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1 A systematic comparison of deforestation 2 drivers and policy effectiveness across the 3 Amazon biome

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16 Abstract:

17 The Amazon biome, spanning nine countries, has one of the highest rates of deforestation
18 worldwide. This deforestation contributes to biodiversity loss, climate change, the spread of
19 infectious diseases, and damage to rural and indigenous livelihoods. Hundreds of articles have
20 been published on the topic of deforestation across Amazonia, yet there has been no recent
21 synthesis of deforestation drivers and deforestation-control policy effectiveness in the region.
22 Here we undertook the first systematic review of papers published between 2000 to 2021 that
23 have causally linked proximate and underlying drivers and policies to deforestation outcomes
24 in Amazonia. In the 155 articles that met our inclusion criteria, we find that causal research is
25 concentrated in Brazil, and to a lesser degree Peru, Ecuador, and Bolivia. There has been little
26 study of the Guianas, Venezuela or Colombia. Large- and small-scale agriculture linked to
27 improved market access and high agricultural prices are frequently researched underlying
28 drivers of deforestation across the heavily researched regions. In the Guianas research focuses
29 on mining with little focus on underlying causes. Research on infrastructure expansion, mining,
30 and oil extraction and on technological, sociocultural, and institutional factors remains sparse.
31 Many public and private policies have been found to be effective in controlling deforestation
32 across the biome, with protected areas standing out as particularly successful in slowing
33 deforestation, vis-à-vis supply chain approaches. Frontier age, land tenure, and policy
34 interactions are key moderating factors affecting the outcomes of different underlying causes
35 and policies. Our findings indicate a greater need for research on i) additional deforestation
36 drivers beyond agriculture and economic factors, ii) the complex interactions between different
37 drivers and deforestation control policies, iii) causes underlying deforestation in low or new
38 deforestation areas, and iv) the dynamics between Amazonian subregions and countries.

Understanding the extent and diversity of deforestation drivers and effectiveness of existing deforestation mitigation policies across Amazonia is a necessary first step toward designing policies to further reduce deforestation in the biome.

Keywords: sustainable development, South America, land systems science, policy

1. Introduction

The Amazon biome holds the largest tropical forest in the world and contains more than 10% of known plant and animal species [1]. It is home to 47 million people [1] and spans 6.7 million km², divided between Brazil (62%), Peru (11%), Bolivia (8%), Colombia (6%), Venezuela (6%), Guyana (2%), Suriname (2%), Ecuador (2%) and French Guiana (1%) [2]. The biome, also known as Amazonia, is defined by a set of biogeographical criteria and includes the vast lowland rainforest region of the Amazon River basin and several subregions, such as the Guiana shield to the north, the Planalto and Gurupi regions to the south west and east, and a part of the Andes to the west [3].

Although often reduced to one entity, Amazonia holds immense environmental and sociocultural diversity. Its regions are connected through the Amazon river and its tributaries as well as through the transboundary movement of nutrients, animals and people [4,5]. The biome provides its inhabitants with food, raw materials, fresh water and regional climate regulation, estimated to be worth between US\$56 to US\$737 per hectare per year [6]. It also serves as the foundation for the spiritual and cultural identity of various ethnicities that in turn preserve functions of the biome [7]. Additionally, the forest is also a major global climate regulator through its large and rapid influence on the water cycle, planetary energetics and global atmospheric composition [8].

Since the 1970s, the clearing of Amazonian forests has intensified and over 17% of primary forest cover has been lost [2]. Deforestation in the Amazon biome contributes to widespread biodiversity loss [9], regional and global climate change [10], changes to the hydrological cycle [11–13], the spread of infectious diseases [14], and damage to the livelihoods of indigenous people [15]. Today, the Amazon biome is the most threatened ecoregion in the world in terms of the area cleared each year, accounting for over 3.4 million ha of forest loss per annum [2]. Climate change, forest degradation and deforestation are increasing the likelihood of an irreversible ecosystem dieback across the entire Amazon, with drastic regional and global consequences [16,17].

The number of studies and initiatives providing crucial information on patterns and processes of land use and land cover change has substantially increased over the past decades. As one of the first attempts to synthesize evidence of causes of deforestation across the tropics, Geist and Lambin [18] created a framework of proximate and underlying drivers, which has since been widely adapted in land use science. For example, Armenteras et al. [19] used it to show that deforestation drivers and rates vary substantially across South American countries and forest types. In response to the issue of widespread

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3 79 deforestation, studies on conservation and policies against deforestation have gained relevance. As with
4 80 research on land use change, large-scale syntheses of policy knowledge provide insightful evidence.
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6 81 Börner et al. [20] for instance found that the effectiveness of conservation policies differs between
7
8 82 regions, highlighting the importance of local contexts when assessing policies.

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10 83 Existing deforestation studies that cover the entire Amazon region have been conducted on the global
11 84 or macroregional level (e.g., Armenteras et al. [19] for South America), often focusing on land use
12 85 change mostly due to agriculture [21,22] and sometimes examining links to underlying factors such as
13 86 global markets [23,24]. Other studies concentrate on analyzing patterns of one direct cause of
14 87 deforestation [25,26] or the effectiveness of different policy options to reduce deforestation [20,27]
15
16 88 across regions.

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19 89 Several initiatives have produced and connected knowledge on deforestation across the Amazon
20 90 biome in recent years, such as the MapBiomas Amazonia project [2], the Monitoring of the Andean
21 91 Amazon Project [28], and the georeferenced socio-environmental information network (RASIG) [29].
22 92 Additionally, in 2020, the UN Sustainable Development Solutions Network convened The Science
23 93 Panel for the Amazon [30], a large interdisciplinary team of researchers, to assess the state of the entire
24 94 Amazon and give recommendations for improvements in an extensive report. This massive undertaking
25 95 provided complex, in-depth, expert driven syntheses across a wide range of topics and disciplines
26 96 relevant to understanding the history and current trajectory of the Amazon biome, across both urban
27 97 and rural contexts.

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30 98 Despite these major advancements in monitoring and synthesizing land and policy processes, there
31 99 is no systematic, concise, multi-region overview of the present-day deforestation drivers and policies,
32 100 or how these factors vary across Amazonian regions. Older pantropical studies, such as Geist and
33 101 Lambin [18] and Rudel [31] may no longer reflect the realities of Amazonia in the early 21st century,
34 102 which has seen exponential growth in global commodities demand and international trade and finance,
35 103 both of which have exerted significant pressure on the tropics. Specifically, current research on
36 104 deforestation in the Amazon faces a gap between macroregional or global studies and local case studies.
37 105 A systematic review is particularly well placed to bridge this gap by gathering existing cases at smaller
38 106 scales and studying them under a uniform analytical and coding framework [32,33].

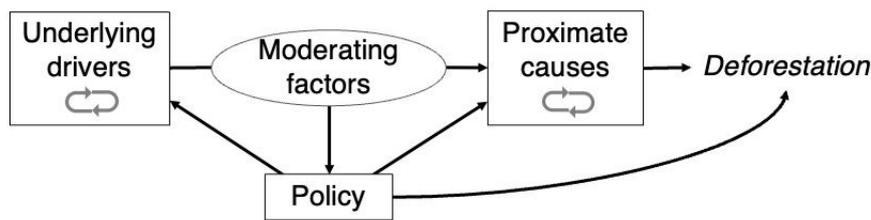
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41 107 The objective of this study is to systematically identify and qualitatively compare current knowledge
42 108 of the drivers of deforestation and deforestation-control policies (e.g., protected areas, supply chain
43 109 policies, public regulations) across Amazonian countries and regions, and to point out critical
44 110 knowledge gaps about these issues, both regionally (i.e., between Amazonian nations) and by topic (i.e.,
45 111 by specific drivers, causes, moderating factors, and policy effectiveness). We build on the framework
46 112 of proximate and underlying drivers by Geist and Lambin [18], which has been widely adapted and
47 113 become the gold standard for meta-studies in land use science [33]. Deforestation policy effectiveness
48 114 is defined as the degree to which the policy successfully reduces deforestation in the target region [34].
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3 115 **Proximate causes** are land use changes that directly cause deforestation on the local or regional level.
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5 116 These include specific types of agriculture, mining, and other extractives (oil, timber), and infrastructure
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7 117 that directly replace forests. In our study we also examine drivers that cause deforestation indirectly by
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9 118 displacing other proximate deforestation causes. This would include, for example, crop expansion
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119 displacing cattle ranching to a forested place.

11 120 **Underlying drivers** are indirect forces that trigger deforestation through encouraging a proximate
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13 121 cause. These include both socioeconomic and bio-physical underlying drivers. For continuity we use
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15 122 the overall groupings of drivers used by Geist and Lambin [18]: economic, institutional, cultural,
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17 123 demographic, environmental, and technological. Economic drivers include both macroeconomic trends
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19 124 (i.e., employment) and microeconomic household conditions. Institutional drivers relate to formalized
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21 125 rules and land use, including land tenure and policies that directly encouraged deforestation and
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23 126 agricultural settlement. Cultural drivers relate to the shared meanings, norms, and values that actors
24
25 127 have around particular land use behaviors. Demographic drivers pertain to both the household
26
27 128 composition and broader population trends. Environmental drivers have to do with the climate and soils.
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29 129 Technological drivers are anything that influence the total factor productivity (i.e., output per input).
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31 130 For continuity with Geist and Lambin [18], land quality and farm size is also considered a technological
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33 131 driver.

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35 132 Ultimately the broad framework depicted in Figure 1 is based on some more specific theoretical
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37 133 models of land use behavior. On the one hand there is the rational actor, income maximization model
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39 134 that suggests an actor will clear land for a particular use (the proximate driver) if the net present value
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41 135 of the expected income less the clearing costs is greater than its value as a standing forest and all other
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43 136 land uses [35]. Aspects like culture extend this model in terms of what constitutes value in a non-
44
45 137 monetary sense [36]. Environmental and technological factors influence the overall productivity of a
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47 138 particular land use in a particular place, affecting its profitability [37,38]. Institutions that influence
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49 139 access to land also affect the expected income [39].

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51 140 When considering smallholder agriculture, the behavioral model is often based around meeting
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53 141 subsistence food, fuel, and housing needs considering leisure-labor tradeoffs, which is why household
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55 142 microeconomic factors are also important, as well as overall demographic pressure in an area [40]. A
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57 143 lack of productive assets is a frequent constraint on behavior [41,42]. Capital availability (to buy land
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59 144 or hire labor or equipment to clear) is one of these key constraints. On the demographic side, labor
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145 availability can be another key constraint.



146
147 *Figure 1: Conceptual model of major mechanisms influencing deforestation*

149 To this framework we added other *policies* that directly provide incentives to pursue deforestation
150 via rewards or penalties. Building from the categories used by Börner et al. [20] and Echeverri et al.
151 [43] these types of policy approaches include: area-based conservation (i.e., protected areas, indigenous
152 areas, restricted use areas, mixed conservation areas), public policies (i.e., taxes, quotas), financial
153 instruments (i.e., payments for environmental services), and supply chain policies (e.g., certifications,
154 standards, commitments and pledges). Deforestation policy effectiveness is defined as the degree to
155 which the policy successfully reduces deforestation in the target region. By suggestion of Magliocca et
156 al. [44] and van Vliet et al. [33] we also include *moderating factors*, as those factors that influence the
157 way in which an underlying driver affects a proximate cause or the implementation and effectiveness
158 of a policy.

159 Using this framework (Figure 1), we answer the following questions as a means of synthesizing the
160 existing research results: (1) *How do the proximate causes, underlying drivers, and moderating factors*
161 *of deforestation vary across the Amazon Biome and over time?* (2) *How effective have deforestation*
162 *control policies been across different regions and over time?*

163 To assess research gaps and uncertainties to guide further studies we ask: (3) *How does the*
164 *availability and rigor of evidence vary across regions and topics (i.e., drivers, causes, moderating*
165 *factors, and policy effectiveness)?* (4) *How does the evidence generated by case study (i.e., local, state,*
166 *or national) research differ from regional (Amazon-wide) studies?*

167 Our findings provide a comprehensive, yet accessible overview on the geographic heterogeneity of
168 deforestation drivers and combined governance effectiveness in Amazonia. Combining drivers,
169 moderating factors, and policies in one synthesis enables us to characterize deforestation pathways and
170 policy effects in more detail and within context and to identify major remaining knowledge gaps.

172 2. Methods

173 To answer our research questions, we systematically reviewed the available literature on Amazonian
174 deforestation and deforestation policies published between 2000 and 2021 and coded this evidence base
175 using a framework derived from our theoretical model to identify specific deforestation drivers and
176 policies and to link them to individual Amazonian regions.

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3 177 Studies were distinguished by whether they were analyzed at administrative level 0 (i.e., national),
4 178 administrative level 1 (i.e., state/province/department), or administrative level 2 or higher (i.e.,
5 179 municipalities, districts, or individual communities). The journal search started from the year 2000 as
6 180 the purpose was to provide an updated view since Geist and Lambin [18] and Rudel [31] whose work
7 181 focused on earlier periods. This was also necessary to constrain the workload and ensure the relevance
8 182 of our findings. However, we included studies that covered time periods extending before 2000 if they
9 183 were published since 2000. We coded the drivers and policy outcomes analyzed in these papers using
10 184 a uniform analytical and coding framework based on the theoretical approach described above (Figure
11 185 1), as well as location and time span. Included studies were further divided into *cases*, with an individual
12 186 case referring to a specific study area, meaning a single study including several study areas was counted
13 187 as several cases.

14 188 Because deforestation processes are highly complex and context-specific, we chose a qualitative
15 189 approach to adequately review the literature. This also allowed us to carefully navigate the aggregation
16 190 and generalization of findings, as case studies vary across spatial and temporal scales and
17 191 methodological spectra. After reviewing selected case studies, we compared this bottom-up synthesis
18 192 of sub-Amazonia scale work to top-down knowledge from Amazon-wide studies on deforestation
19 193 drivers and control policies. Qualitative systematic reviews offer an inclusive yet differentiated
20 194 synthesis of existing literature, while highlighting areas that require more research attention.

21 195 We followed recommendations for systematic review best practice, proposed by Haddaway et al.
22 196 [32] and van Vliet et al. [33], to avoid bias and to improve the transparency and consistency of the
23 197 analysis. The two largest peer-reviewed article databases “Scopus” and “Web of Science” were used to
24 198 retrieve peer-review journal articles, as they allowed advanced article searches with complicated search
25 199 terms (see Table S1 for the search strings). Additionally, we used the platform SciELO to capture
26 200 articles. We included English as well as Spanish and French, however, we did not identify any French
27 201 articles that met our selection criteria. We did not do a separate search in Portuguese because our
28 202 selection already included a vast number of studies in Brazil (where studies in Portuguese were located).
29 203 We did not use Google Scholar since it only offers simple searches and also contains inconsistently
30 204 catalogued grey literature in this field, which could lead to biases if included [45]. We did not include
31 205 grey literature due to the high number of peer-reviewed papers already captured by Scopus, Web of
32 206 Science, and SciELO and our inability to consistently capture grey literature.

33 207 To ensure no relevant paper was missed, a broad search query was defined, using a range of terms to
34 208 capture deforestation (land cover change, forest loss, forest clearing etc.) and searching all available
35 209 database sections (title, abstract, keywords). Furthermore, to limit the search to the study area of the
36 210 Amazon biome while enabling a country-level comparison, all states, departments and provinces (from
37 211 here on all referred to as states) that are part of the Amazon biome, were included in the query. The
38 212 border of the Amazon biome was defined according to the MapBiomias Amazonia map [2]. Finally, to
39 213 ensure papers identified were relevant to current conditions, the search was limited to electronically

214 accessible articles published since 2000 to the time of the search (2021). The search query was then
215 adapted to work identically for the advanced search machines of the two databases.

216 The search in March 2021 identified 7,401 articles from both platforms combined, of which 5,229
217 articles were unique. Using the same search term, the search was repeated in December 2021 to include
218 more literature in Spanish and French. The search identified 160 unique articles in Spanish (16
219 duplicates) and 61 unique articles in French (10 duplicates) from Scopus and Web of Science combined.
220 A different search term consisting of numerous synonyms was used as SciELO does not allow for more
221 elaborate search term formulation like Scopus and Web of Science. The synonyms were translated with
222 the aid of expert native speakers, to include as many terms for deforestation as possible. The search on
223 SciELO in December 2021 yielded 65 Spanish articles, of which 10 were duplicates from Web of
224 Science and Scopus. The evolution of the search term and results is documented in Tables S1 and S2.

225 The retrieved articles were first filtered by reading the abstract to determine if they fit the overall
226 topic of the review. Those articles that appeared relevant were compared to the following set of criteria
227 to ensure only comparable studies with a rigorous methodological standard were included in the review:
228 i) the study area had to be within the Amazon biome, ii) the study had to include a scale of analysis at
229 the national level or below, and iii) the study had to include a rigorous causal analysis of the impact of
230 at least one driver or policy on actual (i.e., past) deforestation.

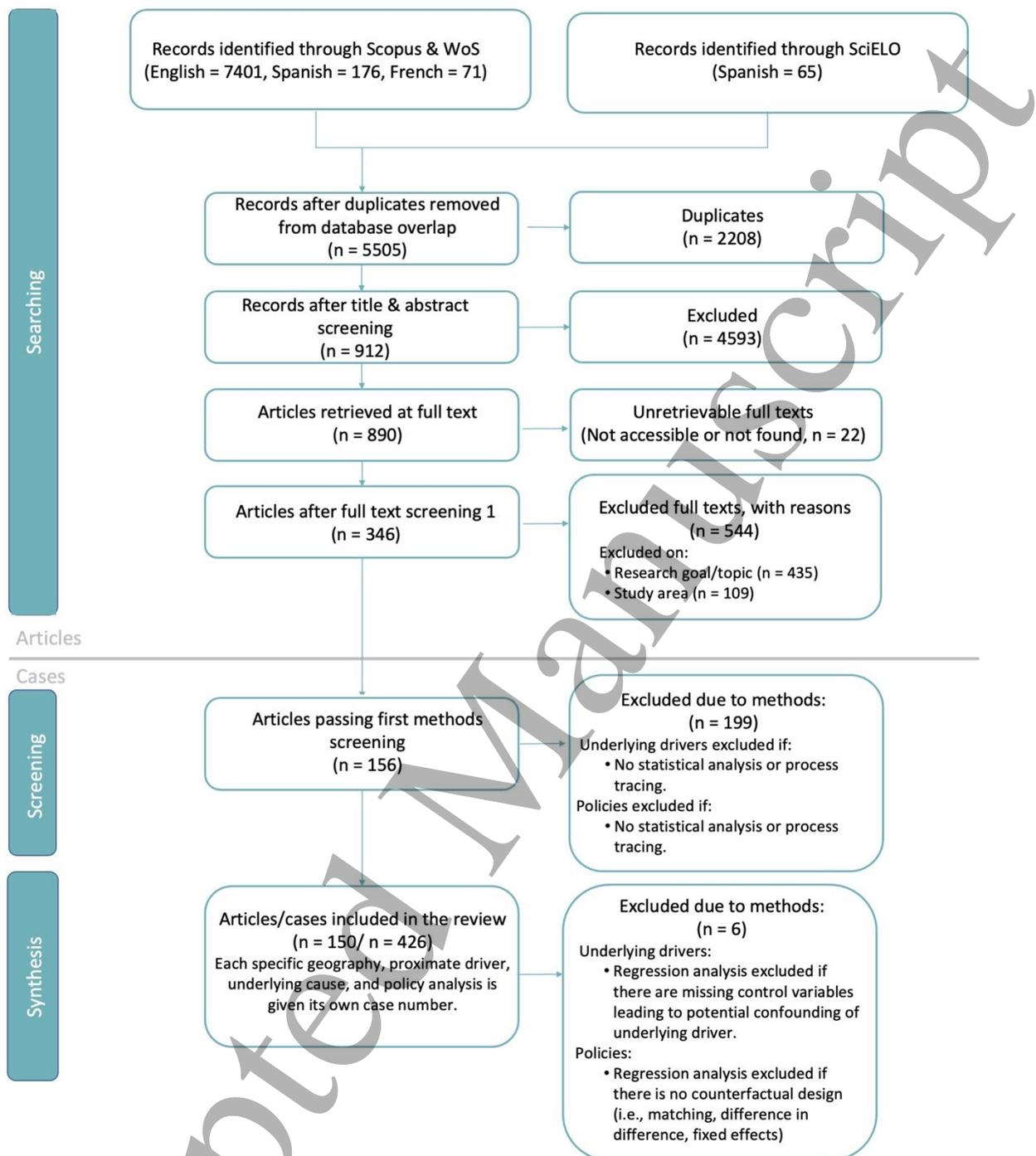
231 For papers on proximate causes of deforestation, we included: i) remotely sensed estimates of land
232 use transitions from forests, ii) correlation studies between deforestation and different land uses (i.e.,
233 where a deforestation map is developed first and then deforested areas within a region are assigned to
234 the land uses that are most correlated with forest loss or allocated by another method, such as
235 suitability), and iii) drivers quantified in interviews. While remotely sensed estimates that directly
236 measure transitions from forest to a specific land use within a given pixel provide the strongest certainty
237 that a specific land use is a primary cause of deforestation, we also include studies that look at the
238 correlations between deforestation at higher spatial scales since remotely sensed data on the specific
239 land uses that replace forests are more recent (at the biome scale such data were not available until the
240 development of MapBiomias. We include interview data in addition to remotely sensed or statistical
241 estimates as we believe that farmers' self-reporting of land use provides valid data. While self-reporting
242 is likely to underestimate overall deforestation levels, there is less reason to suspect such bias regarding
243 the specific cause of deforestation where such deforestation is detected.

244 For underlying drivers and policy impacts where causality is more difficult to assess, a robust
245 econometric analysis that controlled for confounding factors was required. A significant relationship
246 ($p < 0.05$) between the dependent variable (deforestation or deforestation for a specific land use) and
247 the independent variable (i.e., the underlying driver or policy) was necessary for the case to be recorded
248 in the synthesis. The inclusion criteria for policy analyses further required that a sound counterfactual
249 method was used to ascertain what would have happened in the absence of the policy. This could include
250 matching, difference-in-difference, or fixed effects econometric methods. We had intended to include

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3 251 qualitative process-tracing papers as well but identified no papers using this method. Examples and
4
5 252 reasons for excluded articles are given in S4. The full coding scheme is presented in S5 in a separate
6
7 253 tab from the results.

8 254 Each specific geography, proximate driver, underlying cause, and policy analysis is given its own
9
10 255 case number. Articles excluded for one part of the conceptual model (i.e., underlying cause, policy
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12 256 effectiveness) can still appear in the coded data for the proximate causes, but the cases pertaining to the
13
14 257 underlying cause or policy effectiveness would not be included in the analysis if they did not meet the
15
16 258 inclusion criteria. This approach demonstrates not only how many places are studied, but also for a
17
18 259 given geography, how many different causes, drivers, and policies have been rigorously examined.

19 260 All titles and abstracts were screened by the first author. Before starting the coding process, ten
20
21 261 articles were coded independently by the first author and two other authors. The scheme was then
22
23 262 adjusted to guarantee future consistency in the coding process. After that only the first author coded the
24
25 263 articles. The final coding scheme included fields describing the study region and period, the impact and
26
27 264 links of analyzed drivers and policies, and the methodology of the causal analysis. After coding, drivers
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29 265 and policies were aggregated and organized in clusters according to Geist and Lambin [18], the policy
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31 266 mechanism and how they were framed in the studies.
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268 *Figure 2: ROSES flow diagram for systematic reviews. Using the template provided by Haddaway et*
 269 *al. [46].*

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271 3. Results

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273 Our review methodology identified 150 articles in the literature search. From these articles we
 274 identified 426 cases (spreadsheet S5 includes all coded data with separate tabs for cases meeting the
 methodological criteria and those that did not). The amount of research per year included in the

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3 275 review increased steadily since 2000, with 80% of reviewed studies published after 2010 (see Figure
4 276 3a). All papers were published between 2000 and 2021, but the studies have analyzed deforestation
5 277 and land use data as early as 1970. A majority of the studies focused on the period after 2000. Figure
6 278 3b shows the geographic distribution of the 426 analyzed cases (since 1970) and deforestation as
7 279 estimated in MapBiomas since 1985 (since data is not available before then) across the entire Amazon
8 280 biome.
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15 282 *(1) How do the proximate causes, underlying drivers, and moderating factors of deforestation vary*
16 283 *across the Amazon Biome and over time?*

17
18 284 *Proximate causes*

19 285 Agricultural expansion was found to be the main proximate cause of deforestation in all countries
20 286 except the Guianas. Among studies that examined multiple proximate causes, *large-scale agriculture*,
21 287 which includes both cattle ranching and commodity crops, was found to have a larger impact on forests
22 288 compared to other forms of agriculture. *Pasture expansion* was the primary proximate cause of
23 289 deforestation in the Bolivian, Ecuadorian, Colombian and Brazilian Amazon across the entire period.
24 290 In Southern Amazonia the rise of *commodity crops* (soy, oil palm) as a driver of deforestation is
25 291 documented by several cases, especially until the mid-2000s. Studies in the Brazilian Amazon reported
26 292 a considerable indirect effect of soy expansion displacing pasture into new frontier areas [47,48]. The
27 293 limited cases show oil palm expansion is linked to deforestation in Peru, especially in Ucayali, but not
28 294 in Pará, Brazil. *Small-scale or subsistence agriculture* was identified as a small but consistent driver in
29 295 the Bolivian, Peruvian and Brazilian Amazon and in French Guiana. Regionally, small-scale agriculture
30 296 caused larger amounts of deforestation in Brazil versus other Amazonian countries, specifically in
31 297 Roraima and Amapá [49,50], and in Peru in the department of Amazonas [51]. In French Guiana, small-
32 298 scale agriculture was the main agricultural driver and overall caused intermediate amounts of forest loss
33 299 [52].

34 300 *Coca cultivation* was found to be a minor and decreasing cause of deforestation in Guaviare,
35 301 Colombia, and was often replaced with pasture [53,54]. Coca is mostly grown in more accessible
36 302 territories outside of Amazonia.

37 303 Small-scale *gold mining* (including both alluvial and open cast mining) was reported as the main,
38 304 increasing cause of deforestation in Guyana and Suriname and had an intermediate impact on forest
39 305 loss in French Guiana and in the Peruvian Amazonia. Mining spread along the gold-bearing greenstone
40 306 belt of the Guiana Shield and along rivers [55]. In Peru, forest loss was caused equally by small- and
41 307 large-scale gold mining. Although mining (inclusive of all types) had a very small direct impact on
42 308 Amazonian forests in Brazil [50], its indirect impact on deforestation was a lot higher through road
43 309 building that allowed agriculturalists to access new forest areas [56–59]. *Oil exploration* was analyzed
44 310 as a minor local cause of deforestation in Amazonas, Brazil [60], but was also mentioned by several
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3 311 studies as a pioneer cause opening up areas to development and deforestation in the Colombian [61]
4 312 and Ecuadorian Amazon [62–65].

5
6 313 Deforestation for *urban areas, settlements, roads and hydropower dams* in the Brazilian, Bolivian,
7 314 Colombian, and Peruvian Amazon was very small compared to other drivers [50]. Expansion often
8 315 occurred onto agricultural or other previously cleared areas and therefore avoided direct deforestation
9 316 [51,66–68]. In contrast, one study found that due to the generally low levels of deforestation in French
10 317 Guiana, the construction of a hydropower station in 1994 contributed to a large amount of the country's
11 318 forest loss. Urban expansion however remained a minor driver along the coast [52].

12
13 319 While infrastructure is a proximate driver of deforestation according to the framework, studies mostly
14 320 treated it as an indirect driver (by enabling other proximate drivers). In the Brazilian Amazon, the
15 321 construction of large-scale hydropower stations caused considerable local impacts and further induced
16 322 agricultural and urban expansion in the area, thus indirectly causing more deforestation [69–71]. Several
17 323 studies observed that during and after the construction of the interoceanic highway (IOH) in Madre de
18 324 Dios, Peru, deforestation due to agriculture and mining increased in the surrounding areas [72].
19 325 Similarly, many studies across the biome linked forest loss to roads, reflected by a distinctive fish-bone
20 326 pattern of deforestation. In fact, accessibility to markets via roads was identified as an important
21 327 underlying driver in the Amazon [73–76]. Infrastructure thus acts as a twofold driver, both proximate
22 328 and underlying, and often dictates the pattern of deforestation.

23
24 329 Only a few studies found that *logging* was associated with complete deforestation (rather than forest
25 330 degradation). Studies in French Guiana [77], and in the Ecuadorian [65] and Brazilian Amazon [78–80]
26 331 show that logging had minor direct impacts on deforestation. In Ecuador timber extraction was found
27 332 to be a byproduct and additional incentive for agricultural expansion, as it provided income to hire labor
28 333 for clearing [81]. Similarly, *fire* is not necessarily referred to as a proximate cause of deforestation [50].
29 334 In the Brazilian and Colombian Amazon, fire was used to clear the remaining or regrowing forest after
30 335 the logging for agriculture (e.g., in slash-and-burn practices) or for resource extraction. Escaping fires
31 336 affecting surrounding forests were also reported but their impact was not quantified [53,82–84]. In
32 337 extractive reserves very small-scale deforestation by extractivists was observed in the 1980s (in patches
33 338 of up to 1.8 ha) in Acre [80] and in the 1990s and 2000s (in patches up to 4.2 ha) in Amapa [79], both
34 339 in Brazil. But this deforestation was still deemed be associated with crop production by the extractivists
35 340 and was replaced by larger-scale ranching and agriculture over time. Besides anthropogenic
36 341 deforestation, *natural causes* for forest loss such as flooding and wind were observed and quantified as
37 342 a small impact on Amazonian forests in Peru [72], Bolivia [67] and Brazil [50,56].

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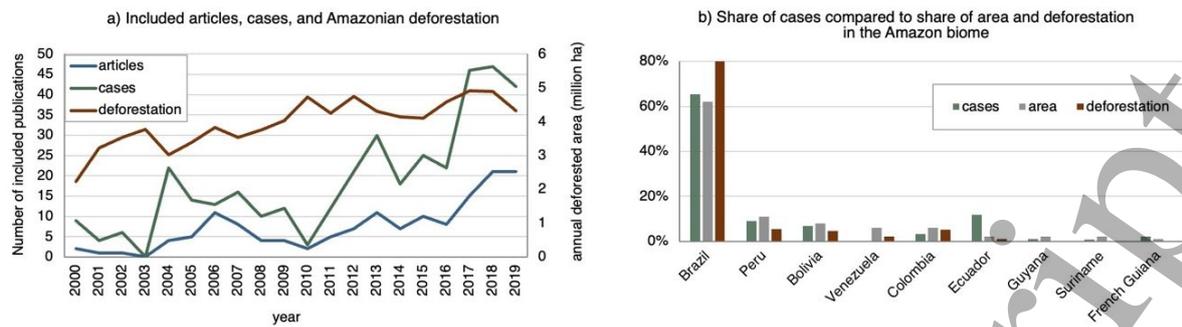


Figure 3: Number of proximate causes (top) and underlying drivers (bottom) identified through causal analyses in the selected studies, including the main drivers per category and the number of links between them (represented also by the thickness of the arrows; not including moderating agents).

Underlying drivers

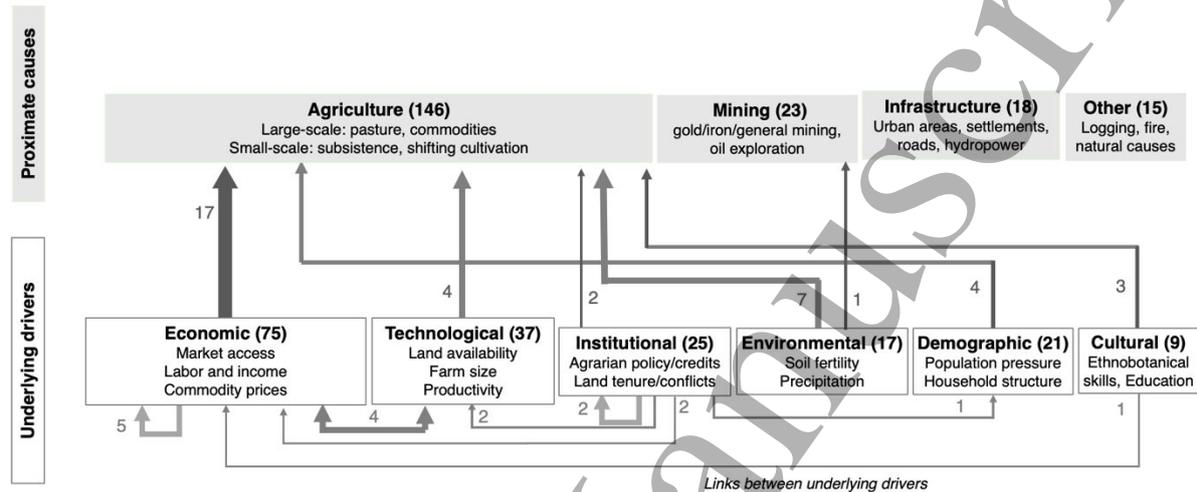
Economic: The most prominent economic drivers affecting deforestation in the Amazon are related to market conditions that influence the prices that land users receive for the proximate deforestation driver (i.e., agriculture, mining, etc.) (Figure 4). This aligns with the classic net present value model of deforestation behavior (whereby land users are assumed to be motivated by differences in the discount flow of income they can get from different land uses). Studies in Peru [51,73], Bolivia [84,85], Ecuador [76,86] and Brazil [87,88] find that *accessibility via roads* and the resulting *lowered transportation costs to local markets* was associated with increased deforestation, especially due to pasture expansion and small-scale agriculture. Deforestation driven by local market access often spreads in a recognizable fishbone pattern from roads in less remote areas [89–91]. In more remote and less developed areas in the Guianas and earlier in Colombia, Peru and Pará, Brazil, roads had not extended into forests to the same degree and thus deforestation occurred *predominantly along navigable rivers* as this was the primary market access route [55,92].

Since the 1990s, corresponding to the growth in commodity agriculture for export markets, temporal variations in *commodity prices* have increasingly affected deforestation for agriculture and mining, especially in Brazil [93–95]. However, one study suggests that where pasture and cattle ranching are already present, an increase in the relative crop-to-beef price index has been associated with a reduction in deforestation by raising agricultural input prices, driving low productivity cattle ranchers out of agriculture [96]. Conversely, high *wood prices* were correlated with decreased deforestation in the Brazilian Amazon, as forests had investment value and were not fully cleared for agriculture (rather selectively logged) [94,97,98]. In the Ecuadorian Amazon however, extracted timber volume correlated with deforestation, as income from selling timber as a byproduct was often used for further agricultural expansion [81].

Several variables that pertain in one way or another to wealth and capital availability have contradictory results (both positive and negative effects) at the Brazilian Amazon-scale (measured by

375 GDP per capita) [78,94,99–102] and at the micro-scale (measured by many different proxies for
376 household income, wealth, and off-farm work) in Brazil, Ecuador, and Peru [97,98,103–109].

377 The availability of labor was consistently linked to small-scale deforestation in the Peruvian [103],
378 Brazilian [104,110], Ecuadorian [65,86] and parts of the Bolivian Amazon [111] during the 1990s. This
379 finding corresponds more closely to subsistence theories of deforestation (whereby decisions are
380 heavily influenced by leisure and labor tradeoffs and overall input scarcity), rather than rational-actor
381 theories.



382
383 Figure 4: Number of proximate causes (top) and underlying drivers (bottom) identified through causal analyses in the
384 selected studies, including the main drivers per category and the number of links between them (represented also by the
385 thickness of the arrows; not including moderating agents).

386 **Demographic:** Population size and growth were found to be important underlying drivers of
387 deforestation at the regional scale in Colombia [92] and Brazil [100]. Yet, at more local scales in Brazil
388 and Ecuador the impact of household size and composition had varied impacts on deforestation with no
389 clear directionality [65,81,86,98,105,112,113].

390 **Technological:** Technological drivers were most often studied in the context of agricultural
391 expansion and thus implicitly linked to agriculturally caused deforestation, even where the proximate
392 cause was not directly analyzed. With more forest available there are more opportunities to deforest and
393 the price of land usually lower. This was confirmed by regional and local studies in Ecuador since the
394 1980s [65,76,81,109] and in more recent cases in the Brazilian Amazon [88,97,114] and in Madre de
395 Dios, Peru [115]. Farm size was also associated with greater deforestation, but in these cases, it was
396 often due to the higher amount of forest available in those farms, rather than farm size itself being a
397 reason for greater deforestation. In the Brazilian Amazon, however, two earlier cases around 1995 found
398 smaller farms deforested more than large ones [104].

399 The relationship between intensification and deforestation depended on capital availability. In
400 general, farmers with higher agricultural productivity and higher efficiency had higher deforestation.
401 However, very inefficient farmers also had high deforestation because they lacked capital for
402 intensification and thus relied on deforestation to expand production. While higher agricultural

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3 403 productivity often led to more deforestation, agricultural diversification, e.g. , through agroforestry, was
4 found to decrease deforestation in colonization frontiers in Rondônia [106–108,116].

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6 405 **Environmental:** Soil quality, precipitation, and slope, were found to be important underlying drivers
7 of deforestation in Brazil, Ecuador, and Bolivia, by determining the environmental suitability for
8 406 different land uses. However, the importance of different environmental factors depended on the
9 407 proximate cause in question. Across the Brazilian Amazon, earlier studies noted that deforestation was
10 408 positively correlated with increased precipitation and low soil fertility, as the ease of initial clearing
11 409 appeared to be a more significant driver of land clearing than the relative productivity of the following
12 410 land use. Less fertile, washed-out soil had less vegetation, which facilitated clearing but required more
13 411 land to achieve the same production [99]. After 1995, however, when commodity agriculture became a
14 412 more relevant proximate cause of deforestation, lower precipitation, good soil quality, and less drought
15 413 attracted greater agricultural conversion and more deforestation [76,87,88,94].

16 414
17 415 **Institutional:** Institutional drivers focused on agricultural settlement policies and land tenure and
18 416 conflict. Only one study examined the broader political context as a driver of deforestation. It found
19 417 that corruption and campaign funding in municipal re-election cycles increased deforestation in Brazil
20 418 [117].

21 419 Land contention has remained an unresolved issue for several decades and was linked repeatedly to
22 420 increased deforestation in the Brazilian Amazon. Land insecurity, inequality and occupation originating
23 421 from inconsistencies in the institutional framework for property rights have affected deforestation
24 422 behaviors. Full property rights are costly to attain and do not protect against encroachment.
25 423 Consequently, competing landholders and squatters cleared "unproductive" forest to claim the land.
26 424 Landholders also used preemptive clearing to avoid land occupation movements and expropriation by
27 425 INCRA [95,97,114,118].

28 426 Once secured, land tenure has been shown to have decreased deforestation in the Brazilian Amazon.
29 427 Where they had the right to clear land, renters and sharecroppers converted more forest to agriculture
30 428 than land owners, who were usually implementing more long-term activities requiring less land
31 429 [78,97,116]. However, in the Ecuadorian Amazon the opposite relationship was found. Land titles were
32 430 required as collateral to acquire credits for cattle ranching in the 1990s, thus deforestation increased
33 431 when that credit was made available [86].

34 432 **Cultural:** Cultural drivers were mostly assessed in the research hotspot of Beni, Bolivia, regarding
35 433 the deforestation practices of the Tsimane' indigenous people. Traditionally, Tsimane' practice
36 434 subsistence shifting cultivation as a secondary livelihood to fishing and hunting and thus only clear very
37 435 small areas at a time. Generally, it was found that people with good ethnobotanical skills used land
38 436 more efficiently and cleared less for agriculture [111,119], while those farmers and villages that shifted
39 437 towards more Western values had higher deforestation [90,91].

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3 438 Finally, greater *education* was consistently associated with increased deforestation, both in the
4 Brazilian [114] and Ecuadorian Amazon, probably due to the need to increased land clearing to generate
5 439 income in order to fulfill aspirational consumption needs [76,86].
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9 441 *Temporal variance and moderating factors*

10 442 Pasture-driven expansion is documented in the studies since 1975 [99,100], often encouraged by
11 443 government-supported colonization programs, whereas commodity agriculture only accelerated in the
12 444 1990s and 2000s due to economic liberalization and rising market demand [99,120]. In the Brazilian
13 445 Amazon, the establishment of agrarian settlements [88,94] and the availability of agricultural credit
14 446 greatly increased deforestation during the 1990s and 2000s [94,95,104,105,121].

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16
17 447 In the Peruvian Amazon, agrarian policy changed with regimes but especially supported cattle
18 448 ranching and more recently infrastructure projects (e.g. the IOH) for trading, causing more deforestation
19 449 [89]. Gold mining has been increasing since 1984, especially in Madre do Dios, Peru [122–124]. In
20 450 Colombia changes in pasture-driven deforestation (in 1985-2001) were linked to changes in
21 451 demographic pressures [92].

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25 452 In Brazil both pasture-driven and commodity-driven deforestation peaked in the early 2000s prior to
26 453 the introduction of improved deforestation-control policies and increases or decreases in rates were
27 454 associated with similar movements in the price of the relevant in the domestic currency [93–95].
28 455 Deforestation only started to rise again in 2019 [2].

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31 456 The effects of changes in both market access and population density on deforestation over time have
32 457 been shown to be moderated by the extent of previous deforestation and remaining availability of forest
33 458 and land. In frontiers with high forest cover, roads lead to greater deforestation, as found in Ecuador
34 459 [109] and Brazil [75]. In contrast, in areas with less forest remaining, improvements to existing roads
35 460 were associated with lower deforestation [100] and a higher population density leads to high
36 461 competition costs associated with obtaining forest land, resulting in lower deforestation [109].
37 462 Relatedly, high urbanization, which can be seen as an almost late (or at least later)-frontier stage in
38 463 development, was associated with decreased coca-related deforestation in the Colombian Amazon [54]
39 464 and overall deforestation in Brazilian municipalities [114]. Market access and population factors also
40 465 interact with each other in the context of frontier areas; infrastructure improvements have been found
41 466 to attract migration and increased demographic pressure on forests.
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52 468 *(2) How effective have deforestation control policies been across different regions and over time?*

53 469 **Area-based conservation:** The first legal conservation area in the Amazon basin was established in
54 470 1929 in Guyana, but the largest number of legal conservation areas were declared over the last twenty
55 471 years. Today, about 50% of Amazonian lands are under protection [125]. In the Brazilian Amazon,
56 472 comparative studies are not entirely conclusive about the most effective type of conservation area. As
57 473 protected areas technically do not allow for any deforestation and received more regulatory attention,
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3 474 they were deemed most effective, followed by indigenous areas and sustainable use areas [126–129].
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5 475 But two more recent studies found that indigenous areas faced more deforestation pressure and thus
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7 476 provided greater additionality compared to remotely located protected areas and less stringent
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9 477 sustainable use areas [130,131]. In some places, indigenous and protected areas even caused positive
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11 478 spillovers, by reducing deforestation outside of the area. Sustainable use areas generally had the smallest
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13 479 or no impact on deforestation, as they still allow limited clearing [132].

14 480 The additionality of conservation areas also varied depending on the deforestation risk, which in turn
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16 481 was a function of the location of the area. In Bolivia, the impact of conservation areas also depended
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18 482 on the proximate cause of deforestation. Conservation areas appear to have effectively stopped
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20 483 commodity agriculture expansion, but pasture and small-scale agriculture was found within such areas.

21 484 A few studies assessed the impact of land titling within indigenous areas. While one Brazilian
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23 485 Amazon study did not find any effect [133], studies in Brazil [134] and Peru [135] found reductions in
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25 486 deforestation after receiving property titles. Titling might have increased the regulatory attention and
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27 487 enforcement power in these areas which made them more effective. In Ecuador, complex tenure forms
28
29 488 i.e., indigenous areas overlapping with patrimony or protected forest (restricted-use, public and private
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31 489 lands) showed a higher conservation effect than single tenure forms.

32 490 **Financial instruments:** PES pays participants, compensating them for the deforestation reliant
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34 491 economic activities that are being avoided and often targets additional issues besides forest
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36 492 conservation. PES policies have been rigorously analyzed in Guyana, and in the Brazilian, Peruvian
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38 493 and Ecuadorian Amazon and include both UN facilitated REDD+ (*Reducing Emissions from*
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40 494 *Deforestation and forest Degradation* in developing countries) projects as well as integrated PES+ICDP
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42 495 (*Integrated Conservation and Development Projects*), which include the Peruvian National Forest
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44 496 Conservation Program and the Ecuadorian Socio Bosque program.

45 497 The *Norway-Guyana REDD+ program* resulted in lowered deforestation due to mining in Guyana
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47 498 by 35% between 2010 and 2015, counteracting the influence of rising gold prices [136]. Yet,
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49 499 deforestation rates were still rising during the program and increased rapidly once it had ended. The
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51 500 policy is also suspected to have caused leakage into Suriname, as forest loss was concentrated along the
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53 501 border [136]. In Brazil, twelve *REDD+* projects across the entire Brazilian Amazon focused on
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55 502 agriculturally-driven deforestation did not provide additionality (outcomes beyond business as usual)
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57 503 in 2008-2018 [137]. Similarly, a regional REDD+ project in a previously blacklisted (i.e., cut off from
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59 504 agricultural credit) municipality in Mato Grosso did not find any significant additional effect as
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505 deforestation rates were already lowered by the government blacklisting policy as part of PPCDAm
506 (explained below) [138].

507 The Peruvian *National Forest Conservation Program* PES+ICDP decreased deforestation in other
508 areas, but not within enrolled zones in the same community, as forests with low threats were chosen for
509 enrollment. Furthermore, communities with historically high deforestation rates were underrepresented
510 and the conservation effect was only significant in the first year of the five-year contract [139].

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3 511 However, a local PES+ ICDP program for the protection of the Moyobamba watershed significantly
4 512 reduced deforestation by supporting the transition to agroforestry and more sustainable agricultural
5 513 production [140]. Similarly, the REDD+ Sustainable Settlements Project in Pará, Brazil, reduced
6 514 deforestation rates by 50% among participating smallholders, at the expense of pasture rather than
7 515 cropland [141]. Ecuador's *Socio Bosque program* also reduced deforestation [142] and even remained
8 516 effective when payment to participants was suspended [62]. Yet, participation in the program was more
9 517 common for more remote and larger properties compared to low-income smallholders for whom forest
10 518 clearing remained an important economic safety cushion [64].

11 519 **Public policies:** Federal enforcement of environmental policies through fines and embargos in the
12 520 Brazilian Amazon was found to effectively decrease deforestation [102]. Two studies conclude that
13 521 these policies have helped to weaken the relationship between global markets and commodity-driven
14 522 deforestation in the two major hotspots of Pará and Mato Grosso [143].

15 523 In 2004 the *PPCDAm* (Action Plan for the Prevention and Control of Deforestation in the Legal
16 524 Amazon) reversed the steep increase of deforestation in the beginning of the 2000s. After 2008, the
17 525 blacklisting of municipalities and other mechanisms within the PPCDAM provided significant
18 526 additionality in decreasing forest loss by up to 60%, likely due to improved law enforcement and
19 527 monitoring [144–147]. Blacklisting even reduced deforestation in neighboring municipalities,
20 528 indicating positive spillovers of the policy [148]. The PPCDAm disproportionately targeted cattle
21 529 ranching and led to intensification of this land use, sparing forests [146]. Supporting the implementation
22 530 of these policies, the local green municipality program in Pará was partially effective, for example in
23 531 Paragominas, the initial municipality the policy was implemented in [149], but beyond [150].

24 532 Outside of Brazil there are few causal studies of public policies. A broad study covering the entire
25 533 Guiana Shield [55] found that *militarized repression of mining*, starting in 2002, had significantly
26 534 reduced deforestation due to gold mining in French Guiana and the neighboring Brazilian Amapá.
27 535 However, leakage into Suriname is assumed where mining was not regulated. Lack of regional
28 536 cooperation and strategies and the heavy reliance of local economies on gold mining were considered
29 537 the main policy challenges in addressing gold mining as the primary driver of deforestation in the
30 538 Guianas.

31 539 **Supply chain policies:** With the rise of commodity agriculture causing vast deforestation in the
32 540 Brazilian Amazon, it is also the only place where private policies aiming to eliminate deforestation in
33 541 supply chains have been fully implemented [151]. As the most effective supply chain policy, the Soy
34 542 Moratorium reduced soy-related deforestation by 35-55% as it was introduced after 2006 [151,152].
35 543 Evidence regarding leakage due to the Soy Moratorium is mixed. One study by Heilmayr et al. [152]
36 544 finds no leakage, but the most recent study by Villoria et al. (2022) indicates that up to 50% of the
37 545 avoided deforestation in the Amazon was offset by deforestation leakage to other Brazilian forests. The
38 546 local Responsible Soy Project in Pará also successfully decreased deforestation among soy-supplying
39 547 smallholders but was not effective on larger farms [154]. On the contrary, the G4 and TAC cattle

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3 548 ranching agreements did not provide additionality in preventing deforestation. Leakage under the two
4 549 agreements from properties that enrolled earlier to properties that enrolled later was responsible for the
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6 550 overall zero effect of the policy [155].
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8 551 Only one study analyzed supply chain policies for other products. A study on the Forest Stewardship
9 552 Council certification, a *third-party standard*, found it was not effective at reducing deforestation in the
10 553 Peruvian Amazon, due to low stringency of the certification criteria and low enforcement of compliance
11 554 to these criteria within certified operations [156].

12 555 **Policy interactions:** Public policies and institutions are key moderating factors for deforestation
13 556 policy effectiveness. The success of REDD+ and voluntary policies was highly dependent on
14 557 redundancies, complementarities, and antagonisms between: 1) public deforestation control policies
15 558 like PPCDAm and 2) the presence of indigenous areas (both of which reduced deforestation, creating a
16 559 certain redundancy that reduced the additionality of the said program), and 3) the type of land tenure
17 560 present. Policies also indirectly affect deforestation through underlying drivers. For example, road
18 561 building linked to deforestation is often an explicit function of public road building investments, making
19 562 public policy a key underlying driver of infrastructure.
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30 564 (3) *How does the availability and rigor of evidence vary across regions and topics (drivers, causes,*
31 565 *moderating agents, and policy effectiveness)?*

32 566 As shown in Figure 4, agriculture (including cattle ranching), was the most researched proximate
33 567 cause, with 146 cases covering the entire period of 1970-2019. Mining and infrastructure, and their
34 568 linkages, were the next most researched proximate causes. There was no literature on large-scale gold
35 569 mining, despite its presence in the Guiana Shield. There was also no literature specific to other types of
36 570 mining (e.g., iron ore, copper).

37 571 The research hotspots for deforestation from *large-scale (soy) agriculture* are Santa Cruz in Bolivia
38 572 (5 cases) and the Arc of Deforestation in Brazil (36 cases). Large- and small-scale expansion of *oil palm*
39 573 is less well researched (8 cases). Subsistence agriculture has received considerably less research (30
40 574 cases) and this is mostly in the form of very local case studies. Coca research is largely absent in
41 575 Amazonia but is more prevalent outside of the Amazon where coca production is also more common.

42 576 Economic factors, such as commodity prices, access to markets (e.g., proximity or cost-distance to
43 577 riverways, urban centers or export terminals), GDP, and income are the most frequently studied
44 578 underlying drivers (75 cases) and were linked to deforestation driven by large- and small-scale
45 579 agriculture, as well as mining. Economic factors were analyzed in all countries especially in Brazil (44
46 580 cases), but not in Colombia and Venezuela.

47 581 Technological aspects were covered in 37 cases, mainly in the Ecuador (8 cases) and Brazil (27
48 582 cases). Most studies on these drivers refer to the context of agricultural production, but usually just link
49 583 technology to overall deforestation instead of a specific proximate deforestation cause.
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3 584 Institutional factors focused on agricultural settlement policies and land tenure and conflict and were
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5 585 analyzed in 25 cases. Most of these cases (21) were in Brazil where land conflict and tenure were
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7 586 commonly analyzed factors. Environmental factors were discussed in 17 cases, mostly in Brazil and
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9 587 Bolivia. Demographic factors were discussed in 13 cases, limited to the Ecuadorian and Brazilian
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11 588 Amazon. Culture was rarely analyzed (9 cases) and this was predominantly in Beni, Bolivia and various
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13 589 provinces in Ecuador.

14
15 590 Many underlying drivers and policies were linked to general deforestation, but not via a specific
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17 591 proximate cause. Sixty cases examined moderating factors between underlying drivers and proximate
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19 592 causes or policies. Nearly half of these cases (28) had to do with policies or institutions moderating
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21 593 other underlying drivers. Economic factors (mostly market access) and technological ones (relating to
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23 594 the remaining forest area) were the next two largest clusters of moderating factors.

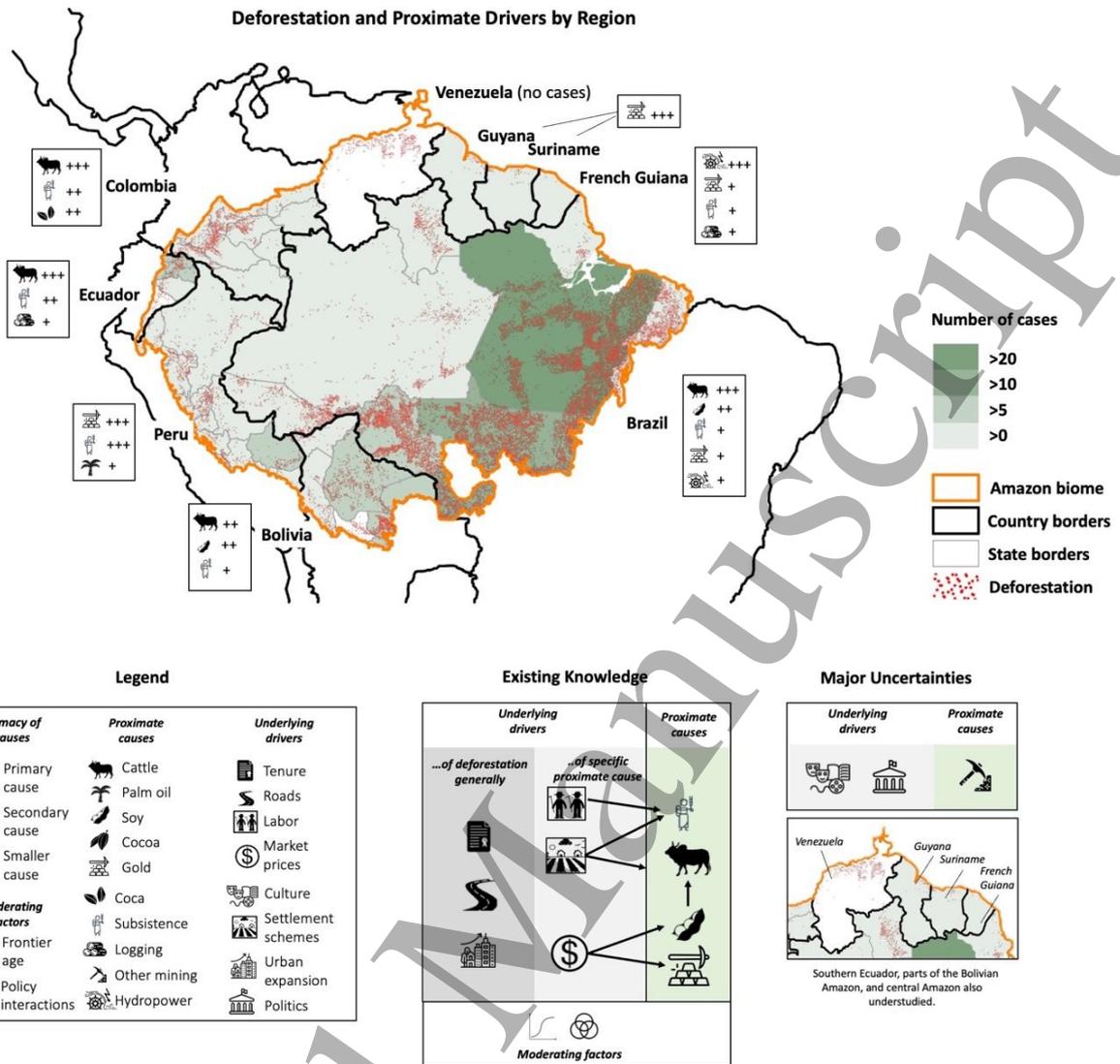
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26 596 *(4) How does the evidence generated by local case study research differ from regional (Amazon-wide)*
27 597 *studies?*

28 598 *Commodity agriculture in Brazil and its links to macroeconomic conditions is the leading focus of*
29 599 *both types of research in Amazonia*

30 600 As indicated in global and Amazon-wide analyses [21,22], deforestation, especially at large scales,
31 601 has become driven mostly by the expansion of commodity agriculture (particularly beef and soy
32 602 production). Brazil has the largest Amazonian territory and overall highest levels of deforestation,
33 603 corresponding to the largest cattle herd and soy production areas. Accordingly, existing literature on
34 604 deforestation drivers and policies follows these broad trends, with 65% of the cases focused on
35 605 deforestation in Brazil, 31% of which were focused on analyzing agriculture as a proximate cause of
36 606 deforestation. Bolivia and Peru are the next biggest hotspots of research on agriculturally-caused
37 607 deforestation.

38 608 Global synthesis reports link this commodity expansion to growth in global and regional
39 609 consumption of these products [1]. Similarly, we observe a shift in the focus of case study research over
40 610 time from smallholder-driven deforestation and underlying household attributes to commodity-driven
41 611 deforestation and macroeconomic underlying drivers. This corresponds closely to a shift in tropical
42 612 deforestation agents described by Rudel [31]. Nevertheless, causal case study research obscures
43 613 important findings from biome-wide studies on how political narratives about modernization and
44 614 economic development influence clearing in the Amazon [157]. Such political economy dimensions are
45 615 inherently more difficult to study at only local scales with causal inference methods. “Chain of
46 616 explanation” approaches, which evaluate the influence of variables interacting at a number of scales,
47 617 are a more common approach to understanding the effects of power relations on land use [158].
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620 *Figure 5: Distribution of cases (since 1970) and cumulative deforestation (since 1985) across the Amazon*
 621 *biome, synthesis of proximate drivers by region, existing knowledge, and major uncertainties. Map of the*
 622 *Amazonian border and deforestation data by MapBiomass Amazonia project (2022).*

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625 Despite an increasing focus on commodity agriculture, both Amazon-wide studies and case studies
 626 show that smallholder farming is still a relevant driver, especially in the western and northern regions
 627 of the Amazon (Figure 5)[21,22]. Kalamandeen et al. [159] found that small-scale clearing has
 628 significantly increased in 2008-2014 in all Amazonian countries and has spread with lower density into
 629 isolated areas. There is some speculation that large-scale farmers may be attempting to evade policy
 630 enforcement by clearing smaller, non-detectable areas [160], but this has not been definitively proven.
 631 Regardless of who the agents of change are, our review brings into focus that such small-scale patches
 632 of deforestation are relatively poorly studied despite their ongoing importance to deforestation and
 633 degradation in the Amazon biome.

633

634 *Case studies illuminate the importance of mining as a key deforestation cause in Peru and the*
635 *Guianas*

636 Gold mining has been confirmed as a major cause of deforestation in the Guianas and in Madre de
637 Dios, Peru and comprised 12% of the cases in these two regions. Yet it still shows up as relatively minor
638 overall cause of deforestation [22] and is largely obscured as a subset of commodity-driven
639 deforestation [21] in pan-tropical or Amazon wide studies, with the exception of the very recent study
640 by Giljum et al. [161] focused entirely on industrial mining.

641 Additionally, our findings point to growing mining hotspots in eastern Venezuela and the Tapajós–
642 Xingu region in Brazil. Forest loss due to gold mining in these regions increased dramatically since
643 2001 and is moving into remote areas of higher conservation priority that overlap with greenstone areas
644 [25]. Mining activity in official concessions can be small- or large-scale but illegal mining is rife and
645 difficult to police due to accessibility issues and lack of resources by relevant countries. So far, military
646 operations against mining in French Guiana and Madre de Dios (e.g. Operation Mercury) were effective
647 but have displaced the cause to other areas [55,162]. As Alvarez-Berríos and Aide [25] confirm,
648 deforestation due to gold mining is also connected global markets and rising gold prices.

649
650 *Many areas of Amazonia are completely ignored*

651 On the national Amazon level, the distribution of cases is comparable to the share of deforestation
652 and the share in Amazonian area. Ecuador has a particularly high density of cases while Brazil has the
653 largest amount of forest loss per area and cases (Figure 3b, 4) [2,163]. Within countries, research
654 hotspots are visible. Outside of Brazil, well-studied areas with a high density of local and regional cases
655 include Madre de Dios in Peru (8 cases), Beni (6 cases) and Santa Cruz (5 cases) in Bolivia and Northern
656 Ecuador (12 cases). In Brazil, both forest loss and number of local and regional cases were particularly
657 high in the Pará, Mato Grosso, Rondônia (136 cases collectively). These three states constitute the
658 majority of the Brazilian Arc of Deforestation, a region named for its historically high levels of
659 deforestation. Most of these hotspots have higher rates of forest loss compared to the rest of their
660 respective countries [2,163].

661 Despite the presence of deforestation in the Venezuelan Amazon, the Southern Ecuadorian Amazon
662 (departments of Zamora Chinchipe and Morona Santiago) and Bolivian Amazon (departments of
663 Cochabamba and La Paz) [2], the reviewed literature entirely misses these areas. In the remaining
664 countries, at least one case study researched the entire Amazon region of the respective country. This
665 could indicate that research is done where most urgently needed. However, areas with high forest cover
666 and low deforestation and research coverage, such as the Guianas or the central Amazon region, should
667 not be neglected. The sparse, small-scale forest loss in this area is not monitored and analyzed with the
668 same attention as deforestation in other parts. Therefore, these regions are at high risk to ineffective or
669 delayed conservation, especially if they are reached by large-scale deforestation (e.g., if existing
670 colonization frontiers develop into commodity frontiers as has occurred in Brazil's Arc of Deforestation

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3 671 to the South). Furthermore, ecological impacts of forest loss on intact, remote forests are unknown given
4 672 their poor coverage, but may be even more detrimental than in areas that are already degraded or
5 673 affected to some degree. For example, the 227 “hyperdominant” tree species that make up over half of
6 674 all Amazonian trees tend to be located in specific regions, whilst over 10,000 trees are rare, poorly
7 675 known and potentially threatened [164]. In addition, remote areas tend to have high degrees of
8 676 endemism: for example there are over 40% endemic plant species in the Kaieteur Falls National Park
9 677 in Guyana [165].

15 678 *Strengths and shortcomings in forest conservation*

16 679 As a global review by Börner et al. [20] showed, policy analyses of forest conservation are generally
17 680 lacking in the Guianas and Venezuela while Brazil has the highest number of studies worldwide. Forest
18 681 conservation is also frequently studied in Peru and Ecuador. Accordingly, the distribution of policy
19 682 research in Börner et al. [20] is consistent with the distribution of policy cases in this review.

22 683 The meta-analysis by Börner et al. [20], which also considered effect sizes, found that indigenous
23 684 territories avoid more deforestation than protected areas or any other policy studied, on average. Our
24 685 synthesis (Figure 4) does not consider effect sizes given the wide differences in the policy types and the
25 686 spatial scopes they cover. We instead only consider whether they were additional in providing reduced
26 687 deforestation beyond business as usual within their target areas. As summarized in Figure 6, we find
27 688 that indigenous areas, restricted use areas, protected areas, PPCDAm (a suite of deforestation control
28 689 policies in Brazil), militarized repression of mining in the Guiana Shield, and Socio Bosque (a national
29 690 integrated conservation and development program in Ecuador) were all, on average, effective in
30 691 providing reduced deforestation beyond business as usual. Several interventions (indigenous areas,
31 692 mixed use areas, and ZDCs had cases of negative spillovers, while only protected areas and ICDPs
32 693 included only positive spillovers.

35 694 Global reviews indicate that PES schemes have been moderately effective in reducing forest loss in
36 695 Latin America but have struggled to be impactful in terms of coverage area, in the absence of
37 696 deforestation risks, and when confronted with insecure land tenure and weak governance. This was
38 697 confirmed by our review in the Peruvian and Ecuadorian Amazon. As PES are voluntary and used for
39 698 compensating opportunity costs of avoided deforestation, their effectiveness also depends on the global
40 699 willingness to pay for these schemes and institutional settings such as the level of market integration,
41 700 an important underlying driver in this review, and who receives the payment (governments or steward
42 701 populations). The effectiveness of ICDP approaches generally seems to have improved over the past
43 702 years, as newer reviews now indicate their increased additionality [166–169]. Overall, the diversity of
44 703 outcomes and range of spillovers across different policy types calls into question any clear “best” single
45 704 policy approach, instead supporting calls for broader conservation and development policy mixes
46 705 [170,171].

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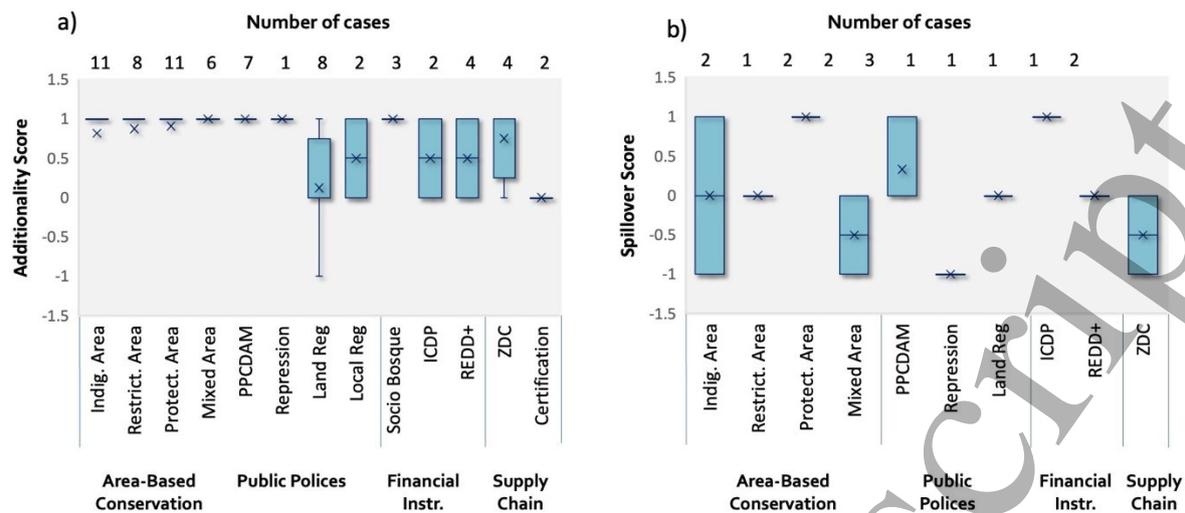


Figure 6: Comparison of the additionality (a) and spillovers (b) of different deforestation control policies in the Amazon. Box plots show the mean and distribution of additionality scores (1=positive additionality/spillover, 0=no additionality/spillover, -1=negative additionality/spillover). The number of cases per policy is listed at the top. Notes: Indig. = indigenous, restrict. = restricted use, protec. = protected, PPCDAm = the Plan for the Prevention and Control of Deforestation, repression = militarized repression of mining, land reg = land registration, Green muni = the green municipality program, ICDP = integrated conservation and development programs, Socio Bosque = a national ICDP in Ecuador, REDD+ = reducing emissions from deforestation and degradation programs, ZDC = zero-deforestation commitments.

6. Conclusion

This review combines and synthesizes an analysis of deforestation drivers and the impacts of policies targeting deforestation in Amazonia while providing a comprehensive, yet accessible overview on the topic. It demonstrates the geographic (and temporal) heterogeneity and complexity of deforestation drivers and governance issues and concludes that not all drivers are equally at play across the region. The various policy instruments and their successes also vary. The review therefore shows the importance of understanding local context and regional differences to better devise appropriate governance options.

Our results show that the distribution of case study research parallels rates of deforestation across the biome. This focus on deforestation hotspots results in a lack of research in regions with lower, but significant rates of deforestation and areas where new drivers such as mining and infrastructure expansion are emerging. There is no causal research of drivers or policies on Venezuela, southern Ecuador and parts of the Bolivian Amazon, and very little in the Guianas or the central Amazon region.

Besides large-scale agricultural expansion and economic underlying factors, other drivers have received comparatively little research attention. Mining, smallholder farming and infrastructure expansion as well as logging and fires may have a smaller direct impact on forests but are also less

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3 734 understood in today's evidence base. Given the emerging importance of critical minerals for electric
4 735 batteries [172] and increasing extent of forest degradation in Amazonia [173], the large blind spots
5 736 regarding drivers and impacts of mining, logging, and fire in Amazonia are very worrying.

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8 737 As research is concentrated in deforestation hotspots and where forest loss occurred, areas with low
9 738 deforestation rates and high forest cover such as the Guianas, are less researched. However, these
10 739 regions are still at risk of increased forest loss, especially as new drivers like mining and infrastructure
11 740 emerge without suitable control policies.

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14 741 Research on sociocultural, behavioral, institutional, and technological factors as underlying causes
15 742 is scarce, and concentrated in either Brazil or certain research hotspots. Very few studies connect
16 743 proximate and underlying drivers and policies, via moderating factors. This reduces the understanding
17 744 of how drivers and policies affect each other and how their interactions affect deforestation. For
18 745 example, analyzing which drivers are impacted by a policy is essential to understand how its
19 746 effectiveness can be improved. More interlinked research that focuses on the complex pathways and
20 747 networks of deforestation processes, particularly indirect land use change, is needed.

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23 748 Finally, only few studies compare or connect specific regions of the Amazon. As the biome is
24 749 integrated in many ways, effects of a local policy or driver on forests may carry across borders to other
25 750 countries and biomes. Studying these dynamics and comparing local contexts may provide essential
26 751 knowledge for conservation in the Amazon.

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29 752 Based on our review and the comparison of our results to additional literature, we define the
30 753 following main areas in need of more research in the Amazon:

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35 754 i. proximate causes of deforestation that are not agricultural, such as mining and infrastructure,
36 755 ii. links between forest degradation and deforestation,
37 756 iii. deforestation in areas with low deforestation rates and high forest cover,
38 757 iv. cultural, institutional, and technological underlying drivers of deforestation,
39 758 v. research on the politics and policy pathways underlying deforestation increases and
40 759 reductions, and
41 760 vi. comparative studies between Amazonian subregions and countries.

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44 761 These findings may help to guide future research and conservation in the Amazon, for instance by
45 762 providing a basis for improving the design of future conservation policies in the region, moving beyond
46 763 existing approaches that are not suited to the local context or miss key drivers.

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57 770 (<https://www.glp.earth>).

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3 7714 772 **Author contributions**

5 773 AH, SL, and RG designed the research. AH and RG analysed the data with support from SL. AH
6 774 wrote the original draft. AH and RG created the figures. SL, RG, JB, XR, IB, and DA all helped
7
8 775 develop this draft and contributed to writing, reviewing, and editing. RG, AH and SL undertook the
9
10 776 major revisions with review and edits from all other co-authors. RG finalised the manuscript.
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