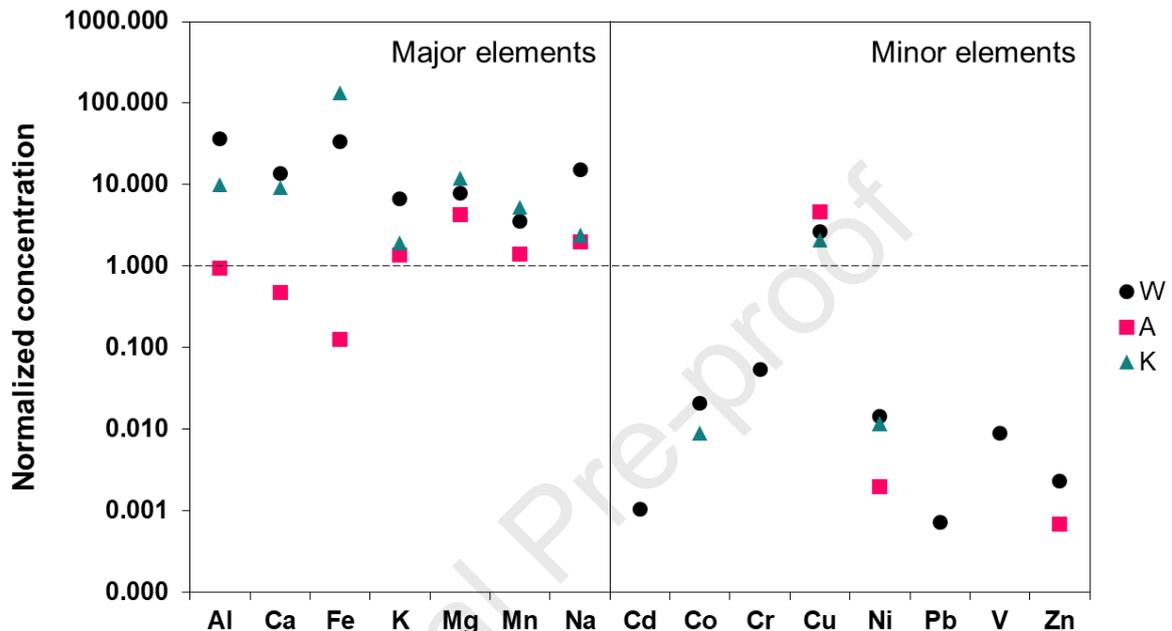
133 **Table 1** Sample and collection information for the volcanic ash samples used in this study.

Volcano	Country	Eruption date	Magma type	Collection date	Collection location
Ambae	Vanuatu	17/03/2018	Basalt to trachybasalt ¹	17/03/2018	Vinangangwe, West Ambae
Kīlauea	USA	10-28/05/2018	Basalt ²	28/05/2018	Ka'ū Desert, Hawai'i
Whakaari/White Island	New Zealand	27/04/2016	Andesite ³	28/04/2016	North rim of the crater

134 ¹Moussallam et al. (2019), ²Neal et al. (2019), ³Mayer et al. (2015).

135 The sample selection was based on available mass to test a number of different leaching parameters,
 136 and because samples were previously characterised for their water-leachable element content. All
 137 three ash samples have relatively high leachable concentrations. Element concentrations (in mg/kg dry
 138 weight ash) were determined by WL for 1 h at a solid to liquid ratio of 1:100 (Damby et al., 2018;
 139 Stewart C., unpublished) using a standardised protocol by Stewart et al. (2020). These concentrations
 140 were normalized to mean concentrations reported in a global dataset on water-extractable elements
 141 from volcanic ash (Ayrís and Delmelle, 2012) (**Fig. 1**, Appendix A). All three samples have abundant
 142 soluble major elements (> 5 mg/L global mean concentration), particularly the ash from
 143 Whaakari/White Island volcano. Water-soluble minor elements (< 5 mg/L global mean
 144 concentration), except for Cu, are present in lower concentrations than global means and some were

145 below detection limits (BDL) in Ambae and Kīlauea ash (namely Co in Ambae, Zn in Kīlauea, Cd,
 146 Cr, Pb and V in both). These 1 h WL data are compared with the concentrations determined in the
 147 present study using the same ash to leachant ratio (1:100) for 24 h WL and SLF at different time
 148 points in *Section 3.4*.



149
 150 **Figure 1** Water-extractable major (> 5 mg/L) and minor (< 5 mg/L) element abundances in analysed
 151 ash samples (black circles – Whaakari/White Island, red squares – Ambae, green triangles – Kīlauea)
 152 normalized to global mean values (Ayriss and Delmelle, 2012). Elements not shown were below
 153 detection limit (Appendix A).

154 2.2 Selection of leaching parameters

155 The recommendations for ash leaching in SLF provided by the expert working group convened by
 156 IVHHN to standardize the leachate protocols are taken as a starting point for our SLF testing (Stewart
 157 et al., 2013). These include an SLF based on Gamble's solution (Table 2), a contact time of 24 h and a
 158 1:100 ash (mass) to leachant (volume) ratio. In the following sections, we summarise the parameter
 159 modifications tested in this study.

160 2.2.1 Leachant

161 In general, the dissolution rate of elements is dictated by their solubility in different media (*e.g.*, water
162 or SLF) which is predominantly controlled by the pH and composition of the solution (Kastury et al.,
163 2017; Misra et al., 2012). Earlier studies argued that simple leachants, such as water, are not
164 ‘physiologically based’ and, thus, are not representative of pulmonary exposure (reviewed in Kastury
165 et al. (2017)). It is thought that a WL may underestimate the lung bioaccessibility of metal
166 components due to the absence of organic compounds.

167 In addition to deionized water, we used four different SLF solutions of a near-neutral pH, all with the
168 same base composition (see **Table 2**). This range of solution compositions allows determination of the
169 effects of inclusion, or exclusion, of commonly used molecules (glycine, citrate) and a surfactant
170 (DPPC) to deduce whether they are essential components for assessing the ash hazard. For these
171 experiments, the 24 h time-point and 1:100 solid to liquid (S:L) ratio were set as constant.

172 2.2.2 Extraction time

173 To be relevant for inhalation exposures, the sample extraction time *in vitro* should be representative of
174 the residence time of particles in the lung. This is difficult to constrain as particle removal depends on
175 the deposition site within the lungs and clearance mechanisms involved (mucociliary transport, uptake
176 by phagocytic cells, *in situ* dissolution, etc.). These processes operate on the order of hours to days
177 and months (Bailey et al., 2007; Morgan et al., 2004; Gehr et al., 1990). It has also been argued that
178 the short-term toxic effect of particles that release ions at a fast rate could be identical to those of the
179 dissolved ions (Studer et al., 2010), whereas, for particles that release ions at a slow rate, there is a
180 greater likelihood that the particles will be the cause of the observed adverse effects (Oberdörster,
181 2000).

182 The contact times used in previous lung bioaccessibility studies on non-volcanic material vary greatly,
183 ranging from 5 min to 1 year (Kastury et al., 2017), with most studies using ≤ 24 h. Reported
184 timeframes in ash-water leachate studies range from 5 min to 2 years (Stewart et al., 2020). A shorter

185 duration timepoint is supported by previous work on volcanic ash leachate studies. Ash leachate
186 studies using water demonstrated that the majority of surfaces phases were dissolved within the first
187 10-15 minutes (Olsson et al., 2013; Duggen et al., 2007), and most sulphate salts were dissolved
188 within 10 minutes (Tomašek et al., 2019).

189 To determine how extraction varies over a relevant timeframe for volcanic ash in SLF, we tested
190 contact times of 10 min, 1 h, 4 h, 24 h and 48 h, each as a separate experiment. The 24 h time-point
191 was kept as a constant parameter for the S:L ratio and leachant composition tests. This is the time-
192 point at which the measured concentrations of all elements became approximately stable and provides
193 a direct comparison with cellular toxicity tests which commonly use 24 h exposures.

194 **2.2.3 Solid to liquid ratios**

195 The mass of particles that deposits in the lung (*i.e.*, particle loading) following inhalation is variable
196 and largely dependent on ambient particle concentrations, size distribution and personal exposure. To
197 determine *in vitro* S:L ratios relevant for real human exposure, we calculated the potential particle
198 loading following a single ash exposure. We assume 100% particle deposition and a daily inhaled air
199 volume of 25 m³, corresponding to a healthy, moderately-active adult (ICRP, 1994), for airborne ash
200 concentrations of 0.02 and 1 mg/m³, reported as minimum and maximum daily averages in the
201 literature (Searl et al., 2002). Using 20 mL as the total volume of lung lining fluid (Macklin, 1955),
202 when ambient concentrations of ash are low (0.02 mg/m³) the S:L ratio would correspond to 1:4000,
203 whereas when ambient concentrations are high (1 mg/m³) it would be 1:800. In case of an exposure to
204 higher ambient concentrations, which could be experienced during ash clean-up activities (*e.g.*, 10
205 mg/m³) (Searl et al., 2002), the ratio could be even larger and equate to 1:80.

206 The S:L ratios used in lung bioaccessibility studies to date range from 1:20 to 1:50000, with the
207 majority of studies using ratios < 1:100 (Kastury et al., 2017), whereas ash-water leachate studies use
208 ratios from 1:5 to 1:1000 (Stewart et al., 2020). The ratios tested in this study (1:10, 1:20, 1:100,
209 1:500, 1:1000) reflect this wide range, accounting for experimental/analytical constraints, and include

210 the existing IVHHN recommended ratios of 1:20 and 1:100 (Stewart et al., 2020). The 1:100 ratio was
211 selected as a constant in the time-series and varying leachant composition tests.

212 **2.2.4 Other method parameters**

213 Other parameters that may affect extraction and have direct relevance to the respiratory system but
214 were not specifically tested in the framework of this study are temperature and particle size. Whereas
215 most SLF studies incubate particles at 37 °C to replicate body temperature (Kastury et al., 2017), we
216 performed the extractions at room temperature (25 °C) as a system to maintain the elevated
217 temperature and agitate the samples simultaneously was not available. This was justified through
218 preliminary experiments that compared extraction at 25 °C and 37 °C (without agitation) and resulted
219 in little difference in leaching efficiency (*unpublished data*). Therefore, we leached all samples at
220 room temperature to reduce experimental complexity.

221 Considering that the focus of the current exercise was on the methodological parameters affecting
222 bioaccessibility rather than sample properties, we choose to perform the extractions on bulk (un-
223 sieved) ash samples. Isolation of respirable (sub-4 µm) material from bulk ash, in amounts sufficient
224 for leachate analysis, is time-consuming and often impractical. The percentage of sub-4 µm particles
225 also varies greatly among samples, depending on the sample collection distance from the vent and the
226 magnitude and explosivity of the eruption, but is typically < 17 % (Horwell, 2007; Horwell and
227 Baxter, 2006). This is the case for sub-10 µm particles as well, which are generally used in lung
228 bioaccessibility assessment studies (Kastury et al., 2017). While it is assumed that analysis of the
229 respirable fraction may be more predictive of real bioaccessibility, and is likely to give higher
230 concentration values than those of bulk extractions due to higher particle surface area, it is not always
231 possible to demonstrate the size effect in isolation from other properties (Misra et al., 2012).

232 **2.3 Leaching experiments**

233 Assay parameters evaluated (see **Table 3**) included composition of the leachant, contact time and ash
234 to leachant (S:L) ratio to determine their influence on the leaching efficiency (**Table 3**). Each test was
235 performed in triplicate. Depending on the experiment, different amounts of ash were weighed into 50-

236 mL polypropylene centrifuge tubes and corresponding volumes of leachant were added. Samples were
 237 then agitated on a platform (horizontal) shaker at 60 rpm at room temperature throughout the
 238 extraction duration. Subsequently, samples were centrifuged for 5 min at $3392 \times g$ and filtered
 239 through $0.45 \mu\text{m}$ cellulose acetate membrane filters (VWR Chemicals, Belgium) into 15 mL
 240 polypropylene tubes using syringe filtration. Leachates were acidified with concentrated nitric acid
 241 (HNO_3) and stored at $4 \text{ }^\circ\text{C}$ until analysis.

242 In order to keep the background concentrations low, all reagents used to prepare the SLF (**Table 2**)
 243 were of analytical grade (*AnalaR® NORMAPUR®*), purchased from VWR Chemicals (Belgium). The
 244 solutions were prepared in deionized water (Milli-Q®, resistivity of $18.2 \text{ M}\Omega$). The pH of the solution
 245 was adjusted to 7.40 ± 0.05 using concentrated hydrochloric acid (HCl).

246 **Table 2** Composition of the simulated lung fluid (SLF) solutions used in this study (see *Section*
 247 *2.2.1*). Base composition (SLF1) is after Stewart et al. (2013), and modifications are to include lung
 248 surfactant (DPPC; SLF2) and to remove citrate (SLF3) or glycine (SLF4). All concentrations are as
 249 mg/L.

	SLF1	SLF2	SLF3	SLF4
NaCl	6400	6400	6400	6400
$\text{CaCl}_2 \cdot 2\text{H}_2\text{O}$	255	255	255	255
Na_2HPO_4	150	150	150	150
NaHCO_3	2700	2700	2700	2700
NH_4Cl	118	118	118	118
$\text{MgCl}_2 \cdot 6\text{H}_2\text{O}$	212	212	212	212
$\text{Na}_2\text{SO}_4 \cdot 10\text{H}_2\text{O}$	179	179	179	179
$\text{Na}_3 \text{ citrate} \cdot 2\text{H}_2\text{O}$	160	160	-	160
Glycine	190	190	190	-
DPPC* (0.01%)	-	100	-	-

250 *DPPC = 1,2-Dipalmitoyl-sn-glycero-3-phosphocholine

251 **Table 3** Experimental parameters: leachant, extraction time and S:L ratio. Each test was performed in
 252 triplicate.

SLF1	SLF2	SLF3	SLF4	WL*
------	------	------	------	-----

10 min	1:100
1 h	1:100
4 h	1:100
	1:10
	1:20
24 h	1:100 1:100 1:100 1:100 1:100
	1:500
	1:1000
48 h	1:100

253 *WL = water leach, using deionized water (Milli-Q[®], resistivity of 18.2 M Ω , pH 7.95 \pm 0.05)

254 **2.4 Trace elements analysis**

255 We analysed a large suite of PTEs (Al, Cd, Co, Cr, Cu, Fe, Mn, Ni, Pb, V, Zn) and cations that are
 256 the main constituents of ash surface coatings (Ca²⁺, Na⁺, Mg²⁺, K⁺). All element concentrations were
 257 measured using high resolution inductively coupled plasma mass spectrometry (HR-ICP-MS; Thermo
 258 Finnigan Element II) at the facilities of the Analytical, Environmental and Geochemistry group of the
 259 Vrije Universiteit Brussels. Calibration curves of the selected elements were made from dilutions of
 260 an acidified multi-element stock solution (ICP-MS Calibration standard 2, VWR Chemicals, Belgium)
 261 and single element standards (Certipur[®] 1000 ppm, Merck, Belgium), with Rh103 as internal
 262 standard. The procedural blanks and experimental samples were diluted 10-fold in 2 % HNO₃ solution
 263 prior to the analysis. Operational parameters are listed in the Appendix B (**Tables B.1** and **B.2**). The
 264 limits of quantification (LOQs) were calculated as 3-fold standard deviation of the mean elemental
 265 concentration measured in the leachates of procedural blanks which underwent the whole extraction
 266 procedure (**Table 4**).

267 **Table 4** Limits of quantification (LOQ) for selected elements in $\mu\text{g/L}$ obtained by HR-ICP-MS in
 268 deionized water (DI) and SLF solutions, calculated as three times the SD for the mean of n=8
 269 procedural blanks for SLF1 and n=3 for DI, SLF2, SLF3 and SLF4.

	DI	SLF1	SLF2	SLF3	SLF4
Na	100	94831	506282	412779	425928
Ca	5	1340	7445	6320	6451
Mg	0.7	634	3471	2824	2752
K	1.3	5.4	35	5.1	204
Al	2.5	31.3	3.4	3.7	14.1
Fe	1.2	3.3	10.9	2.6	1.2
Mn	0.06	1.10	0.27	0.12	0.11
Cu	0.11	0.21	0.11	0.13	0.08
Cd	0.003	0.008	0.004	0.007	0.009
Co	0.006	0.005	0.010	0.005	0.002
Cr	0.23	0.30	0.26	0.27	0.25
Ni	0.24	0.33	0.28	0.26	0.26
V	0.006	0.027	0.009	0.004	0.008
Pb	0.008	0.116	0.027	0.019	0.030
Zn	0.84	10.77	1.59	1.00	1.32

270

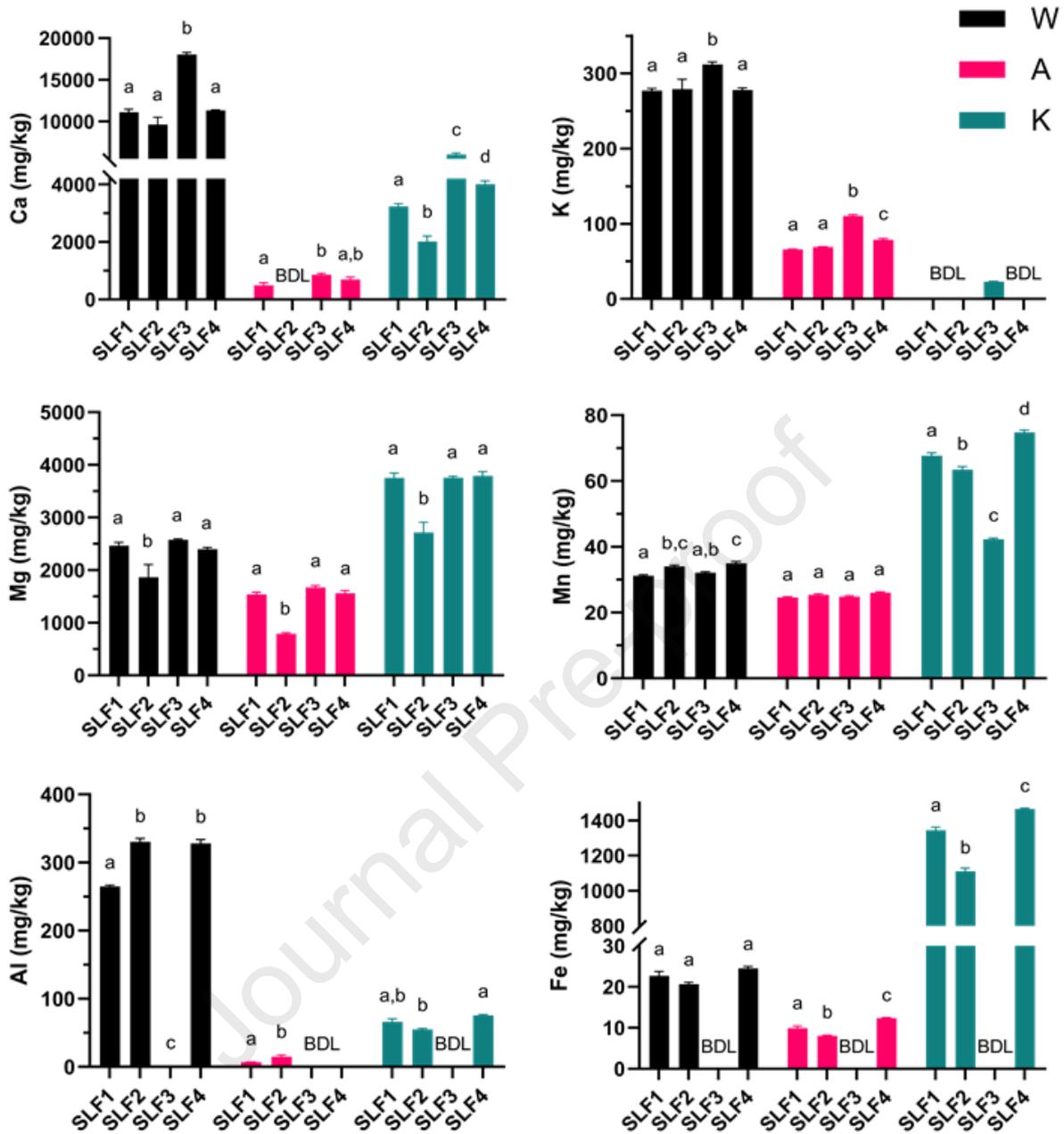
271 **2.5 Data processing and statistical analysis**

272 Graphical representation and statistical analysis of the data were performed using GraphPad Prism
273 (version 8.3.0; *GraphPad Software*, San Diego, CA). Statistical significance between different
274 experimental parameters was determined using a one-way analysis of variance (ANOVA) with
275 subsequent Tukey's tests. The alpha value was set at 0.05. In the figures, significant differences are
276 denoted by lowercase letters; for all parameters with the same letter, the difference between the means
277 is not statistically significant ($p > 0.05$), whereas, for parameters with a different letter, the difference
278 is statistically significant ($p < 0.05$).

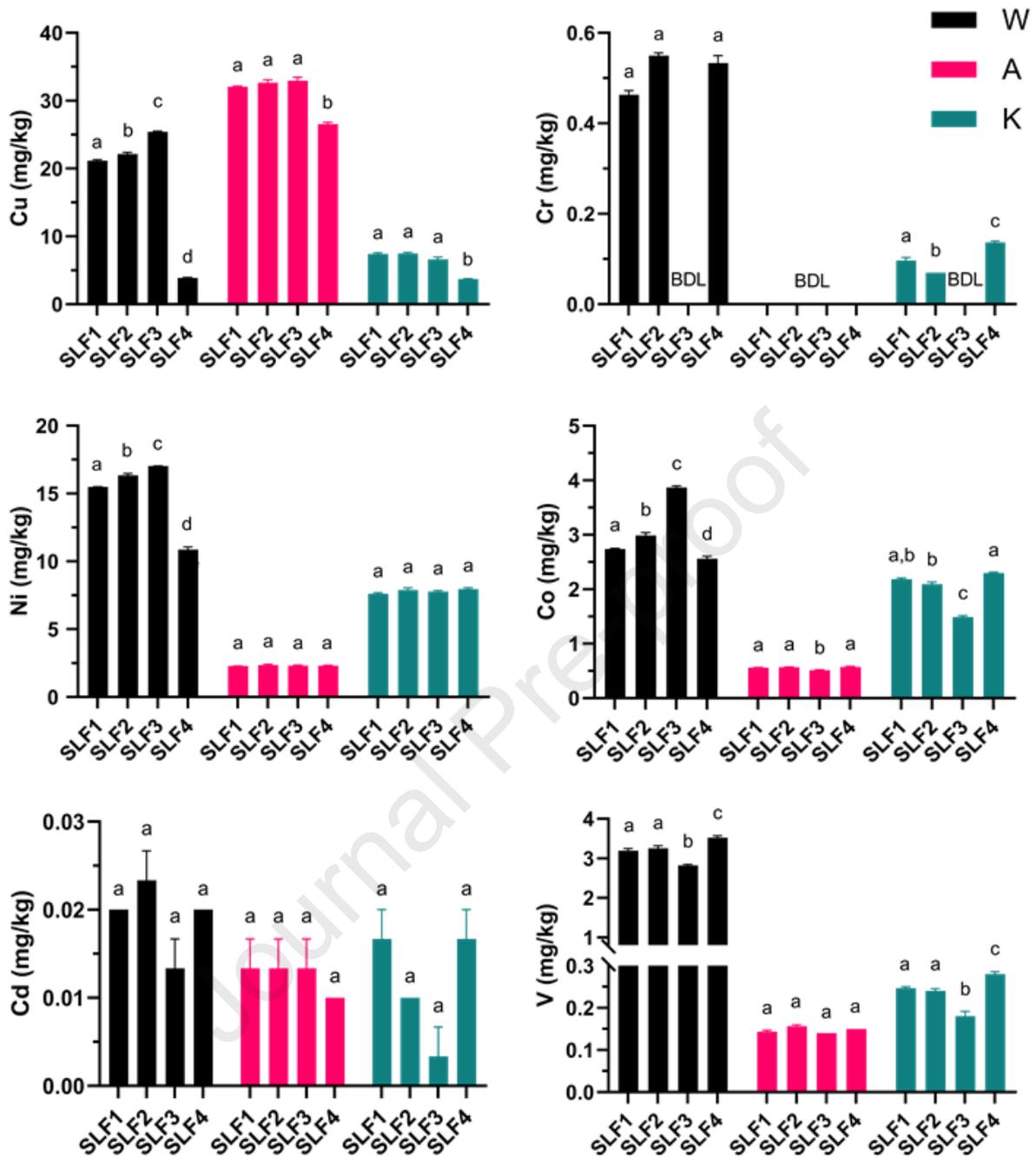
279 **3 Results**

280 **3.1 Influence of leachant composition on PTE release**

281 The results of ash extraction in four preparations of SLF (SLF1-SLF4) are shown in **Fig. 2** (major
282 elements) and **Fig. 3** (minor elements). All extractions were for 24 h at 1:100. Across the four SLFs
283 tested, there were notable differences in measured concentrations that were consistent for all ash
284 samples: Al, Fe (**Fig. 2**) and Cr (**Fig. 3**) were all found BDL in SLF3; Cu was significantly ($p < 0.05$)
285 lower in SLF4 (**Fig. 3**); Mg was significantly ($p < 0.05$) lower in SLF2, but was otherwise equivalent
286 ($p > 0.05$) across all leachants (**Fig. 2**); Ca and K were higher in SLF3 (**Fig. 2**). The concentration of
287 leached Cd was the same ($p > 0.05$) in all SLFs (**Fig. 3**). There were no differences in concentrations
288 of K (**Fig. 2**) and minor elements determined in SLF1 and SLF2 (**Fig. 3**). While low levels of Pb and
289 Zn were leached in water (Appendix C), their concentrations were BDL in all SLFs. The substantial
290 component of Na in SLF solutions (**Table 4**) prevents the accurate quantification of Na (Appendix C).
291 Regarding the extraction behaviour among the three samples, some exceptions could be observed
292 across the different leachants. The concentration of Mn was largely comparable in different SLFs for
293 Whaakari/White Island and Ambae ash but showed significant ($p < 0.05$) differences in Kīlauea ash
294 (**Fig. 2**). Although following the same qualitative pattern for all samples, the concentration of Fe was
295 not statistically different ($p > 0.05$) across all SLFs for Whaakari/White Island, whereas it was
296 significantly different ($p < 0.05$) across all SLFs for Ambae and Kīlauea ash (**Fig. 2**). While
297 concentrations of Cu and Ni were statistically equivalent in SLF1-SLF3 in Ambae and Kīlauea ash,
298 this was not the case in leachates of Whaakari/White Island ash (**Fig. 3**). Leaching of V was similar
299 for Whaakari/White Island and Kīlauea ash, with lower concentrations in SLF3, in contrast to Ambae
300 ash where the concentrations were equivalent across all SLFs (**Fig. 3**). Concentrations of Co indicated
301 different magnitudes of leaching across all samples (**Fig. 3**).



302
 303 **Figure 2** Major element concentrations (> 5 mg/kg, global mean concentration) in three ash samples
 304 (black – Whaakari/White Island, red – Ambae, green – Kīlauea) obtained through the extractions in
 305 four different simulated lung fluids (24 h, 1:100 S:L). Data are reported as mg element per kg of ash
 306 dry weight and represented as the mean of three replicates for each sample. Error bars are the standard
 307 error of the mean. Lowercase letters indicate a significant difference ($p < 0.05$) between the mean
 308 concentrations of leachants for each ash sample.

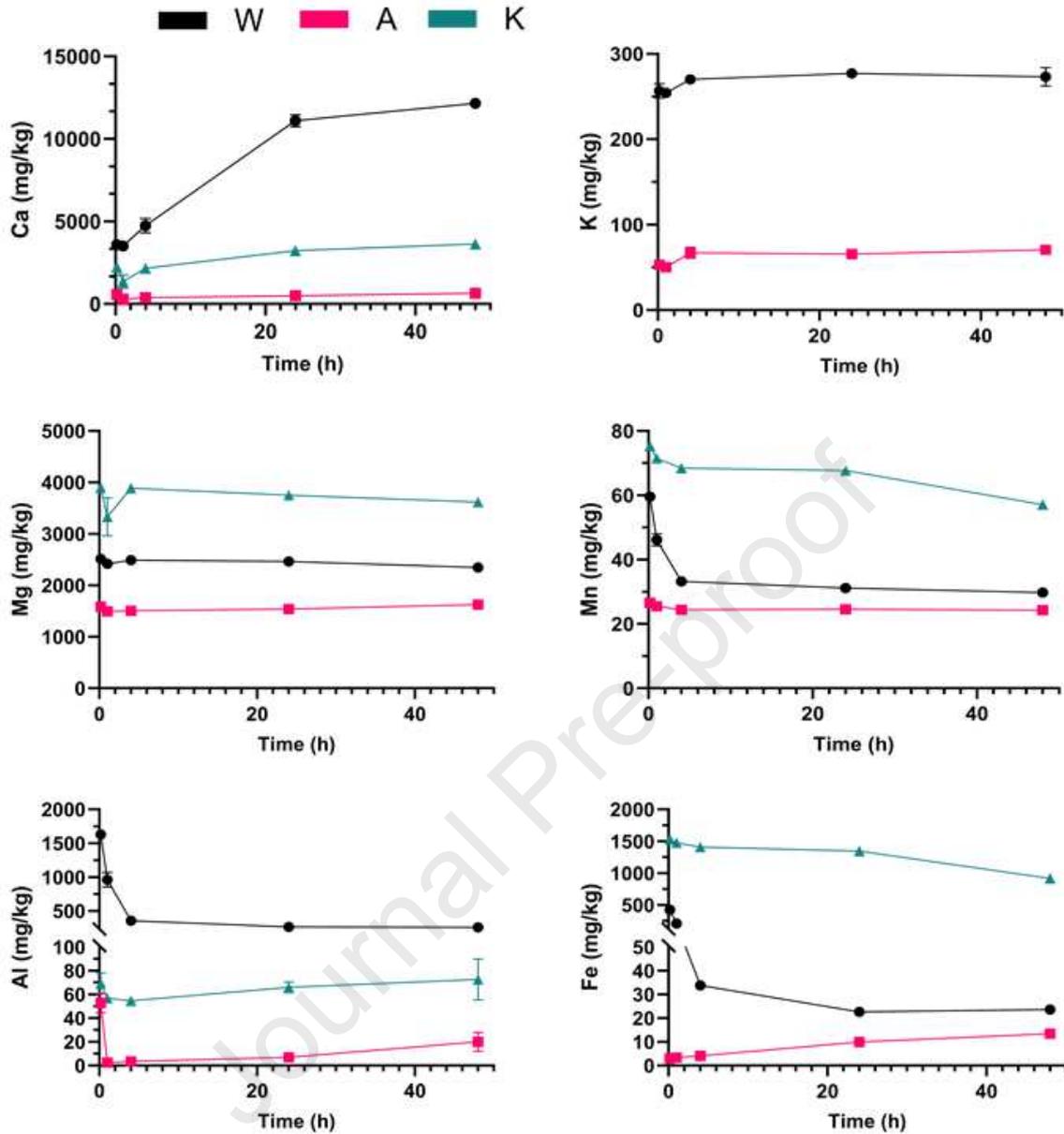


309

310 **Figure 3** Minor element concentrations (< 5 mg/kg, global mean concentration) in three ash samples
 311 (black – Whaakari/White Island, red – Ambae, green – Kīlauea) obtained through the extractions in
 312 four different simulated lung fluids (24 h, 1:100 S:L). Data are reported as mg element per kg of ash
 313 dry weight and represented as the mean of three replicates for each sample. Error bars are the standard
 314 error of the mean. Lowercase letters indicate a significant difference ($p < 0.05$) between the mean
 315 concentrations of leachants for each ash sample.

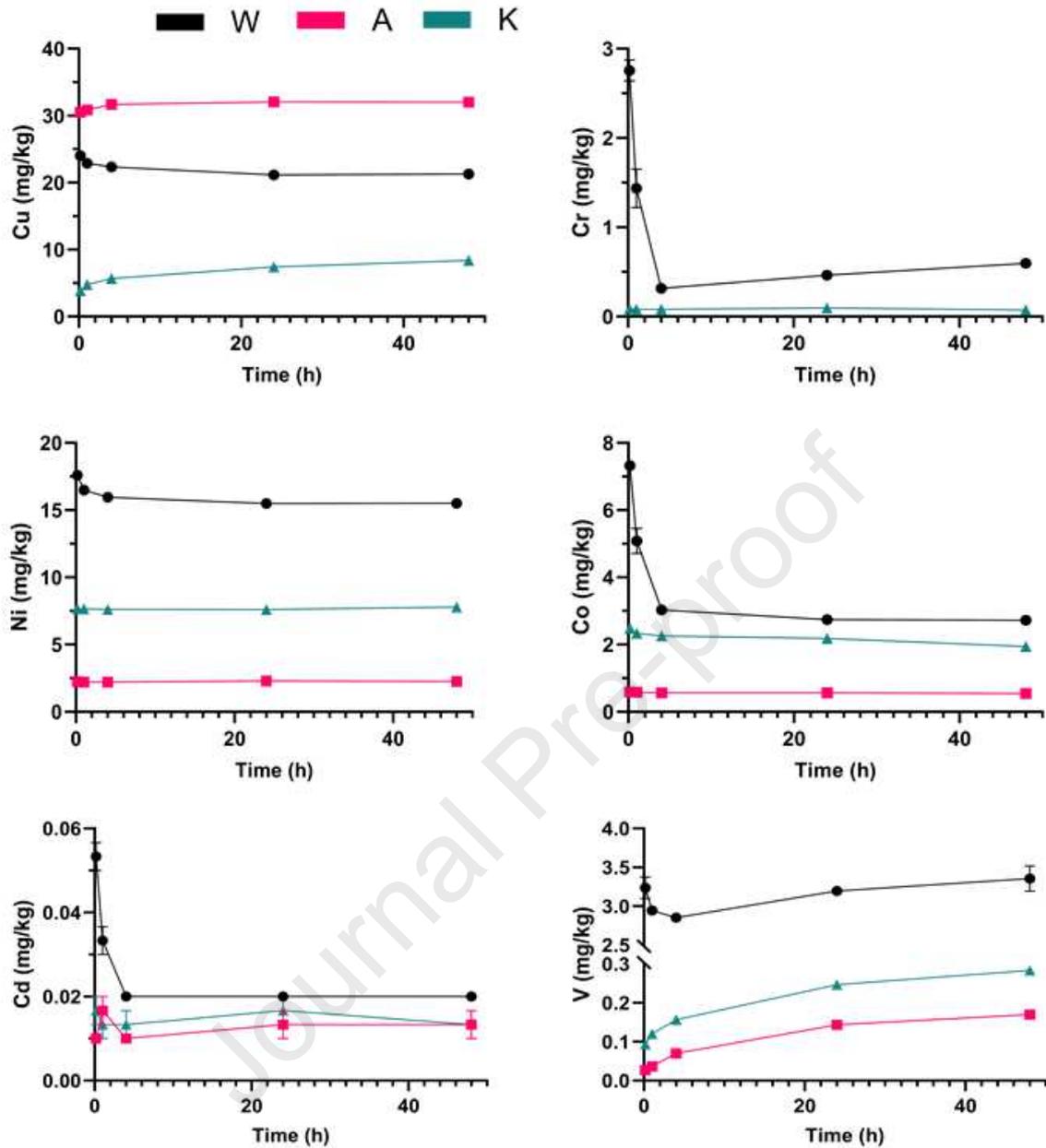
316 3.2 Influence of extraction time on PTE release

317 Release of PTEs in SLF was tracked over five time intervals, from 10 min to 48 h. Overall, the
318 concentration of individual elements over different time intervals varied among the different samples.
319 Only Mg concentrations remaining constant ($p > 0.05$) across all measured time-points in all samples
320 (**Fig. 4**). Except for Fe in Ambae ash, the time-series for Al, Fe and Mn were largely consistent with
321 their highest concentrations recorded after the initial 10 min leaching period, after which
322 concentrations decreased by the 4 h time-point and then remained stable, as seen from the little
323 difference ($p > 0.05$) between concentrations at 4 h, 24 h and 48 h time-points (**Fig. 4**). This trend was
324 the most prominent for Whaakari/White Island ash leachates and could also be observed for leached
325 Cd, Co, Cr and Ni, whereas their concentrations were constant over time in Ambae and Kīlauea ash
326 (**Fig. 5**). Similar behaviour in Whaakari/White Island was exhibited for Cu and V, whereas their
327 concentrations slightly increased by the 48 h time-point in Ambae and Kīlauea ash (**Fig. 5**). Leached
328 Ca and K showed a similar qualitative pattern, with a decrease at 1 h compared to the initial
329 concentration at 10 min then followed by an increase over time (**Fig. 4**).



330

331 **Figure 4** Major element concentrations (> 5 mg/kg, global mean concentration) in three ash samples
 332 (Whaakari – black, Ambae – red, Kīlauea – green) obtained through the extractions over varying
 333 time-points (for 1:100 S:L in SLF1). Data are reported as mg element per kg of ash dry weight and
 334 represented as the mean of three replicates for each sample. Error bars are the standard error of the
 335 mean. Where error bars are not visible, they are less than the size of the symbol.

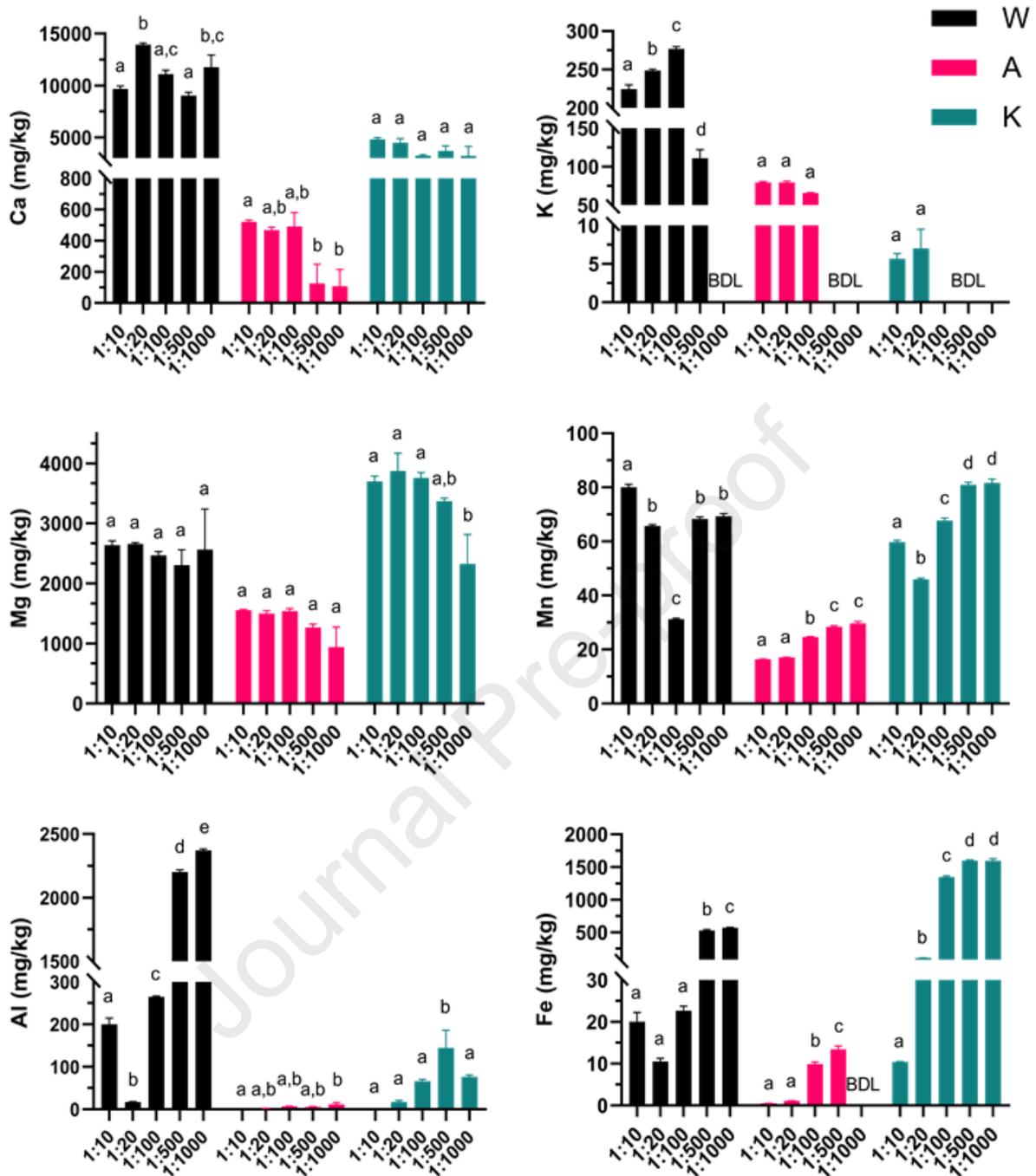


336
337

338 **Figure 5** Minor element concentrations (< 5 mg/kg, global mean concentration) in three ash samples
 339 (Whaakari – black, Ambae – red, Kīlauea – green) obtained through the extractions over varying
 340 time-points (for 1:100 S:L in SLF1). Data are reported as mg element per kg of ash dry weight and
 341 represented as the mean of three replicates for each sample. Error bars are the standard error of the
 342 mean. Where error bars are not visible, they are less than the size of the symbol.

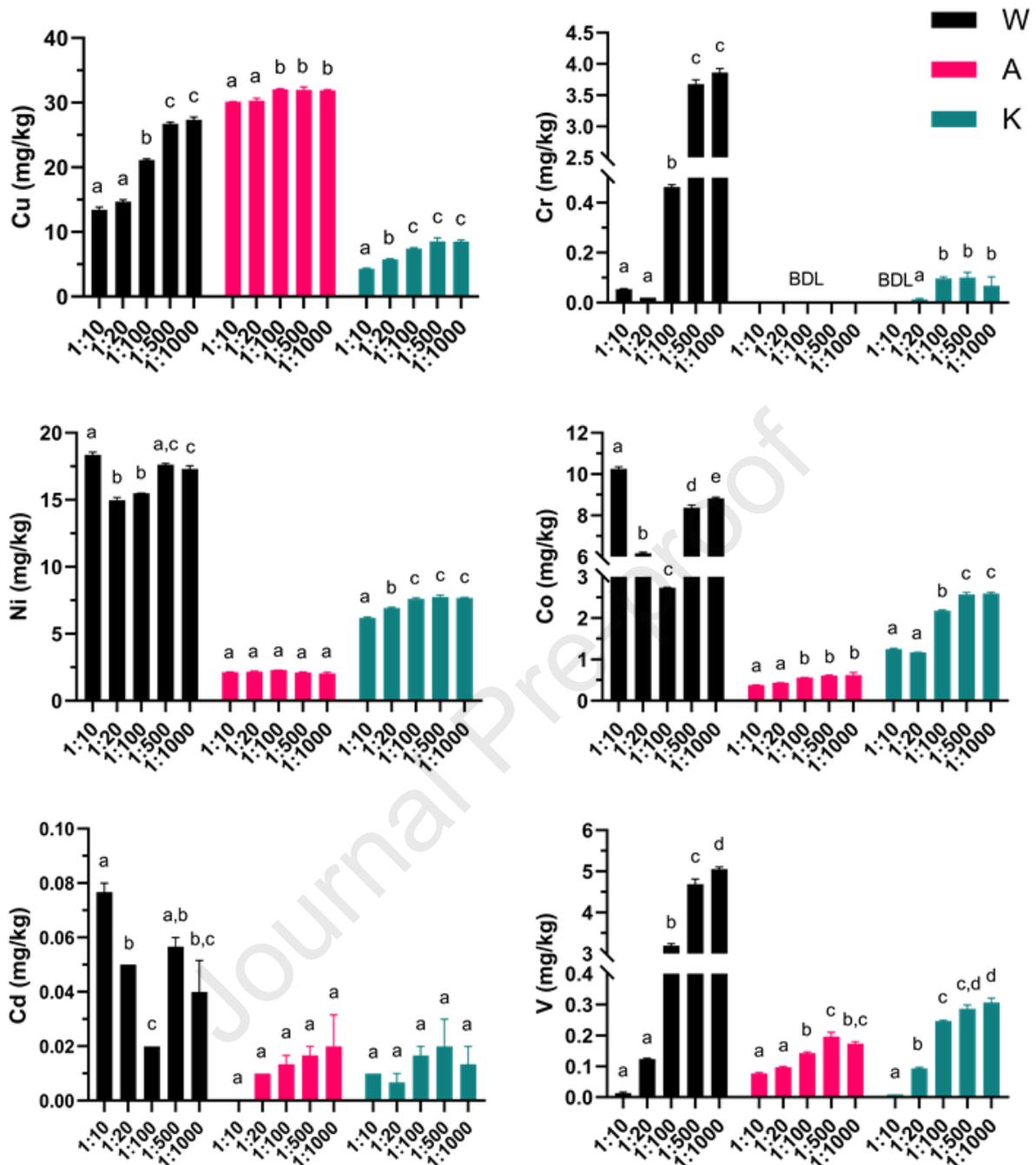
343 3.3 Influence of the ash to leachant ratio on PTE release

344 Five S:L ratios, ranging from 1:10 to 1:1000, were used to assess the influence of ash to leachant ratio
345 on the PTE dissolution from ash in SLF. Among the different tested ratios, concentrations of Mg were
346 found to be stable in all samples (**Fig. 6**). Concentrations at the two lowest S:L ratios, 1:500 and
347 1:1000, showed no significant differences ($p > 0.05$) for Cd, Cr, Cu, Mg, Mn and Ni across all
348 samples, whereas K was lower or BDL in these ratios. As an overall trend, recorded element
349 concentrations increased with decreasing sample loading for Al, Co, Cr, Cu, Fe, Mn and V, and were
350 largely independent of S:L ratios in a range 1:100 to 1:1000 for Ambae and Kīlauea ash, with the
351 exceptions observed in Whakaari/White Island leachates which generally had higher concentrations
352 for these elements. For Whakaari/White Island ash, there was a drop in concentration at 1:20 and/or
353 1:100 for Al, Co, Fe and Mn. This was the case for Cd as well, which was found to be constant in all
354 S:L ratios for Ambae and Kīlauea samples (**Fig. 7**). Ni concentrations varied across different samples
355 (**Fig. 7**).



356

357 **Figure 6** Major element concentrations (> 5 mg/kg, global mean concentration) in three ash samples
 358 (Whaakari/White Island – black, Ambae – red, Kīlauea – green) obtained through the extractions in
 359 varying ash to leachant (S:L) ratios (24 h in SLF1). Data are reported as mg element per kg of ash dry
 360 weight and represented as the mean of three replicates for each sample. Error bars are the standard
 361 error of the mean. Lowercase letters indicate a significant difference ($p < 0.05$) between the mean
 362 concentrations of leachants for each ash sample.



363

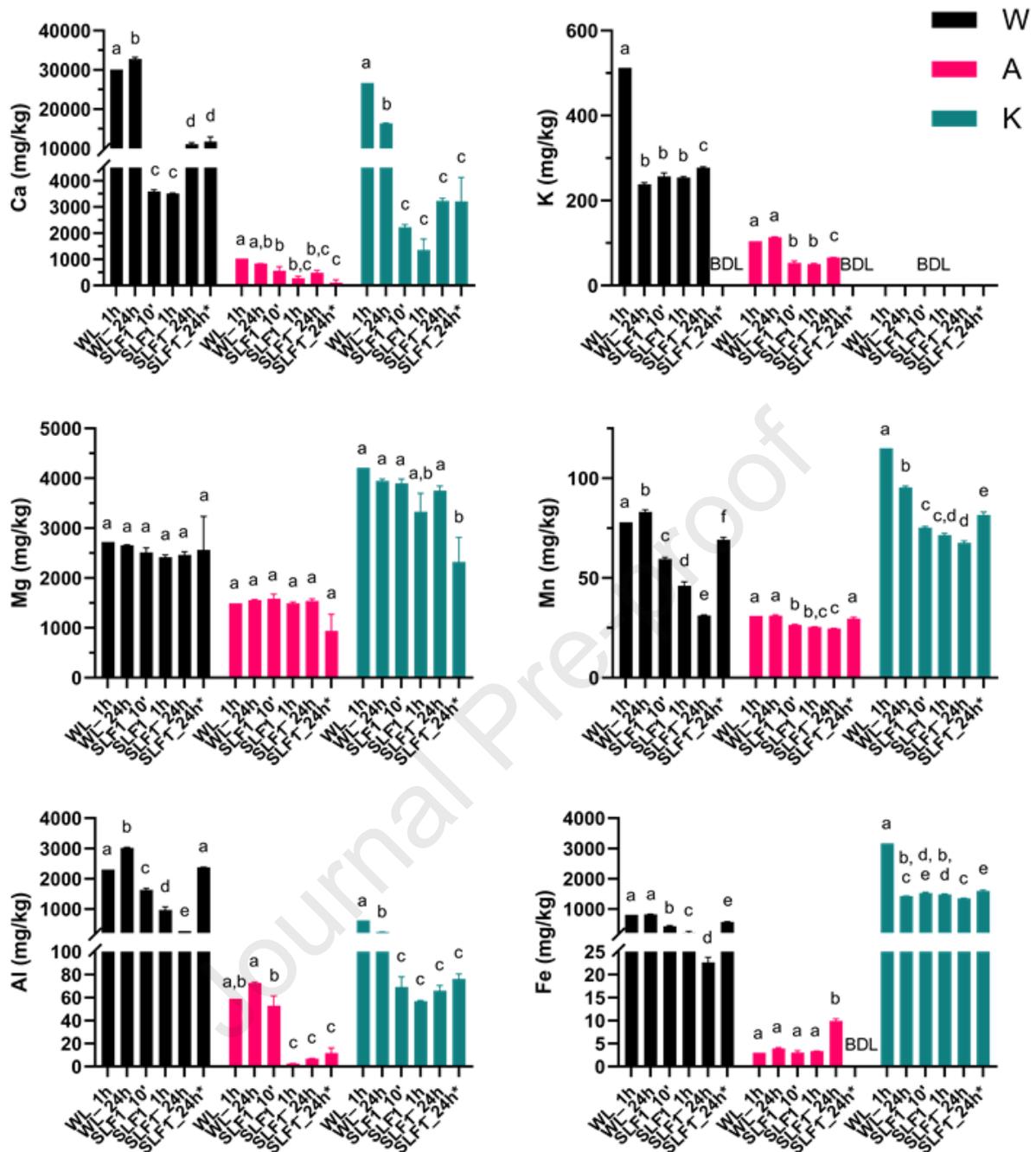
364 **Figure 7** Minor element concentrations (< 5 mg/kg, global mean concentration) in three ash samples
 365 (Whaakari/White Island – black, Ambae – red, Kīlauea – green) obtained through the extractions in
 366 varying ash to leachant (S:L) ratios (24 h in SLF1). Data are reported as mg element per kg of ash dry
 367 weight and represented as the mean of three replicates for each sample. Error bars are the standard
 368 error of the mean. Lowercase letters indicate a significant difference ($p < 0.05$) between the mean
 369 concentrations of leachants for each ash sample.

370 3.4 Comparison of DI water and SLF leach

371 A comparison of the concentrations determined in standard WL after 1 h (from Damby et al. (2018)
372 and Stewart C. (unpublished)) and 24 h, and SLF1 leach for 10 min, 1 h and 24 h (at 1:100 and 1:1000
373 S/L) is shown in **Fig. 8** (major elements) and **Fig. 9** (minor elements).

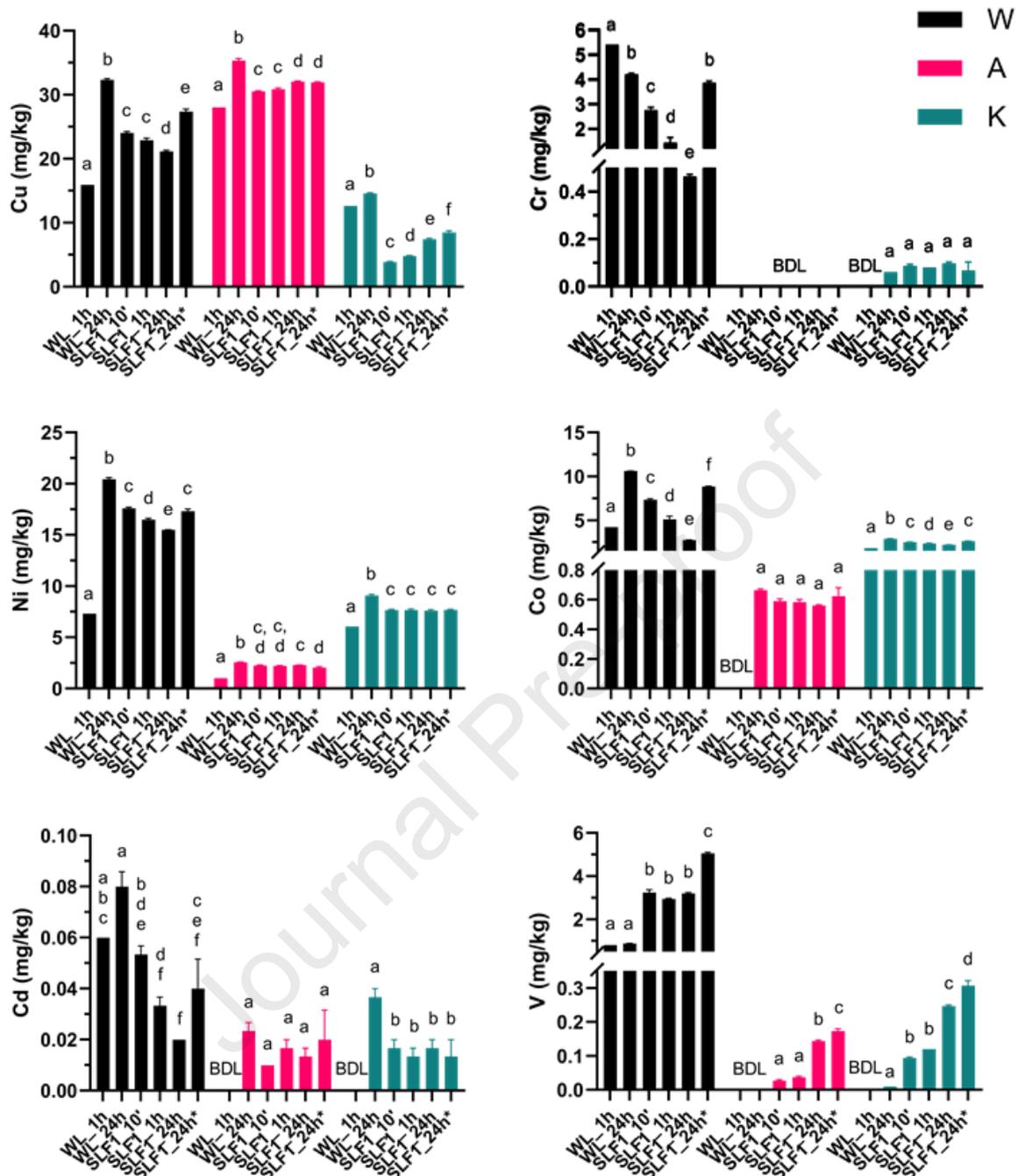
374 The results demonstrate that most of the water-soluble elements in the analysed ash samples were
375 found in similar or higher concentrations at 24 h than those measured in 1 h WL, with the exception
376 of K and Cr from Whaakari/White Island ash, and Al, Ca, Mn and Fe from Kilauea ash, which were
377 higher in the 1 h WL (**Fig. 8**).

378 The overall trends indicated that the SLF1 leached concentrations of major elements were either
379 comparable to, or significantly ($p < 0.05$) lower than, those in the 1 h and 24 h WL, with only Mg
380 concentrations being the same across the compared parameters (**Fig. 8**). The concentrations of minor
381 elements in SLF1 leachates were either comparable to the WL or were higher than the 1 h WL, but
382 lower than the 24 h WL (**Fig. 9**), except for V, which was significantly ($p < 0.05$) lower or BDL in the
383 WL.



384

385 **Figure 8** Major element concentrations (> 5 mg/kg, global mean concentration) in three ash samples
 386 obtained through the extractions in water (WL) and SLF1 over varying time-points. All data are for
 387 S:L 1:100, except SLF1 at 24 h denoted with a star (*), which was extracted at S:L 1:1000. Data are
 388 reported as mg element per kg of ash dry weight and represented as the mean of three replicates for
 389 each sample, except for WL at 1 h (n=1). Error bars are the standard error of the mean. Lowercase
 390 letters indicate a significant difference ($p < 0.05$) between the mean concentrations of leachates for
 391 each ash sample.



392

393 **Figure 9** Minor element concentrations (< 5 mg/kg, global mean concentration) in three ash samples
 394 obtained through the extractions in water (WL) and SLF1 over varying time-points. All data are for
 395 S:L 1:100, except SLF1 at 24 h denoted with a star (*), which was extracted at S:L 1:1000. Data are
 396 reported as mg element per kg of ash dry weight and represented as the mean of three replicates for
 397 each sample, except for WL at 1 h (n=1). Error bars are the standard error of the mean. Lowercase
 398 letters indicate a significant difference ($p < 0.05$) between the mean concentrations of leachates for
 399 each ash sample.

400 **4 Discussion**

401 **4.1 Influence of leaching parameters on release of elements from volcanic ash**

402 The results demonstrate the influence of operational conditions on the measured PTE concentrations
403 in SLF. In most cases, changing the parameters had an equivalent effect across all samples, which
404 allowed for an overview of the direct impact of each method parameter.

405 The affect of specific compounds in the solubilisation of certain elements in SLF was clearly
406 demonstrated. The observed differences in the release of PTEs among the four SLF compositions used
407 (**Fig. 2** and **Fig. 3**) can likely be attributed to differences in their mobilities and tendencies to form
408 soluble complexes with organic components present in the tested SLF solutions (**Table 2**). This was
409 shown for glycine, a component of SLF1-SLF3, which likely formed soluble complexes with Cu (**Fig.**
410 **3**). Similarly, Al, Cr and Fe have high affinities for the citrate ion and were not as efficiently released
411 in SLF3 in the absence of citrate (**Fig. 2** and **Fig. 3**). Furthermore, there was little difference in
412 leachate concentrations between SLF1 and SLF2, suggesting that inclusion of lung surfactant (DPPC)
413 in SLF imparts limited impact on extraction of PTEs associated with volcanic ash (**Fig. 3**). The
414 relevance of including DPPC in an SLF has been repeatedly highlighted in the literature (see review
415 by Kastury et al., 2017), mostly due to it being the dominant component of the lung surfactant by
416 mass and its ability to promote dissolution efficiency. While it was shown that DPPC increases the
417 bioaccessibility of certain elements, *e.g.*, Pb, Zn and Sr (Boisa et al., 2014; Caboche et al., 2011;
418 Pelfrêne et al., 2017), it has also been reported that the addition of DPPC results in no significant
419 changes to bioaccessibility (Pelfrêne et al., 2017).

420 The concentrations of PTEs in the leachates in the time-series experiments (**Fig. 4** and **Fig. 5**) follow
421 three general trajectories up to 24 h, when most PTE concentrations reach a plateau: increasing, flat,
422 or decreasing. Whereas increasing concentrations can be explained by prolonged release, and a stable
423 concentration explained by rapid dissolution, a decrease in concentration requires sequestering of
424 previously leached elements, which decreases the fraction that remains in the solution over time. This

425 decrease for some elements is possibly a consequence of formation of new, insoluble phases and/or
426 adsorption onto the particles. Such formation of new, insoluble complexes or, alternatively, lack of
427 dissociation of soluble complexes, can be facilitated by the neutral pH of the SLF solution (Pelfrêne et
428 al., 2017; Schaider et al., 2007; Marschner et al., 2006). These processes may be also associated with
429 the differences in PTE extraction in varying S:L ratios. The results indicated that PTE concentrations
430 in SLF1 were largely independent of ratio in the range 1:100 to 1:1000 at 24 h (**Fig. 6** and **Fig. 7**),
431 except for Whakaari/White Island ash, where the concentrations were higher in 1:1000 (and 1:500)
432 than 1:100 for some elements (Al, Fe, Cr, Cu, V). It is possible that, in this case, the saturation effect
433 was reduced due to lower particle mass loading in the solution during the extraction.

434 Overall, it could be seen that some element releases in SLF changed with the different substrate
435 (sample), but the observed leaching behaviours could be summarised as follows:

- 436 1) The concentrations of Mg measured in the SLF leachates were the most stable across the
437 different experiments in all samples;
- 438 2) The dissolution of Ni in SLF was the least affected by variation in test parameters among
439 analysed PTEs and samples;
- 440 3) The other elements showed different sensitivities to changes in test parameters and exhibited
441 variability among the three ash samples, with the strongest effects observed for Al, Fe and
442 Mn.

443 Any sample-specific deviations from these broad effects are likely due to the differences in ash
444 leachable burden (**Fig. 1, Table 1**), including the state in which PTEs are bound to the ash particles
445 (Wolf et al., 2011).

446 **4.2 Comparison of SLF and water leach**

447 Our results showed that leachants of similar near-neutral pH with different complexity in their
448 chemical composition affect the release patterns of elements from volcanic ash (**Fig. 8** and **Fig. 9**).
449 The solubilisation potential of SLF solutions was particularly noticeable for V, which increased over

450 time in SLF1 and was measured in all SLF leachates (**Fig. 3**) yet found in significantly ($p < 0.05$)
451 lower concentrations (or was BDL) in the WL at 1 h and 24 h (**Fig. 9**).

452 Comparison of WL and SLF data demonstrates that the initial release of major elements in SLF1 at 10
453 min is, to some extent, comparable to that in WL at 1 h (**Fig. 8**), while minor PTE concentrations are
454 higher in 10 min SLF1 than in 1 h WL, and more similar to those measured in 24 h WL (**Fig. 9**). In
455 some cases, for the 24 h time-point, it seems that the WL overestimates the amounts extractable in
456 SLF, but this apparent lower leaching efficiency of SLF relative to earlier time-points in SLF1 and
457 WL, in general, is likely a consequence of solution reaching saturation state. This may be explained
458 by progressive dissolution of less soluble surface phases and/or dissolution of the ash grains, which, in
459 the case of the WL, does not result in solution saturation, as seen from the differences in PTE
460 concentrations between the 1 h and 24 h WL.

461 In summary, the results showed that:

- 462 1) The concentrations of Mg measured in the leachates (WL and SLF) were the most stable in all
463 samples;
- 464 2) The dissolution of V in SLF was found to be more efficient than in WL regardless of the
465 sample;
- 466 3) The WL reflects the SLF-soluble PTEs for shorter contact times, except for Cu, Co and Ni in
467 some samples.

468 **4.3 SLF analytical challenges**

469 Although we successfully characterised the leaching behaviour of most of the analysed elements in
470 SLF, determining the concentration of Na yielded poor results (Appendix C). These results are
471 probably due to the initial levels of Na in the SLF solutions (**Table 2**), so the measured concentrations
472 are likely to be less reliable since they are generally similar to those in the blanks: WL concentrations
473 for Na from the 3 samples are 8.1-67.3 mg/L (Appendix C), whereas the limits of quantification for
474 Na are approximately 100-500 mg/L (**Table 4**). Pb and Zn were BDL in the SLF leachates, even
475 though they were measured in the WL (**Fig. 1**). The finding of minor elements BDL of the method but

476 present in WL likely results from the solution being too dilute for quantification, or because the
477 concentrations measured in samples are, again, too similar to, or lower than, the blank concentration.

478 We note, though, that WL concentrations of Pb and Zn were comparatively low in our samples
479 relative to other ash samples (**Fig. 1**). Therefore, we cannot comment in detail on whether
480 solubilization of these elements would increase if present in higher concentrations, or whether
481 previously observed increases from other materials were specific to the sample matrix. The mean and
482 median concentrations of soluble elements in ash are generally found at the lower end of their
483 reported ranges (Ayriss and Delmelle, 2012), but it is possible that other ash types might exhibit higher
484 PTE concentrations than those analysed here. The range of PTEs investigated in this study are
485 generally the most abundant in ash (Ayriss and Delmelle, 2012), and are also commonly analysed in
486 lung bioaccessibility studies (Kastury et al., 2017; Plumlee et al., 2003). Other elements usually
487 present in ash leachates, such as As, Ba, Li, Mo, Se, Si, Sr, were not considered here and their
488 leaching behaviour in SLF is yet to be investigated.

489 **4.4 Implications for the ash hazard assessment**

490 Based on our experiments, an overview of possible steps towards operationalizing an SLF method for
491 ash leaching is given below:

- 492 1) **Leachant composition.** While previous SLF studies have excluded use of phospholipids such as
493 DPPC without justification or explanation, here we clearly show that non-inclusion in a modified
494 Gamble's solution (SLF1, **Table 2**) is a reasonable SLF modification when determining
495 bioaccessible concentrations of elements from volcanic ash. More profound effects on extraction
496 efficiency were noted for some of the analysed elements in the absence of citrate (SLF3) or
497 glycine (SLF4), confirming the postulation by Stewart et al. (2020) that omission of these key
498 organic compounds could lead to potential underestimation of the real bioaccessibility of PTEs.
- 499 2) **Extraction time.** While the release of elements is time-dependent, we show that ash-associated
500 PTEs may approximate maximum dissolution in an environment resembling lung lining fluid
501 (**Fig. 5**) in the first 10 minutes of leaching. Therefore, even though ash particles may reside in the

502 lung for extended periods of time, our data suggest that a short contact time (up to 4 h) will
503 adequately estimate the upper limit of PTE release, but a longer extraction period (*e.g.*, 24 h)
504 would be necessary for research considering the steady state of bioaccessible PTEs. While still
505 not too time-consuming for an *in vitro* extraction procedure, 24 h extraction would reflect the
506 availability of more slowly soluble compounds. This would also allow comparison with published
507 SLF dissolution data and acute toxicity data (*e.g.*, Tomašek et al., 2019).

508 3) **S:L ratio.** We show that PTE dissolution in SLF is relatively stable at our lower S:L ratios (1:100
509 through 1:1000), but at ratios 1:500 and 1:1000 the repeatability is lower (Appendix C) and there
510 is increased risk of introducing potential errors due to small abundances of some elements (*e.g.*,
511 Cd, Cr, Pb, Zn), as a consequence of large sample dilution. Therefore, 1:100 could be adopted as
512 an optimal ratio. This ensures sufficient volume of leachant for the subsequent analysis, while
513 using a minimal amount of ash to produce reliable data. In this way, impracticable scaling up with
514 lower ratios, which require large volumes of SLF, is avoided as well.

515 Considering the challenges with the method application outlined in this paper, and for the mentioned
516 practical reasons, the SLF leach may be difficult to implement and include on a routine basis within
517 the standardised IVHHN leachates protocol for rapid hazard assessment (Stewart et al., 2020). Our
518 results showed that the WL largely reflects the SLF soluble element concentrations for shorter contact
519 times. This suggests that the general-purpose WL could be considered a suitable analogue for SLF
520 and used as a conservative estimate of soluble elements, for the purposes of rapid respiratory hazard
521 assessment from leachable elements, while acknowledging that some elements may be underreported
522 (*e.g.*, Cu, Ni, V). The WL is much easier to implement in laboratories, thus offering a practical
523 approach to assessing the potential lung bioaccessible PTEs from ash, especially in time-sensitive
524 situations during volcanic crises. Further, previous rapid ash hazard assessments have used a WL
525 (*e.g.*, Damby et al., 2017, 2013; Horwell et al., 2013), so continued use of a WL allows data to be
526 comparable with past case studies. However, given the observed differences between WL and SLF,
527 SLF leach should still be a preferred method for detailed investigations of PTEs of specific concern to
528 respiratory health outside of a response situation.

529 **5 Conclusions**

530 This study aimed to understand the leaching dynamics of PTEs from volcanic ash to inform a choice
531 of parameters for an *in vitro* protocol to estimate the soluble-element hazard from inhaled ash. The
532 release of PTEs was evaluated in varying formulations of SLF and under varying experimental
533 conditions, and compared to a standardized water leach (Stewart et al., 2020). Our findings show that:

- 534 • Release of elements in SLF is affected by changes in assay parameters, including S:L ratio,
535 extraction time and solution composition;
- 536 • The differences among ash samples are element specific, indicating the role of ash
537 composition in PTE bioaccessibility;
- 538 • The addition of lung surfactant (DPPC) is not necessary when assessing bioaccessible
539 concentrations of elements in volcanic ash;
- 540 • Some major elements (Ca, Na) are less reliably quantified than minor elements, likely due to
541 their initial, high concentrations in SLF;
- 542 • The elements found as the most sensitive to changes in test parameters are Al, Fe and Mn,
543 whereas the least affected were Ni and Mg;
- 544 • SLF is more efficient than WL in extracting V, but also Cu and Ni over shorter time periods
545 (≤ 1 h);
- 546 • A WL may be used as a conservative estimate of lung bioaccessibility in a response situation.

547 This study provides a useful step in the development of a leachate protocol which could form a
548 standard method for volcanic ash respiratory hazard analysis. Future application would allow
549 acquisition of leachate composition data that can be more easily compared to that of other ash
550 characterisation studies, and it will foster the development of a global database of information
551 relevant for informing volcanic health hazard from leachable elements (Stewart et al., 2020).

552 6 Abbreviations

553	BDL	below detection limit
554	DI	deionized water
555	DPPC	dipalmitoylphosphatidylcholine
556	HR-ICP-MS	high resolution inductively coupled plasma mass spectrometry
557	IVHHN	International Volcanic Health Hazard Network
558	LOQ	limit of quantification
559	PTE	potentially toxic element
560	SLF	simulated lung fluid
561	S:L	solid (mass) to liquid (volume) ratio
562	WL	water leach

563 7 Supplementary Material

564	Appendix A: Water-leachable element content of ash samples
565	Appendix B: Instrumental parameters
566	Appendix C: Experimental data

567 **8 Author contributions**

568 **Ines Tomašek:** *Conceptualization, Methodology, Investigation, Formal analysis, Visualization,*
 569 *Writing – original draft.* **David E. Damby:** *Conceptualization, Methodology, Resources, Writing –*
 570 *review & editing.* **Carol Stewart:** *Conceptualization, Resources, Writing – review & editing.* **Claire J.**
 571 **Horwell:** *Conceptualization, Writing – review & editing.* **Geoff Plumlee:** *Conceptualization, Writing*
 572 *– review & editing.* **Pierre Delmelle:** *Conceptualization, Writing – review & editing.* **Christopher J.**
 573 **Ottley:** *Conceptualization, Writing – review & editing.* **Suzette Morman:** *Conceptualization, Writing*
 574 *– review & editing.* **Sofian El Yazidi:** *Investigation.* **Philippe Claeys:** *Funding acquisition, Resources,*
 575 *Supervision, Writing – review & editing.* **Matthieu Kervyn:** *Resources, Supervision, Writing – review*
 576 *& editing.* **Marc Elskens:** *Resources, Supervision, Writing – review & editing.* **Martine Leermakers:**
 577 *Methodology, Investigation, Writing – review & editing.*

578 **9 Declaration of competing interest**

579 The authors declare no conflict of interest.

580 **10 Acknowledgments**

581 This work was funded by the VUB Strategic Research Program granted to PC, ME and MK. IT and
 582 CJH acknowledge the support by the Marie Skłodowska-Curie Actions Initial Training Network
 583 ‘VERTIGO’, funded through the European Seventh Framework Programme (FP7 2007–2013) under
 584 Grant Agreement number 607905. The authors thank Gideon Maleb of Vinanangwe, Ambae, Vanuatu
 585 for providing the bulk 2018 Ambae ash sample, and Mike Rosenberg of GNS Science, New Zealand,
 586 for providing the bulk 2016 Whakaari/White Island ash sample. Thanks to Pierre-Yves Tournigand
 587 (FARD-VUB, Belgium) for graphic design of the manuscript graphical abstract. We are thankful to
 588 three anonymous reviewers for their constructive comments on this manuscript. A USGS internal
 589 review by JoAnn Holloway substantially improved the presentation of content in this manuscript. Any

590 use of trade, firm or product names is for descriptive purposes only and does not imply endorsement
 591 by the U.S. Government.

592 **11 References**

- 593 Ayris, P.M., Delmelle, P., 2012. The immediate environmental effects of tephra emission. *Bull.*
 594 *Volcanol.* 74, 1905–1936. <https://doi.org/10.1007/s00445-012-0654-5>
- 595 Ayris, P.M., Delmelle, P., Pereira, B., Maters, E.C., Damby, D.E., Durant, A.J., Dingwell, D.B., 2015.
 596 Spatial analysis of Mount St. Helens tephra leachate compositions: implications for future
 597 sampling strategies. *Bull. Volcanol.* 77, 1–17.
- 598 Bailey, M.R., Ansoborlo, E., Guilmette, R.A., Paquet, F., 2007. Updating the ICRP human respiratory
 599 tract model. *Radiat. Prot. Dosimetry* 127, 31–34.
- 600 Barone, G., De Giudici, G., Gimeno, D., Lanzafame, G., Podda, F., Cannas, C., Giuffrida, A.,
 601 Barchitta, M., Agodi, A., Mazzoleni, P., 2020. Surface reactivity of Etna volcanic ash and
 602 evaluation of health risks. *Sci. Total Environ.* 143248.
- 603 Boisa, N., Elom, N., Dean, J.R., Deary, M.E., Bird, G., Entwistle, J.A., 2014. Development and
 604 application of an inhalation bioaccessibility method (IBM) for lead in the PM10 size fraction
 605 of soil. *Environ. Int.* 70, 132–142.
- 606 Caboche, J., Perdrix, E., Malet, B., Laurent, A.Y., 2011. Development of an in vitro method to
 607 estimate lung bioaccessibility of metals from atmospheric particles. *J. Environ. Monit.* 13,
 608 621–630.
- 609 Chen, L.C., Lippmann, M., 2009. Effects of metals within ambient air particulate matter (PM) on
 610 human health. *Inhal. Toxicol.* 21, 1–31.
- 611 Colombo, C., Monhemius, A.J., Plant, J.A., 2008. Platinum, palladium and rhodium release from
 612 vehicle exhaust catalysts and road dust exposed to simulated lung fluids. *Ecotoxicol. Environ.*
 613 *Saf.* 71, 722–730.
- 614 Costa, D.L., Dreher, K.L., 1997. Bioavailable transition metals in particulate matter mediate
 615 cardiopulmonary injury in healthy and compromised animal models. *Environ. Health*
 616 *Perspect.* 105, 1053.
- 617 Damby, D., 2012. From dome to disease: The respiratory toxicity of volcanic cristobalite. Durham
 618 University.
- 619 Damby, D.E., Horwell, C.J., Baxter, P.J., Delmelle, P., Donaldson, K., Dunster, C., Fubini, B.,
 620 Murphy, F.A., Natrass, C., Sweeney, S., Tetley, T.D., Tomatis, M., 2013. The respiratory
 621 health hazard of tephra from the 2010 Centennial eruption of Merapi with implications for
 622 occupational mining of deposits. *J. Volcanol. Geotherm. Res.* 261, 376–387.
 623 <https://doi.org/10.1016/j.jvolgeores.2012.09.001>

- 624 Damby, D.E., Horwell, C.J., Larsen, G., Thordarson, T., Tomatis, M., Fubini, B., Donaldson, K.,
625 2017. Assessment of the potential respiratory hazard of volcanic ash from future Icelandic
626 eruptions: a study of archived basaltic to rhyolitic ash samples. *Environ. Health* 16, 98.
- 627 Damby, D.E., Peek, S., Lerner, A.H., Elias, T., 2018. Volcanic ash leachate chemistry from increased
628 2018 activity of Kīlauea Volcano, Hawaii (U.S. Geological Survey data release).
629 <https://doi.org/10.5066/P98A07DC>.
- 630 Dean, J.R., Elom, N.I., Entwistle, J.A., 2017. Use of simulated epithelial lung fluid in assessing the
631 human health risk of Pb in urban street dust. *Sci. Total Environ.* 579, 387–395.
- 632 Duggen, S., Croot, P., Schacht, U., Hoffmann, L., 2007. Subduction zone volcanic ash can fertilize
633 the surface ocean and stimulate phytoplankton growth: Evidence from biogeochemical
634 experiments and satellite data. *Geophys. Res. Lett.* 34.
- 635 Gamble, J., 1967. *Chemical anatomy, physiology and pathology of extracellular fluid*. Harvard
636 University Press, Cambridge.
- 637 Gehr, P., Schürch, S., Berthiaume, Y., HOF, V.I., Geiser, M., 1990. Particle retention in airways by
638 surfactant. *J. Aerosol Med.* 3, 27–43.
- 639 Gray, J.E., Plumlee, G.S., Morman, S.A., Higuera, P.L., Crock, J.G., Lowers, H.A., Witten, M.L.,
640 2010. In vitro studies evaluating leaching of mercury from mine waste calcine using
641 simulated human body fluids. *Environ. Sci. Technol.* 44, 4782–4788.
- 642 Horwell, C.J., 2007. Grain-size analysis of volcanic ash for the rapid assessment of respiratory health
643 hazard. *J. Environ. Monit.* 9, 1107–1115. <https://doi.org/10.1039/010583p>
- 644 Horwell, C.J., Baxter, P.J., 2006. The respiratory health hazards of volcanic ash: a review for volcanic
645 risk mitigation. *Bull. Volcanol.* 69, 1–24. <https://doi.org/10.1007/s00445-006-0052-y>
- 646 Horwell, C.J., Baxter, P.J., Hillman, S.E., Calkins, J.A., Damby, D.E., Delmelle, P., Donaldson, K.,
647 Dunster, C., Fubini, B., Kelly, F.J., Le Blond, J.S., Livi, K.J.T., Murphy, F., Natrass, C.,
648 Sweeney, S., Tetley, T.D., Thordarson, T., Tomatis, M., 2013. Physicochemical and
649 toxicological profiling of ash from the 2010 and 2011 eruptions of Eyjafjallajökull and
650 Grimsvotn volcanoes, Iceland using a rapid respiratory hazard assessment protocol. *Environ.*
651 *Res.* 127, 63–73. <https://doi.org/10.1016/j.envres.2013.08.011>
- 652 ICRP, 1994. Publication 66: Human Respiratory Tract Model for Radiological Protection. A report of
653 a Task Group of the International Commission on Radiological Protection. Elsevier Health
654 Sciences.
- 655 Kastury, F., Smith, E., Juhasz, A.L., 2017. A critical review of approaches and limitations of
656 inhalation bioavailability and bioaccessibility of metal(loid)s from ambient particulate matter
657 or dust. *Sci. Total Environ.* 574, 1054–1074. <https://doi.org/10.1016/j.scitotenv.2016.09.056>
- 658 Macklin, C.C., 1955. Lung fluid, alveolar dust drift, and initial lesions of disease in the lungs. *Can.*
659 *Med. Assoc. J.* 72, 664.

- 660 Marschner, B., Welge, P., Hack, A., Wittsiepe, J., Wilhelm, M., 2006. Comparison of soil Pb in vitro
661 bioaccessibility and in vivo bioavailability with Pb pools from a sequential soil extraction.
662 *Environ. Sci. Technol.* 40, 2812–2818.
- 663 Martin, R., Dowling, K., Nankervis, S., Pearce, D., Florentine, S., McKnight, S., 2018. In vitro
664 assessment of arsenic mobility in historical mine waste dust using simulated lung fluid.
665 *Environ. Geochem. Health* 40, 1037–1049.
- 666 Mayer, K., Scheu, B., Gilg, H.A., Heap, M.J., Kennedy, B.M., Lavallée, Y., Letham-Brake, M.,
667 Dingwell, D.B., 2015. Experimental constraints on phreatic eruption processes at Whakaari
668 (White Island volcano). *J. Volcanol. Geotherm. Res.* 302, 150–162.
- 669 Misra, S.K., Dybowska, A., Berhanu, D., Luoma, S.N., Valsami-Jones, E., 2012. The complexity of
670 nanoparticle dissolution and its importance in nanotoxicological studies. *Sci. Total Environ.*
671 438, 225–232.
- 672 Morgan, L., Pearson, M., De Jongh, R., Mackey, D., van der Wall, H., Peters, M., Rutland, J., 2004.
673 Scintigraphic measurement of tracheal mucus velocity in vivo. *Eur. Respir. J.* 23, 518–522.
- 674 Moussallam, Y., Rose-Koga, E.F., Koga, K.T., Médard, E., Bani, P., Devidal, J.-L., Tari, D., 2019.
675 Fast ascent rate during the 2017–2018 Plinian eruption of Ambae (Aoba) volcano: a
676 petrological investigation. *Contrib. Mineral. Petrol.* 174, 90.
- 677 Neal, C.A., Brantley, S.R., Antolik, L., Babb, J.L., Burgess, M., Calles, K., Cappos, M., Chang, J.C.,
678 Conway, S., Desmither, L., 2019. The 2018 rift eruption and summit collapse of Kīlauea
679 Volcano. *Science* 363, 367–374.
- 680 Oberdörster, G., 2000. Determinants of the pathogenicity of man-made vitreous fibers (MMVF). *Int.*
681 *Arch. Occup. Environ. Health* 73, S60–S68.
- 682 Oller, A.R., Cappellini, D., Henderson, R.G., Bates, H.K., 2009. Comparison of nickel release in
683 solutions used for the identification of water-soluble nickel exposures and in synthetic lung
684 fluids. *J. Environ. Monit.* 11, 823–829.
- 685 Olsson, J., Stipp, S.L.S., Dalby, K.N., Gislason, S.R., 2013. Rapid release of metal salts and nutrients
686 from the 2011 Grímsvötn, Iceland volcanic ash. *Geochim. Cosmochim. Acta* 123, 134–149.
- 687 Pelfrêne, A., Cave, M.R., Wragg, J., Douay, F., 2017. In vitro investigations of human
688 bioaccessibility from reference materials using simulated lung fluids. *Int. J. Environ. Res.*
689 *Public. Health* 14, 112.
- 690 Plumlee, G.S., Morman, S.A., 2011. Mine wastes and human health. *Elements* 7, 399–404.
- 691 Plumlee, G.S., Ziegler, T.L., Lollar, B.S., 2003. The medical geochemistry of dusts, soils, and other
692 earth materials. *Environ. Geochem.* 9, 263–310.
- 693 Schaidler, L.A., Senn, D.B., Brabander, D.J., McCarthy, K.D., Shine, J.P., 2007. Characterization of
694 zinc, lead, and cadmium in mine waste: Implications for transport, exposure, and
695 bioavailability. *Environ. Sci. Technol.* 41, 4164–4171.

- 696 Searl, A., Nicholl, A., Baxter, P.J., 2002. Assessment of the exposure of islanders to ash from the
697 Soufriere Hills volcano, Montserrat, British West Indies. *Occup. Environ. Med.* 59, 523–531.
- 698 Stewart, C., Damby, D.E., Tomašek, I., Horwell, C.J., Plumlee, G.S., Armienta, M.A., Hinojosa,
699 M.G.R., Appleby, M., Delmelle, P., Cronin, S., 2020. Assessment of leachable elements in
700 volcanic ashfall: A review and evaluation of a standardized protocol for ash hazard
701 characterization. *J. Volcanol. Geotherm. Res.* 106756.
- 702 Stewart, C., Horwell, C., Plumlee, G., Cronin, S., Delmelle, P., Baxter, P., Calkins, J., Damby, D.,
703 Morman, S., Oppenheimer, C., 2013. Protocol for analysis of volcanic ash samples for
704 assessment of hazards from leachable elements. Available at:
705 https://www.ivhhn.org/images/pdf/volcanic_ash_leachate_protocols.pdf.
- 706 Studer, A.M., Limbach, L.K., Van Duc, L., Krumeich, F., Athanassiou, E.K., Gerber, L.C., Moch, H.,
707 Stark, W.J., 2010. Nanoparticle cytotoxicity depends on intracellular solubility: comparison
708 of stabilized copper metal and degradable copper oxide nanoparticles. *Toxicol. Lett.* 197,
709 169–174.
- 710 Tomašek, I., Damby, D.E., Horwell, C.J., Ayris, P.M., Delmelle, P., Ottley, C.J., Cubillas, P., Casas,
711 A.S., Bisig, C., Petri-Fink, A., 2019. Assessment of the potential for in-plume sulphur dioxide
712 gas-ash interactions to influence the respiratory toxicity of volcanic ash. *Environ. Res.* 179,
713 108798.
- 714 Twining, J., McGlenn, P., Loi, E., Smith, K., Gieré, R., 2005. Risk ranking of bioaccessible metals
715 from fly ash dissolved in simulated lung and gut fluids. *Environ. Sci. Technol.* 39, 7749–
716 7756.
- 717 Utembe, W., Potgieter, K., Stefaniak, A.B., Gulumian, M., 2015. Dissolution and biodurability:
718 Important parameters needed for risk assessment of nanomaterials. *Part. Fibre Toxicol.* 12,
719 11.
- 720 Wallenborn, J.G., Schladweiler, M.J., Richards, J.H., Kodavanti, U.P., 2009. Differential pulmonary
721 and cardiac effects of pulmonary exposure to a panel of particulate matter-associated metals.
722 *Toxicol. Appl. Pharmacol.* 241, 71–80.
- 723 Wiseman, C.L., Zereini, F., 2014. Characterizing metal (loid) solubility in airborne PM10, PM2.5
724 and PM1 in Frankfurt, Germany using simulated lung fluids. *Atmos. Environ.* 89, 282–289.
- 725 Witham, C.S., Oppenheimer, C., Horwell, C.J., 2005. Volcanic ash-leachates: a review and
726 recommendations for sampling methods. *J. Volcanol. Geotherm. Res.* 141, 299–326.
727 <https://doi.org/10.1016/j.jvolgeores.2004.11.010>
- 728 Wolf, R.E., Morman, S.A., Hageman, P.L., Hoefen, T.M., Plumlee, G.S., 2011. Simultaneous
729 speciation of arsenic, selenium, and chromium: species stability, sample preservation, and
730 analysis of ash and soil leachates. *Anal. Bioanal. Chem.* 401, 2733.
- 731

Table A.1 Water-extractable element concentrations (in $\mu\text{g}/\text{kg}$ or mg/kg dry weight ash) from samples used in the study, determined by water leach for 1 h at 1:100 solid to liquid ratio (Damby et al., 2018; Stewart C., unpublished) and the mean concentrations reported in the global dataset on water-extractable elements from volcanic ash (Ayris and Delmelle, 2012).

	Whaakari/ White Island ¹	Ambae ¹	Kilauea ²	Global mean ³
$\mu\text{g}/\text{kg}$				
Cd	0.06	BDL	BDL	57
Co	4.2	BDL	1.81	204
Cr	5.4	BDL	BDL	100
Ni	7.3	1	6.06	516
Pb	0.1	BDL	BDL	139
V	0.8	BDL	BDL	91
Zn	9.3	2.7	BDL	4013
mg/kg				
Al	2297	59	620	63
Ca	30062	1022	19900	2172
Cu	16	28	13	6
Fe	809	3	3170	24
K	512	104	148	76
Mg	2721	1493	4210	349
Mn	78	31	115	22
Na	6254	808	972	407

¹Stewart C. (unpublished data); ²Damby et al. (2018); ³Ayris and Delmelle (2012).
BDL = below detection limit.

Table B.1 Instrumental parameters of the ICP-MS used for trace element analysis.

Instrument	ELEMENT2 Thermo Finnigan
Forward power	1,350 W
Reflected power	< 2 W
Nebuliser	Concentric
Solution uptake rate	0.4 mL min ⁻¹ (pumped)
Spray chamber	Cyclonic
Sampling and skimmer cones	Ni (Thermo Finnigan)
Sample gas flow	1 to 1.5 L min ⁻¹
Cool argon flow rate	16 L min ⁻¹
Auxiliary argon flow rate	1.0 L min ⁻¹
Torch	Capacitive decoupling Pt shield torch
RF frequency	27.12 Mhz
Sensitivity	1 x 10 ⁶ cps per 1 ng mL ⁻¹ ¹¹⁵ In (in Low Resolution)
Autosampler	ESI SC 3 Fast
Take-up time	15 s
Wash time	10 s
Number of acquisition	6 (3 runs and 2 pass)
Isotopes in LR	Cd111, Cd114, Pb208 Rh103(IS)
Mass window LR	150% for each isotope
Search window LR	150% for each isotope
Samples per peak LR	20
Integration window LR	80% for each isotope
Isotopes in MR	Al27, V51, Cr52, Mn55, Fe56, Co59, Ni60, Cu63, Zn66 Rh103(IS)
Mass window MR	125%
Search window MR	80%
Samples per peak MR	20
Integration window MR	80%
Scan type	E scan for each isotope
Integration type	Average for each isotope

Table B.2 Instrumental parameters of the ICP-MS used for major cations analysis.

Instrument	ELEMENT2 Thermo Finnigan
Forward power	1,350 W
Reflected power	< 2 W
Nebuliser	Concentric
Solution uptake rate	0.4 mL min ⁻¹ (pumped)
Spray chamber	Cyclonic
Sampling and skimmer cones	Ni (Thermo Finnigan)
Sample gas flow	1 to 1.5 L min ⁻¹
Cool argon flow rate	16 L min ⁻¹
Auxiliary argon flow rate	1.0 L min ⁻¹
Torch	Capacitive decoupling Pt shield torch
RF frequency	27.12 Mhz
Sensitivity	1 x 10 ⁶ cps per 1 ng mL ⁻¹ ¹¹⁵ In (in Low Resolution)
Autosampler	ESI SC 3 Fast
Take-up time	15 s
Wash time	10 s
Number of acquisition	6 (3 runs and 2 pass)
Isotopes in HR	Na23, Mg26, K39, Ca44 Rh103(IS)
Mass window HR	125%
Search window HR	60%
Samples per peak HR	20
Integration window HR	60%
Scan type	E scan for each isotope
Integration type	Average for each isotope

Mg	White Island									Ambae									Kilauea								
	Rep 1	Rep 2	Rep 3	Mean	SD	RSD %	Rep 1	Rep 2	Rep 3	Mean	SD	RSD %	Rep 1	Rep 2	Rep 3	Mean	SD	RSD %	Rep 1	Rep 2	Rep 3	Mean	SD	RSD %			
1:10	2031	2014	2110	2039	120	5.90	1970	1931	1930	1934	22	1.42	3000	3070	3000	3100	190	6.21									
1:20	2000	2010	2000	2000	71	3.50	1950	1921	1930	1927	01	0.07	2950	3070	3070	3070	100	3.20									
1:500	1001	1000	1000	1000	741	73.10	1210	1200	1200	1200	02	0.07	3000	3000	3000	3000	100	3.30									
1:1000	1011	1000	1011	1000	1101	109.00	1402	1310	1300	300	30.10	01.30	2304	3002	1070	2022	601	30.04									
WL (24h)	2010	2010	2010	2010	27	0.30	1970	1900	1977	1900	21	1.07	3000	3000	3000	3000	100	3.30									
SLF1 1:100 24h	2001	2000	2010	2004	110	5.51	1922	1901	1912	1900	10	0.70	3000	3000	3000	3000	100	3.20									
SLF2	2040	1000	1000	1000	410	41.00	1300	1300	1300	1300	04	0.30	2300	3000	2400	2111	300	12.20									
SLF3	2000	2010	2001	2001	00	0.00	1911	1900	1900	1910	00	0.00	3100	3010	3110	3100	100	3.20									
SLF4	2000	2000	2000	2000	00	0.00	1900	1900	1900	1900	00	0.00	3000	3000	3000	3000	100	3.20									
10 min	2000	2000	2000	2000	100	5.01	1900	1900	1900	1900	101	3.32	3000	3000	3100	3000	100	3.30									
1 h	2000	2000	2000	2000	10	0.50	1910	1900	1910	1900	00	0.00	3000	3000	3000	3000	100	3.20									
4 h	2000	2000	2000	2000	100	5.00	1910	1910	1910	1900	00	0.00	3000	3000	3000	3000	100	3.20									
48 h	2000	2000	2000	2000	144	7.20	1910	1900	1910	1900	00	0.00	3100	3000	3000	3000	100	3.00									

Between parameter means, SD, RSD

Mg S:L	Wahaakari			Ambae			Kilauea		
	Mean	SD	RSD %	Mean	SD	RSD %	Mean	SD	RSD %
Leachant	2525	145	5.75	1359	263	19.32	3403	633	18.59
Time	2392	311	13.02	1424	371	26.06	3592	496	13.81
	2447	66	2.71	1551	56	3.59	3698	235	6.36

Cd	White Island									Ambae									Kilauea								
	Rep 1	Rep 2	Rep 3	Mean	SD	RSD %	Rep 1	Rep 2	Rep 3	Mean	SD	RSD %	Rep 1	Rep 2	Rep 3	Mean	SD	RSD %	Rep 1	Rep 2	Rep 3	Mean	SD	RSD %			
1:10	3000	3000	3000	3000	000	0.00	3000	3000	3000	3000	000	0.00	3000	3000	3000	3000	000	0.00									
1:20	10000	10000	10000	10000	200	2.00	3000	3000	3000	3000	000	0.00	3000	3000	3000	3000	100	3.30									
1:500	8017	8029	8013	8020	500	6.20	BDL	BDL	BDL	BDL	--	--	4104	4220	2130	3702	538	12.63									
1:1000	3000	10011	11000	10000	2021	20.21	3000	BDL	BDL	3000	100	10.00	3000	3000	2021	3210	1010	33.00									
WL (24h)	3000	3000	3000	3000	111	3.71	3000	3000	3000	3000	100	10.00	1000	1000	1000	1000	100	10.00									
SLF1 1:100 24h	10000	10000	11011	11000	021	0.21	3000	3000	3000	3000	100	10.00	3000	3000	3000	3000	100	10.00									
SLF2	11000	10100	10000	10000	3000	30.00	1000	BDL	BDL	BDL	--	--	1000	2000	1100	2010	300	15.00									
SLF3	11000	10100	10000	10000	400	4.00	1100	310	000	000	10	0.30	3100	3000	3000	3000	100	3.00									
SLF4	11000	11000	11011	11000	00	0.00	3000	3000	3000	3000	100	10.00	3000	3000	3000	3000	100	10.00									
10 min	3000	3000	3100	3000	100	3.00	3000	3000	3000	3000	100	10.00	3000	3000	3000	3000	100	10.00									
1 h	3000	3000	3000	3000	21	0.70	3000	3000	3000	3000	100	10.00	3000	3000	3000	3000	100	10.00									
4 h	3000	3000	3000	3000	100	3.00	3000	3000	3000	3000	100	10.00	3000	3000	3000	3000	100	10.00									
48 h	12000	11000	11000	12100	200	1.60	3000	3000	3000	3000	100	10.00	3000	3000	3000	3000	100	10.00									

Cd S:L	Wahaakari			Ambae			Kilauea		
	Mean	SD	RSD %	Mean	SD	RSD %	Mean	SD	RSD %
Leachant	11000	1020	9.20	3000	00	0.00	3000	120	4.00
Time	10000	3010	30.10	1100	121	11.00	3000	3010	10.00
	10000	7000	70.00	3000	100	33.30	3000	200	6.60

N	White Island									Ambae									Kilauea								
	Rep 1	Rep 2	Rep 3	Mean	SD	RSD %	Rep 1	Rep 2	Rep 3	Mean	SD	RSD %	Rep 1	Rep 2	Rep 3	Mean	SD	RSD %	Rep 1	Rep 2	Rep 3	Mean	SD	RSD %			
1:10	214	221	200	212	10	4.70	02	10	10	00	2	2.00	0	0	0	0	0	0.00									
1:20	200	201	200	200	0	0.00	00	10	10	00	0	0.00	12	0	0	0	0	0.00									
1:500	00	100	102	101	10	10.00	BDL	BDL	BDL	--	--	--	BDL	BDL	BDL	--	--	--									
1:1000	BDL	BDL	BDL	--	--	--	BDL	BDL	BDL	--	--	--	BDL	BDL	BDL	--	--	--									
WL (24h)	200	200	200	200	1	0.50	110	112	110	110	2	1.80	BDL	BDL	BDL	--	--	--									
SLF1 1:100 24h	212	211	202	208	1	0.50	00	00	00	00	1	1.00	BDL	BDL	BDL	--	--	--									
SLF2	300	200	200	210	20	6.70	00	10	00	00	1	1.00	BDL	BDL	BDL	--	--	--									
SLF3	300	310	310	307	0	0.00	101	111	110	110	0	0.00	22	27	20	20	1	5.00									
SLF4	210	212	202	210	0	0.00	10	02	10	10	0	0.00	BDL	BDL	BDL	--	--	--									
10 min	241	200	200	214	14	6.40	44	01	04	00	0	0.00	BDL	BDL	BDL	--	--	--									
1 h	200	201	200	200	7	3.50	00	00	00	00	1	1.00	BDL	BDL	BDL	--	--	--									
4 h	200	201	200	200	7	3.50	00	00	00	00	1	1.00	BDL	BDL	BDL	--	--	--									
48 h	201	204	200	202	10	5.00	12	01	10	11	0	0.00	BDL	BDL	BDL	--	--	--									

N S:L	Wahaakari			Ambae			Kilauea		
	Mean	SD	RSD %	Mean	SD	RSD %	Mean	SD	RSD %
Leachant	210	10	4.70	10	00	0.00	0	0	0.00
Time	211	20	9.40	00	20	20.00	20	--	--
	200	10	5.00	00	0	0.00	--	--	--

Nd	White Island									Ambae									Kilauea								
	Rep 1	Rep 2	Rep 3	Mean	SD	RSD %	Rep 1	Rep 2	Rep 3	Mean	SD	RSD %	Rep 1	Rep 2	Rep 3	Mean	SD	RSD %	Rep 1	Rep 2	Rep 3	Mean	SD	RSD %			
1:10	0000	0000	10101	10001	2200	22.00	2101	1000	1000	1121	300	26.00	1101	410	100	91.00											
1:20	0000	0000	0000	0000	010	1.00	2122	000	0000	2010	1901	19.00	1900	00	BDL	1900	3010	190.10									
1:500	BDL	110101	00	30401	00014	14.14	BDL	40000	BDL	40000	--	--	BDL	11011	BDL	11011	--	--									
1:1000	BDL	210002	BDL	210002	--	--	10000	BDL	11000	00000	10.00	10.00	BDL	00000	BDL	00000	--	--									
WL (24h)	6533	6662	7001	6732	242	3.59	1015	1007	826	949	107	11.26	787	823	812	807	18	2.29									
SLF1 1:100 24h	1252	18519	19970	13247	10413	78.61	11635	1917	3462	5671	5222	92.08	BDL	11360	8208	9784	2229	22.78									
SLF2	BDL	BDL	BDL	--	--	--	BDL	BDL	BDL	--	--	--	BDL	BDL	BDL	--	--	--									
SLF3	11339	14135	18542	14672	3631	24.75	9334	15474	9565	11458	3480	30.37	8697	11339	8078	9371	1732	18.48									
SLF4	10404	1265	9199	4965	4965	71.38	BDL	3038	BDL	3038	--	--	BDL	BDL	269	269	--	--									
10 min	BDL	2941	8979	5960	4270	71.64	BDL	18810	8175	13493	7520	55.74	11132	BDL	BDL	11132	--	--									
1 h	11327	6907	4005	7413	3687	49.74	1790	7314	BDL	4552	3906	85.81	BDL	BDL	BDL	--	--	--									
4 h	6342	14786	14721	11950	4856	40.64	5612	BDL	1586	3599	2847	79.10	15296	7638	8864	10599	4113	38.81									
48 h	12239	22207	BDL	17223	7048	40.92	21081	13353	26644	20359	6675	32.79	3707														

Al	White Island						Ambae						Kilauea					
	Rep 1	Rep 2	Rep 3	Mean	SD	RSD %	Rep 1	Rep 2	Rep 3	Mean	SD	RSD %	Rep 1	Rep 2	Rep 3	Mean	SD	RSD %
1:10	191.4	229.1	180.0	200.2	25.7	12.85	1.3	1.2	1.2	1.2	0.1	4.56	1.8	1.6	1.7	1.7	0.1	5.92
1:20	14.7	16.9	18.8	16.8	2.0	12.17	2.3	2.0	2.2	2.2	0.2	7.12	9.7	21.1	20.2	17.0	6.4	37.35
1:500	2232.8	2182.4	2191.6	2202.3	26.8	1.22	312.12*	7.0	4.3	5.7	1.9	33.52	148.7	214.3	70.3	144.4	72.1	49.90
1:1000	2387.3	2351.2	2377.3	2372.0	18.6	0.79	17.5	14.8	3.1	11.8	7.7	65.12	80.3	80.9	67.9	76.4	7.3	9.60
WL (24h)	2976.1	3034.3	3043.4	3017.9	36.5	1.21	73.2	74.1	71.1	72.8	1.5	2.10	222.3	235.1	230.4	229.2	6.5	2.82
SLF1 1:100 24h	263.4	263.0	268.6	265.0	3.1	1.17	6.3	7.1	7.2	6.9	0.5	7.53	75.1	61.4	61.8	66.1	7.8	11.83
SLF2	324.7	325.8	340.6	330.4	8.9	2.69	20.5	11.5	12.1	14.7	5.0	34.04	52.1	57.0	55.7	54.9	2.5	4.55
SLF3	0.9	0.6	1.2	0.9	0.3	29.41	BDL	BDL	BDL	--	--	--	BDL	BDL	BDL	--	--	--
SLF4	332.5	316.3	335.3	328.1	10.2	3.12	BDL	BDL	BDL	--	--	--	74.1	73.9	78.2	75.4	2.4	3.19
10 min	1546.6	1721.1	1622.6	1630.1	87.5	5.37	69.1	41.1	48.6	52.9	14.5	27.36	60.8	86.9	60.2	69.3	15.3	22.05
1 h	1159.6	793.2	933.0	961.9	184.9	19.22	2.8	2.3	2.6	2.6	0.3	10.10	58.3	56.6	55.9	56.9	1.2	2.16
4 h	332.6	362.2	367.4	354.1	18.7	5.29	2.8	5.3	2.3	3.4	1.6	46.86	54.9	55.4	54.0	54.7	0.7	1.29
48 h	254.4	263.2	256.2	257.9	4.6	1.80	9.9	14.4	35.7	20.0	13.8	68.75	50.9	60.2	107.0	72.7	30.1	41.36

Al	Whaakari			Ambae			Kilauea		
	Mean	SD	RSD %	Mean	SD	RSD %	Mean	SD	RSD %
S:L	1011.2	1169.8	115.68	5.5	4.2	75.94	61.1	56.3	92.09
Leachant	Mean	SD	RSD %	Mean	SD	RSD %	Mean	SD	RSD %
	788.5	1253.7	159.00	31.4	35.2	112.10	106.4	82.3	77.34
Time	Mean	SD	RSD %	Mean	SD	RSD %	Mean	SD	RSD %
	693.8	599.6	86.42	17.2	21.2	123.45	64.0	7.8	12.21

Cr	White Island						Ambae						Kilauea					
	Rep 1	Rep 2	Rep 3	Mean	SD	RSD %	Rep 1	Rep 2	Rep 3	Mean	SD	RSD %	Rep 1	Rep 2	Rep 3	Mean	SD	RSD %
1:10	0.05	0.06	0.05	0.05	0.01	10.83	BDL	BDL	BDL	--	--	--	BDL	BDL	BDL	--	--	--
1:20	0.02	0.02	0.02	0.02	0.00	0.00	BDL	BDL	BDL	--	--	--	0.01	0.01	0.02	0.01	0.01	43.30
1:500	3.69	3.55	3.79	3.68	0.12	3.28	BDL	BDL	BDL	--	--	--	0.09	0.14	0.07	0.10	0.04	36.06
1:1000	3.99	3.79	3.80	3.86	0.11	2.92	BDL	BDL	BDL	--	--	--	0.03	0.14	0.03	0.07	0.06	95.26
WL (24h)	4.17	4.15	4.29	4.20	0.08	1.80	BDL	BDL	BDL	--	--	--	0.06	0.06	0.06	0.06	0.00	0.00
SLF1 1:100 24h	0.45	0.46	0.48	0.46	0.02	3.30	BDL	BDL	BDL	--	--	--	0.09	0.11	0.09	0.10	0.01	11.95
SLF2	0.55	0.56	0.54	0.55	0.01	1.82	BDL	BDL	BDL	--	--	--	0.07	0.07	0.07	0.07	0.00	0.00
SLF3	BDL	BDL	BDL	--	--	--	BDL	BDL	BDL	--	--	--	BDL	BDL	BDL	--	--	--
SLF4	0.55	0.50	0.55	0.53	0.03	5.41	BDL	BDL	BDL	--	--	--	0.13	0.14	0.14	0.14	0.01	4.22
10 min	2.58	2.98	2.71	2.76	0.20	7.40	BDL	BDL	BDL	--	--	--	0.08	0.10	0.08	0.09	0.01	13.32
1 h	1.84	1.10	1.37	1.44	0.37	26.07	BDL	BDL	BDL	--	--	--	0.08	0.08	0.08	0.08	0.00	0.00
4 h	0.29	0.33	0.33	0.32	0.02	7.29	BDL	BDL	BDL	--	--	--	0.09	0.07	0.08	0.08	0.01	12.50
48 h	0.58	0.59	0.62	0.60	0.02	3.49	BDL	BDL	BDL	--	--	--	0.06	0.07	0.09	0.07	0.02	20.83

Cr	Whaakari			Ambae			Kilauea		
	Mean	SD	RSD %	Mean	SD	RSD %	Mean	SD	RSD %
S:L	1.61	1.97	122.31	--	--	--	0.07	0.04	58.02
Leachant	Mean <td>SD <td>RSD %</td> <td>Mean <td>SD <td>RSD %</td> <td>Mean <td>SD <td>RSD %</td> </td></td></td></td></td>	SD <td>RSD %</td> <td>Mean <td>SD <td>RSD %</td> <td>Mean <td>SD <td>RSD %</td> </td></td></td></td>	RSD %	Mean <td>SD <td>RSD %</td> <td>Mean <td>SD <td>RSD %</td> </td></td></td>	SD <td>RSD %</td> <td>Mean <td>SD <td>RSD %</td> </td></td>	RSD %	Mean <td>SD <td>RSD %</td> </td>	SD <td>RSD %</td>	RSD %
	1.44	1.84	128.30	--	--	--	0.09	0.03	37.71
Time	Mean <td>SD <td>RSD %</td> <td>Mean <td>SD <td>RSD %</td> <td>Mean <td>SD <td>RSD %</td> </td></td></td></td></td>	SD <td>RSD %</td> <td>Mean <td>SD <td>RSD %</td> <td>Mean <td>SD <td>RSD %</td> </td></td></td></td>	RSD %	Mean <td>SD <td>RSD %</td> <td>Mean <td>SD <td>RSD %</td> </td></td></td>	SD <td>RSD %</td> <td>Mean <td>SD <td>RSD %</td> </td></td>	RSD %	Mean <td>SD <td>RSD %</td> </td>	SD <td>RSD %</td>	RSD %
	1.11	1.02	91.20	--	--	--	0.08	0.01	10.58

Fe	White Island						Ambae						Kilauea					
	Rep 1	Rep 2	Rep 3	Mean	SD	RSD %	Rep 1	Rep 2	Rep 3	Mean	SD	RSD %	Rep 1	Rep 2	Rep 3	Mean	SD	RSD %
1:10	18.3	24.4	17.5	20.0	3.8	18.78	0.5	0.6	0.5	0.5	0.0	7.98	10.7	10.3	10.3	10.4	0.2	2.17
1:20	9.0	11.1	11.5	10.5	1.3	12.52	1.1	1.1	1.2	1.1	0.0	1.77	104.3	106.5	109.7	106.8	2.7	2.56
1:500	530.0	500.6	552.8	527.8	26.2	4.95	14.8	13.4	12.0	13.4	1.4	10.44	1600.0	1610.4	1585.2	1598.5	12.6	0.79
1:1000	566.9	552.2	584.9	568.0	16.4	2.88	BDL	BDL	BDL	--	--	--	1559.1	1658.9	1559.3	1592.4	57.6	3.62
WL (24h)	807.5	817.5	830.8	818.6	11.7	1.43	4.3	3.9	3.7	4.0	0.3	7.86	1414.5	1438.5	1406.2	1419.7	16.8	1.18
SLF1 1:100 24h	22.6	20.8	24.6	22.7	1.9	8.53	9.0	10.5	10.2	9.9	0.8	7.95	1364.7	1359.1	1307.1	1343.6	31.8	2.36
SLF2	20.2	20.1	21.6	20.6	0.8	3.89	7.9	8.4	7.9	8.0	0.3	3.27	1086.7	1142.8	1104.6	1111.4	28.7	2.58
SLF3	BDL	BDL	BDL	--	--	--	BDL	BDL	BDL	--	--	--	BDL	BDL	BDL	--	--	--
SLF4	24.7	23.7	25.2	24.5	0.7	3.05	12.3	12.1	12.7	12.4	0.3	2.51	1459.0	1464.8	1473.2	1465.7	7.2	0.49
10 min	379.2	467.5	419.9	422.2	44.2	10.47	3.8	2.8	2.5	3.0	0.7	23.23	1485.4	1565.4	1532.8	1527.9	40.2	2.63
1 h	282.6	147.6	197.9	209.3	68.2	32.59	3.5	3.3	3.4	3.4	0.1	2.99	1505.4	1437.5	1488.2	1477.0	35.3	2.39
4 h	31.3	35.8	34.4	33.8	2.3	6.76	4.3	4.1	4.0	4.1	0.1	3.49	1425.9	1399.7	1399.8	1408.5	15.1	1.07
48 h	23.3	24.3	23.2	23.6	0.6	2.58	13.9	13.5	13.1	13.5	0.4	2.82	987.8	822.8	941.5	917.4	85.1	9.28

Fe	Whaakari			Ambae			Kilauea		
	Mean	SD	RSD %	Mean	SD	RSD %	Mean	SD	RSD %
S:L	229.8	290.8	126.52	6.2	6.4	102.97	930.4	803.1	86.32
Leachant	Mean <td>SD <td>RSD %</td> <td>Mean <td>SD <td>RSD %</td> <td>Mean <td>SD <td>RSD %</td> </td></td></td></td></td>	SD <td>RSD %</td> <td>Mean <td>SD <td>RSD %</td> <td>Mean <td>SD <td>RSD %</td> </td></td></td></td>	RSD %	Mean <td>SD <td>RSD %</td> <td>Mean <td>SD <td>RSD %</td> </td></td></td>	SD <td>RSD %</td> <td>Mean <td>SD <td>RSD %</td> </td></td>	RSD %	Mean <td>SD <td>RSD %</td> </td>	SD <td>RSD %</td>	RSD %
	221.6	398.0	179.60	8.6	3.3	38.78	1335.1	157.4	11.79
Time	Mean <td>SD <td>RSD %</td> <td>Mean <td>SD <td>RSD %</td> <td>Mean <td>SD <td>RSD %</td> </td></td></td></td></td>	SD <td>RSD %</td> <td>Mean <td>SD <td>RSD %</td> <td>Mean <td>SD <td>RSD %</td> </td></td></td></td>	RSD %	Mean <td>SD <td>RSD %</td> <td>Mean <td>SD <td>RSD %</td> </td></td></td>	SD <td>RSD %</td> <td>Mean <td>SD <td>RSD %</td> </td></td>	RSD %	Mean <td>SD <td>RSD %</td> </td>	SD <td>RSD %</td>	RSD %
	142.3	175.4	123.20	6.8	4.7	69.03	1334.9	243.6	18.25

Ni	White Island						Ambae						Kilauea					
	Rep 1	Rep 2	Rep 3	Mean	SD	RSD %	Rep 1	Rep 2	Rep 3	Mean	SD	RSD %	Rep 1	Rep 2	Rep 3	Mean	SD	RSD %
1:10	18.5	18.6	17.9	18.4	0.4	2.04	2.1	2.2	2.1	2.1	0.0	0.71	6.1	6.3	6.3	6.2	0.1	1.93
1:20	15.3	14.6	15.0	15.0	0.3	2.24	2.3	2.1	2.2	2.2	0.1	3.96	6.8	7.0	6.9	6.9	0.1	1.61
1:500	17.8	17.6	17.5	17.6	0.2	0.89	2.0	2.1	2.2	2.1	0.1	4.83	7.6	8.0	7.6	7.7	0.3	3.40
1:1000	17.8	17.2	16.9	17.3	0.4	2.38	2.1	1.9	2.2	2.0	0.2	8.10	7.8	7.6	7.6	7.7	0.1	1.47
WL (24h)	20.2	20.3	20.8	20.4	0.3	1.56	2.5	2.6	2.6	2.6	0.1	3.14	8.8	9.1	9.2	9.1	0.2	2.29
SLF1 1:100 24h	15.5	15.5	15.5	15.5	0.0	0.16	2.3	2.3	2.3	2.3	0.0	0.67	7.8	7.6	7.5	7.6	0.2	2.14
SLF2	16.6	16.3	16.2	16.4	0.2	1.26	2.3	2.3	2.5	2.4	0.1	4.65	7.6	8.0	8.1	7.9	0.3	3.33
SLF3	17.1	17.0	17.0	17.0	0.1	0.32	2.2	2.2	2.4	2.3	0.1	4.90	7.8	7.9	7.6	7.8	0.1	1.83
SLF4	11.2	10.6	10.9	10.9	0.3	2.80	2.3	2.4	2.3	2.3	0.1	2.41	7.8	8.1	8.0	8.0	0.2	2.22
10 min	17.8	17.4	17.5	17.6	0.2	1.09	2.2	2.2	2.3	2.2	0.1	4.02	7.5	7.8	7.7	7.7	0.2	2.03
1 h	16.7	16.3	16.5	16.5	0.2	1.21	2.2	2.2	2.3	2								

Highlights

- Investigation of a lung fluid leachate method to assess inhalation hazards of ash.
- Element release in leachates is both method-parameter and sample dependent.
- Inclusion of lung surfactant is not necessary when assessing ash leachates.
- Optimal method parameters are extraction up to 24 h at 1:100 ratio.
- Water leach can be used as a conservative estimate of lung bioaccessible elements.

Declaration of interests

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

The authors declare the following financial interests/personal relationships which may be considered as potential competing interests:

Journal Pre-proof